

# Terrestrial protected areas maintain freshwater ecosystem resilience to costly aquatic invasive species in the Panama Canal

Jorge Salgado (✉ [Jorge.SalgadoBonnet@nottingham.ac.uk](mailto:Jorge.SalgadoBonnet@nottingham.ac.uk))

University of Nottingham <https://orcid.org/0000-0003-0670-0334>

**María Vélez**

University of Regina, Department of Geology, 3737 Wascana Parkway, Regina, Saskatchewan, S4S 0A2, Canada.

**Catalina Gonzalez-Arango**

Universidad de Los Andes

**Aaron O'Dea**

Smithsonian Tropical Research Institute, PO Box 0843-03092, Balboa, Republic of Panama.

---

## Brief Communication

**Keywords:** ecosystem resilience, invasive species, Panama canal, freshwater ecology

**Posted Date:** June 30th, 2021

**DOI:** <https://doi.org/10.21203/rs.3.rs-659064/v1>

**License:**  This work is licensed under a Creative Commons Attribution 4.0 International License.

[Read Full License](#)

---

1 **Terrestrial protected areas maintain freshwater ecosystem resilience to**  
2 **costly aquatic invasive species in the Panama Canal**

3 Jorge Salgado\*<sup>1,2,3,5,</sup> María I. Vélez<sup>4,</sup> Catalina González-Arango<sup>1,</sup> Aaron O'Dea<sup>5,6</sup>

4 1. Facultad de Ingeniería, Universidad Católica de Colombia, Bogotá, Avenida Caracas # 46 -72.

5 2. School of Geography, University of Nottingham, Nottingham, UK.

6 3. Laboratorio de Palinología y Paleoecología Tropical, Departamento de Ciencias Biológicas,  
7 Universidad de Los Andes, Carrera 1# 18A - 12, Bogotá, Colombia.

8 4. University of Regina, Department of Geology, 3737 Wascana Parkway, Regina, Saskatchewan,  
9 S4S 0A2, Canada.

10 5. Smithsonian Tropical Research Institute, PO Box 0843-03092, Balboa, Republic of Panama.

11 6. Department of Biological, Geological and Environmental Sciences, University of Bologna,  
12 Bologna, Italy.

13

14 Corresponding author: Jorge Salgado, email [Jorge.SalgadoBonnet@nottingham.ac.uk](mailto:Jorge.SalgadoBonnet@nottingham.ac.uk)

15 ORCID: <https://orcid.org/0000-0003-0670-0334>

16 **Abstract**

17 River damming is expected to proliferate across Tropical American Rivers in the forthcoming  
18 decades with expected declines in ecosystem health to costly invasive species. Historical data  
19 and modern aquatic plant surveys of one of the largest and oldest tropical dam projects (the  
20 > 100 years old Panama Canal) reveal that modern plant communities in areas adjacent to  
21 terrestrial Natural Protected Areas (tNPAs) retain a pre-damming community structure that  
22 is apparently more resistant to invasive species. Establishing tNPAs adjacent to impounded  
23 rivers could be a cost-effective nature-based solution for tropical reservoir management.

## 24 **Main text**

25 The construction of large dam projects (> 15 m high) is predicted to accelerate across Tropical  
26 America for hydropower and water provision<sup>1,2</sup> (Fig. 1a). While these projects can address  
27 important socio-economic needs<sup>3</sup>, they often have far-reaching effects on aquatic  
28 biodiversity<sup>2,4</sup> through the altering of biogeochemical cycles, water quality, and primary  
29 productivity<sup>5,6</sup>, and the restriction of upstream-downstream movement of organisms, water  
30 and nutrients<sup>4</sup>. Less understood, however, is the impact of alien invasive species (IAS) on  
31 dam reservoirs<sup>7,8</sup>. The often-poor water quality of man-made reservoirs, coupled with the  
32 high variability in water levels from regulation, can stress native species while facilitating  
33 conditions for more generalist species<sup>9</sup> and this may explain why reservoirs have been  
34 suggested to be less resilient to IAS than natural lakes<sup>7,9,10</sup>.

35 Despite considerable progress in identifying the threats of river damming to ecosystem  
36 biodiversity in the American tropics<sup>1,2,4,8</sup>, best practices to protect aquatic biodiversity  
37 continue to be debated<sup>11,12</sup>. As riverine systems are primarily defined by their hydrology,  
38 conservation strategies have focused on protecting and restoring hydrological regimes to  
39 preserve the ecosystem function and biodiversity<sup>1,4,13</sup>. Yet, there remain substantial gaps in  
40 our understanding about the effects of land-use change, water pollution, and IAS on the  
41 ecology of dammed rivers<sup>14</sup>. The designation of terrestrial Natural Protected Areas (tNPAs)  
42 has been suggested as an additional solution to maintain ecosystem integrity in  
43 reservoirs<sup>14,15,16</sup>. Simulation of conservation strategies in the Amazon basin, has shown that  
44 the joint implementation of aquatic and tNPAs, could bring overwhelming positive effects in  
45 conserving both terrestrial and aquatic biota<sup>12</sup>. However, whether such an approach can be

46 effective in heavily impacted ecosystems, such as man-made reservoirs, has yet to be  
47 demonstrated.

48 Lake Gatun in the Panama Canal, Panama, is one of the oldest (> 100 years) and largest (425  
49 km<sup>2</sup>) reservoirs in the American Tropics (Fig. 1c). It has been invaded by a suite of species  
50 including a variety of fish (e.g., the peacock bass *Cichla ocellaris*), mollusc (e.g., the red-  
51 rimmed melania snail *Melanooides (Thiara) tuberculata*, and the South American apple snail  
52 *Pomacea bridgesii*), mammals (e.g., the American manatee *Trichechus manatus*), and  
53 macrophytes (e.g., the water thyme *Hydrilla verticillata* and the water hyacinth *Eichhornia*  
54 *crassipes*), either accidentally or deliberately<sup>17,18</sup>. Furthermore, since the canal was  
55 completed in 1914, it came under the protection of the US government and extensive forest  
56 areas were safeguarded in the northern half of the lake by the creation of a nature reserve  
57 in Barro Colorado Island (BCI) in 1923<sup>19</sup>. The subsequent inclusion of five adjacent  
58 peninsulas in 1946 created what is currently known as the Barro Colorado Nature  
59 Monument–BCNM, whilst two more National Parks (Soberanía and the Chagres) were  
60 established later in the early 1980s on the eastern side of the lake<sup>19</sup> (Figure 1c). These  
61 unique long-running sanctuaries are currently characterised by the growth of mature  
62 secondary forest that positively influence in the hydrology and water quality of the  
63 associated tributary rivers<sup>19,20,21</sup>. They attenuate flood peaks and extreme droughts, help  
64 retain sediments, nutrients, and runoff<sup>19,20,21</sup>. However, the extended effects of tNPAs on  
65 the functioning and ecology of the Gatun Lake has yet to be addressed. Gatun Lake  
66 therefore presents a unique opportunity to explore the ecological effects of tNPAs on  
67 aquatic communities and their resilience to anthropogenic and natural change over a  
68 relatively long period of time; an approach that may provide a lense to help predict future

69 dynamics of the many hundreds of dam projects that are proposed across lowland tropical  
70 areas and to offer useful insights into best practices to maintain ecosystem integrity.

71 In this study, we combined long-term paleoecological data from a sedimentary core with  
72 high-resolution spatial surveys of modern aquatic plant (macrophyte) communities in lake  
73 areas adjacent and not adjacent to tNPAs (Fig. 1c) to explore the role of tNPAs and the  
74 resilience of the aquatic ecosystems. To achieve this, we (1) defined a baseline, pre-dam  
75 community and ecosystem conditions, (2) surveyed the macrophyte composition of the  
76 tNPAs and other non-tNPAs, and (3) compared the macrophyte communities relative to pre-  
77 dam species (for detailed methods see<sup>17</sup> and supplementary material).

78 Our findings show that areas of Gatun Lake that are included within a tNPA sustain  
79 macrophyte communities that are more similar in their composition to baseline  
80 communities, than those from outside tNPAs (Fig. 2a). Macrophyte communities that are  
81 adjacent to tNPAs are often dominated by native macrophyte species such as *Nymphaea*  
82 *ampla*, *Ceratophyllum demersum*, *Cabomba piauhyensis*, and *Utricularia vulgaris* (Fig. 2b).  
83 These baseline taxa rarely occur in areas not adjacent to tNPAs or in the parental Chagres  
84 River (2b). In contrast, macrophyte communities in the Chagres River and areas not adjacent  
85 to tNPAs are typically dominated by the invasive *H. verticillata* and *E. crassipes*.

86 Communities in lake areas located within transition zones of tNPAs and not protected land  
87 are characterized by a mixture of both native and IAS. In addition, macrophyte communities  
88 occurring in lake areas adjacent to tNPAs have higher beta diversity (i.e., a greater variation  
89 in species composition and abundances between sampling points) than the more  
90 homogenous, IAS-dominated lake areas not adjacent to tNPAs and in the Chagres River (Fig.  
91 2a). Thus, tNPAs are suggested to have a striking positive effect on their adjacent aquatic

92 ecosystems by promoting diversity, increasing community heterogeneity, and potentially  
93 enhancing the ecosystem ability to resist the establishment of IAS. While recent simulations  
94 predict a positive impact of tNPAs on lowland tropical freshwater ecosystems<sup>12</sup>, our result is  
95 surprising because the Gatun Lake is a reservoir with a long history of human impacts and  
96 has been heavily invaded<sup>17,18</sup>.

97 Reservoirs are complex hydrological systems that combine numerous features of both  
98 riverine and lacustrine environments. This river-lake continuum has proved to influence the  
99 distribution of macrophytes in other similar large, lowland, Tropical American dams such as  
100 Itaipu in Brazil, where floating plants dominate in the riverine fraction of the reservoir,  
101 submerged plant species thrive across the lake portion, and both co-exist in the riverine-lake  
102 fraction<sup>7,22</sup>. In Gatun Lake, physical-chemical characteristics supports such riverine-lake  
103 transition (Fig. 2c) and the studied lake areas adjacent to tNPAs are located within the river-  
104 lake portion. Thus, our results could be reflecting such river-lake gradient. However, the  
105 alike macrophyte IAS dominated communities in the Chagres River and in lake areas not  
106 adjacent to tNPAS (i.e., the lake portion of the reservoir), and modern plant communities in  
107 lake areas adjacent to tNPAs resembled those at pre-dam times. These patterns,  
108 harmoniously suggest that the aquatic flora is not necessarily following a riverine-lake  
109 gradient.

110 Currently the mechanisms explaining the macrophyte patterns between lake areas adjacent  
111 to tNPAs vs not tNPAS are unresolved. However, our results suggest that macrophytes may  
112 be responding to what appears to be a gradient in habitat complexity provided by land  
113 protection. Adjacent tNPAs may promote a range of light conditions in the lake littoral areas  
114 via varying canopy forest cover, retain nutrients and runoff, while promote dissolved

115 organic matter load<sup>19,20,21,23,24</sup>. These various factors could encourage, varying degrees of  
116 water clarity conditions for different plants to exploit. Heterogeneous habitats sustain  
117 macrophyte diversity<sup>25,26,27</sup>, while limit the dominance of macrophyte IAS via species  
118 coexistence<sup>27,28</sup>. Similarly, reductions in water quality<sup>25</sup> and IAS spread<sup>29</sup>, could homogenise  
119 macrophyte communities, through the dominance of few good competitors. The observed  
120 heterogenous water clarity conditions in lake areas adjacent to tNPAs (supplementary  
121 material Fig. 1), therefore, appears to favour habitat structure for pre-dam macrophyte  
122 communities, while limiting the expansion of IAS, which have become to some extent  
123 problematic in the more human influenced areas of the Chagres River and in those lake  
124 areas with not association to tNPAs.

125 The economic and ecological impact of aquatic IAS on man-made lakes across Tropical  
126 America remains a critically understudied problem and long-term records are few and far  
127 between making predictions and recommendations challenging. In other, better studied  
128 (higher latitude and less-biodiverse) systems, costs can frequently exceed many hundreds of  
129 millions of US\$ annually<sup>30</sup> and global estimates place the impact at more than US\$20 billion  
130 a year<sup>32</sup>. And yet, invaded-prone tropical inland freshwater systems receive relatively little  
131 protection against exploitation and development compared to terrestrial and marine  
132 ecosystems. It is therefore critically important that as dam projects proliferate across  
133 developing countries in Tropical America (Fig. 1) and IAS threaten established reservoirs<sup>7</sup>, all  
134 opportunities to curb invasions and protect native biodiversity must be explored effectively  
135 and economically<sup>11</sup>.

136 Our findings illuminate a potentially overlooked role for tNPAs in maintaining aquatic native  
137 biodiversity and ecosystem integrity of tropical impounded systems. Our data is limited to



138 macrophytes, but the implications likely extend more broadly to those invertebrates and  
139 vertebrate community members that rely on macrophyte community structure<sup>32</sup>. Questions  
140 remain however on the universality of the role of tNPAs and the mechanisms driving the  
141 maintenance of pre-dam species populations. Nonetheless, the inclusion of freshwater  
142 ecosystems within tNPAs is suggested to improve the protection of aquatic biodiversity  
143 without undermining terrestrial conservation goals<sup>11,12</sup>. Thus, to prevent river-dammed  
144 ecosystem collapse across the region, we recommend: (i) other tropical protected man-  
145 made reservoirs to be similarly studied to elucidate mechanism and applications; and (ii)  
146 whenever possible, areas in the dam reservoir should be integrated within tNPAs. Such  
147 considerations are critically important given the predicted acceleration of dam projects, the  
148 ever-increasing dispersals of IAS and the fact that highly biodiverse lowland inland  
149 ecosystems are disproportionately sensitive to land use and climatic changes across the  
150 globe<sup>33,34</sup>.

## 151 **ACKNOWLEDGEMENTS**

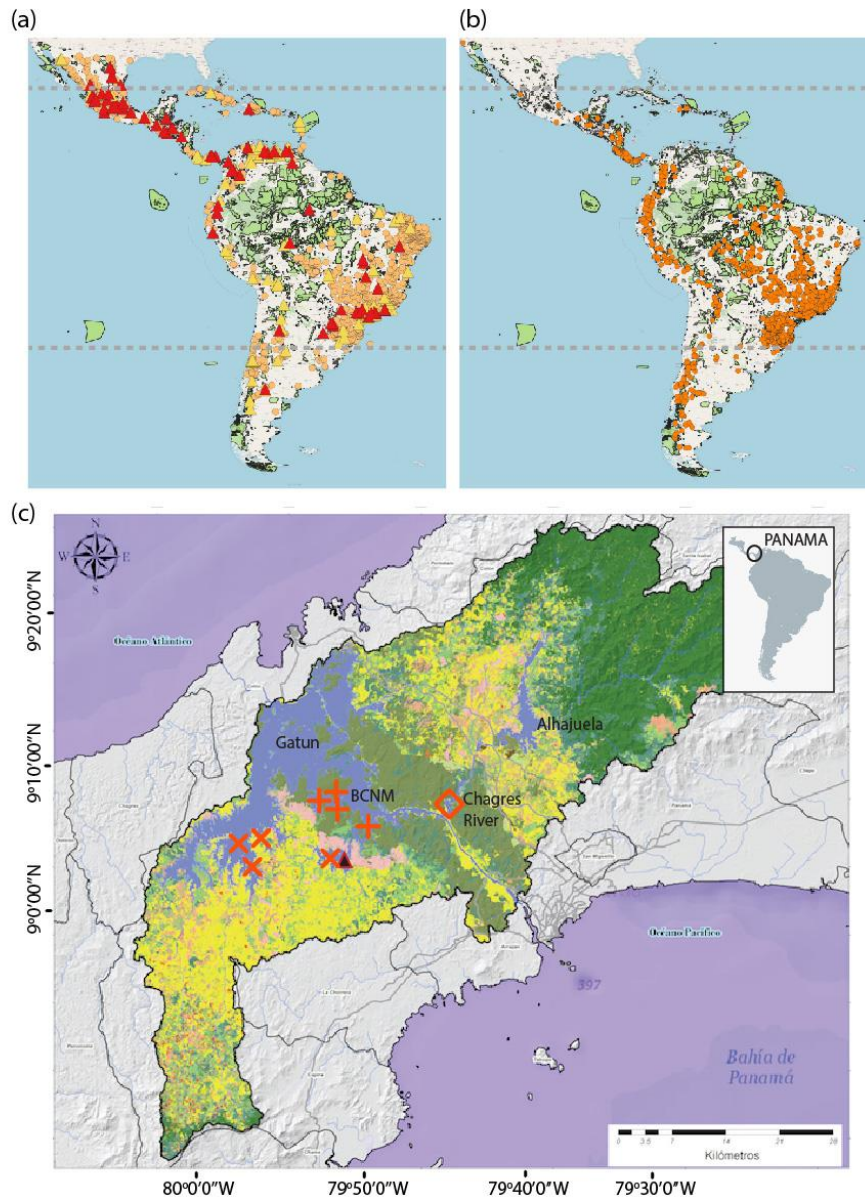
152 We thank the Smithsonian Tropical Research Institute (STRI) for funding fieldwork and  
153 supporting JS through a Latin American Researchers Fellowship. We thank “Autoridad del  
154 Canal de Panamá (ACP)” for their support and site access on Gatun Lake. We thank  
155 Universidad de Los Andes and COLCIENCIAS for supporting JS under the postdoctoral  
156 programme “Es tiempo de volver” Convocatoria 2015. We thank Universidad Católica de  
157 Colombia for supporting JS research the “Convocatoria de Investigación 2020”. We thank  
158 Max Titcomb, Victor Frankel, Felix Rodriguez, Luis J. de Gracia, Marcos Alvarez, Maria  
159 Pinzon, Jorge Morales, Brigida de Gracia and Marcela Herrera for fieldwork, laboratory  
160 assistance and hospitality. We thank Steve Paton for limnological monitoring data provision.

161 The contemporary surveys and collection and exporting of sediment material was assessed  
162 under the ARAP collecting permit No. 25 and under the authorization of the ACP (permit  
163 granted on 13/03/2013). AO was supported by the *Sistema Nacional de Investigación*  
164 (SENACYT).

## 165 REFERENCES

- 166 1. Angarita, H. et al. *Hydrol. Earth Syst. Sci.* **22**, 2839–2865 (2018).
- 167 2. Winemiller, K. O. et al. *Science* **351**, 128–129 (2016).
- 168 3. Moran, E. F. et al. *Proc. Natl. Acad. Sci. U. S. A.* **115**, 11891–11898 (2018).
- 169 4. Carvajal-Quintero, J. D. et al. *Conser. Lett.* **10**, 708–716 (2017).
- 170 5. Maavara, T. et al. *Nature* **8**, 1–10 (2017).
- 171 6. Agostinho, A. A. et al. *Theoretical Reservoir Ecology and Its Applications*, 1999.
- 172 7. Thomaz, S. M. et al. *Hydrobiologia* **746**, 1–12 (2015).
- 173 8. Pereira, L. S. et al. *Hydrobiologia*, **817**, 71–84 (2018).
- 174 9. Havel, J. E. et al. *BioScience* **55**, 518–525 (2005).
- 175 10. Johnson, P. T. et al. Dam invaders: impoundments facilitate biological invasions  
176 into freshwaters. *Front. Ecol. Environ.* **6**, 357–363 (2008).
- 177 11. Abell, R. & Harrison, I. J. *Science* **370**, 38–39 (2020).
- 178 12. Leal, C. G. et al. *Science* **370**, 117–121 (2020).
- 179 13. Anderson, E. P. et al. *Climate change and biodiversity in the tropical Andes* 326-  
180 338 (2011).
- 181 14. Finlayson, C. M. et al. *Freshwater ecosystems in protected areas: Conservation*  
182 *and management*. Routledge (2018).
- 183 15. Saunders, D. L. et al. *Conserv. Biol.* **16**, 30–41 (2002).

- 184 16. Arthington, A. H. et al. Managing specific freshwater ecosystems. In *Freshwater*  
185 *ecosystems in protected areas: Conservation and management* (pp. 144-176).  
186 Taylor & Francis (2017).
- 187 17. Salgado, J. et al. *Sci. Total Environ.* **729**, 138444 (2020).
- 188 18. Castellanos-Galindo, G. A. et al. *Nat. Ecol. Evol.* **4**, 1444-1446 (2020).
- 189 19. Condit, R., et al. *BioScience*, **51**, 389–398 (2001).
- 190 20. Ogden, F.L. et al. *Water Resour. Res.* **49**, 8443–8462 (2013).
- 191 21. Stallard, R. F. et al. *Water Resour. IMPACT* **12**, 18-20 (2010).
- 192 22. Mormul, R. P. et al. *Rev. Biol. Trop.* **58**, 1437–1451 (2010).
- 193 23. Suescún, D. et al. *Reg. Environ. Change* **17**, 827-839 (2017).
- 194 25. Fu, H. et al. *Sci. Total Environ.* **687**, 206–217 (2019).
- 195 25. Salgado, J. et al. *Ecosphere* **9**, e02406 (2018).
- 196 26. Salgado, J. et al. *J. Paleolimnol.* **60**, 311– 328 (2018).
- 197 27. Capers, R. S. et al. *Ecology* **88**, 3135-3143 (2007).
- 198 28. Salgado, J. et al. *Riv. Res. Appl.* (accepted)
- 199 29. Muthukrishnan, R. & Larkin, D. J. *Glob. Ecol. Biogeogr.* **29**, 656–667 (2020).
- 200 30. Oreska, M. P. & Aldridge, D. C. *Biol. Invasions* **13**, 305–319 (2011).
- 201 31. Cuthbert, R. N. et al. *Sci. Total Environ.* **775**, 145238 (2021).
- 202 32. Jeppesen, E. et al. *The structuring role of submerged macrophytes in lakes (Vol.*  
203 *131)*. Springer Science & Business Media (2012).
- 204 33. Barlow, J. et al. *Nature* **559**, 517–526 (2018).
- 205 34. Newbold, T. et al. *Nat. Ecol. Evol.* **4**, 1630–1638 (2020).
- 206
- 207



208

209 **FIGURE 1** Map of Central and South America showing the geographical distribution of **(a)**

210 terrestrial natural protected areas (tNPAs; green), documented large (> 15 m high) dams

211 (circles), dams associated with tNPAs (yellow triangles), and dams associated with tNPAs

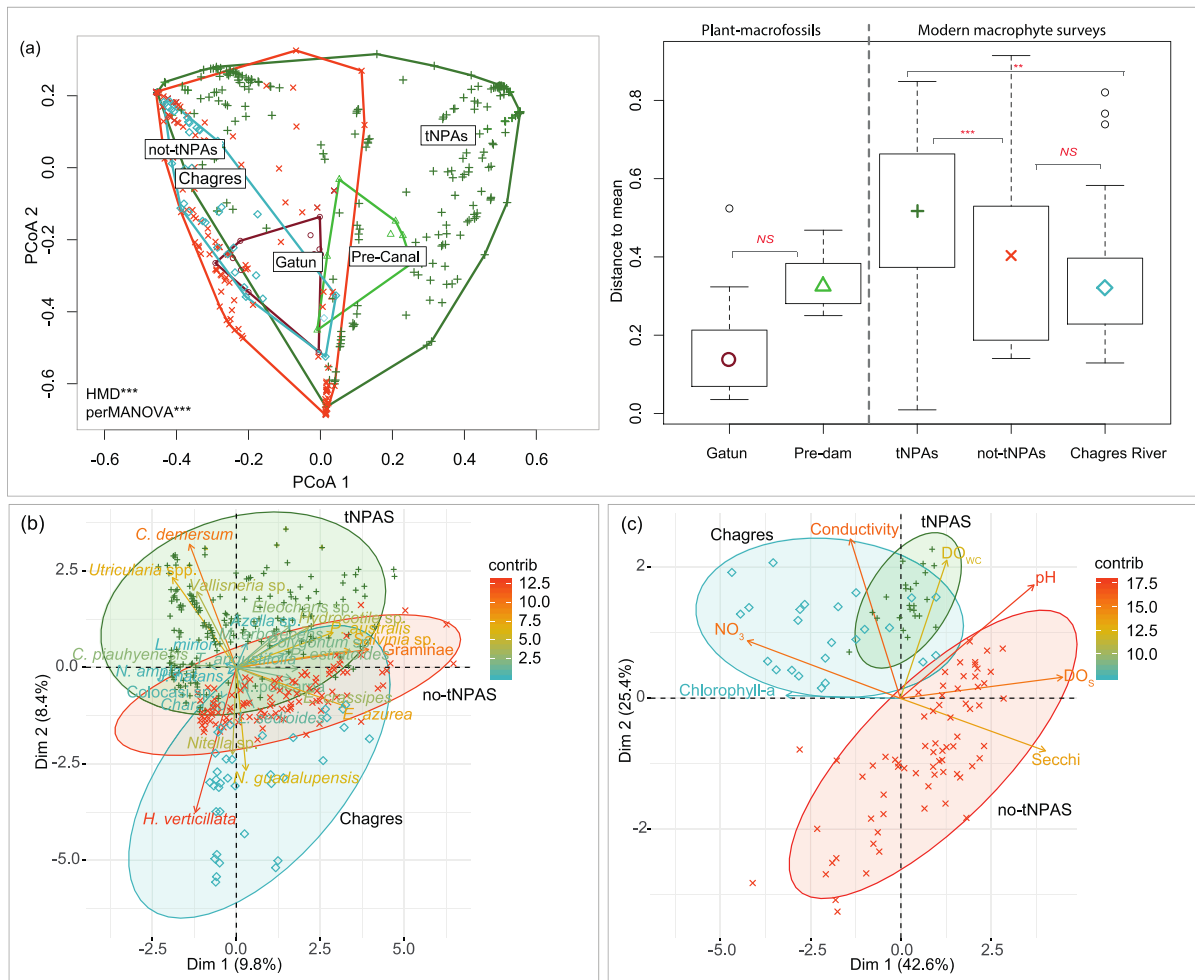
212 with macrophyte invasive species records (red triangles); **(b)** proposed hydropower dams;

213 and **(c)** map showing the location, watershed, and vegetation cover of Gatun Lake (Panama

214 Canal). tNPAs are shown in dark-green, lake sampling areas adjacent to tNPAs in red crosses,

215 lake sampling areas not adjacent to tNPAs in red Xs, and sampling areas in the Chagres River

216 in a red diamond. Coring location of LGAT1<sup>17</sup> core is indicated by a red triangle.



217

218 **FIGURE 2 (a)** Principal coordinate analysis plot (PCoA) of a homogeneity multivariate  
 219 dispersion test (HMD) showing the variations of contemporary macrophyte species  
 220 abundances (community heterogeneity) in three study areas of the Gatun Reservoir: the  
 221 Chagres River (blue diamonds), lake areas adjacent to tNPAs (green crosses), and lake areas  
 222 not adjacent to tNPAs (not-tNPAs; red Xs), and of plant macrofossils at two time periods in  
 223 Gatun Lake (collected in a not-tNPAs): pre-dam times (pale green triangles) and Gatun-lake  
 224 times (dark red circles). The degree of community heterogeneity of each study area is also  
 225 shown as boxplots. The distance to mean of each study area is indicated in the PCoA plot by  
 226 each area's label and in the boxplots by each area symbol. Compositional differences  
 227 between the study areas were assessed via Permutational multivariate analysis of variance

228 (PERMANOVA). Differences in distance to means of each the study areas were assessed via  
229 *post hoc* pairwise comparison via Tukey Honest test under a significance level of  $p \leq 0.05$ .  
230 \* $p \leq 0.05$ ; \*\*  $p \leq 0.01$ ; \*\*\* $p \leq 0.001$ ; NS=not significant.; **(b)** Principal component analysis  
231 (PCA) plot showing the distributions of the macrophyte species in the three study areas; (c)  
232 PCA plot showing the variation of selected physicochemical parameters in the three study  
233 areas. Dissolved oxygen (DO) in the water column (w) and the lake surface (s); Nitrates =  
234  $\text{NO}_3$ .

## Supplementary Files

This is a list of supplementary files associated with this preprint. Click to download.

- [Salgadoetal.Supplementarymaterial.pdf](#)
- [Salgadoetal.Supplementarymaterial.pdf](#)