

A Comparative Study Of Linear And Non-Linear Methods of Isotherm Parameters For Biosorption Of Sodium Diclofenac Onto Calcined Cow Leather (CCL)

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Research

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1 **A comparative study of linear and non-linear methods of isotherm**
2 **parameters for biosorption of Sodium Diclofenac onto Calcined Cow**
3 **Leather (CCL)**

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21 **Abstract**

22 In the current work, the activated biosorbent prepared from Calcined Cow Leather (CCL) was
23 used to study the biosorption potential on anti-inflammatory, namely Sodium Diclofenac and
24 derived from phenylacetic acid from the group of arylcarboxylic acids. Thus, it has been
25 selected because they are widely used and potential associated toxicological effects and also
26 apparently detected in aqueous environments. The biosorption study was carried out to examine
27 the biosorption mechanism by applying the five isotherms models represented by the
28 Langmuir, Freundlich, Elovich, Temkin and Dubinin-Radushkevich. The comparison of linear
29 and nonlinear regression methods was applied by taking into consideration mathematical and
30 statistical properties of these two methods and compare different statistical criteria to determine
31 the best model fitted to experimental data by predicting the optimum isotherms parameters.
32 The following six error functions were used, the coefficient of determination (R^2), the sum of

33 the squares of the errors (SSE) , Hybrid fractional error function (HYBRID), Average relative
34 error (ARE) , Sum of the absolute errors (SAE) and Marquardt's percent standard deviation
35 (MPSD). Whereas, the comparison between different models shows that the nonlinear form of
36 Langmuir model was the best fitted to describe the equilibrium data which was confirmed by
37 the five error functions and the smallest 'sum of the normalized errors (SNE)' parameter.
38 Therefore, the kinetics data was evaluated by various models. It was found that the biosorption
39 process was conducted by the pseudo-second-order model which predicted best the
40 experimental data.

41 **Keywords:** *UV-Visible spectroscopy, Sodium diclofenac, Natural adsorbent, Adsorption, linear*
42 *and non-linear isotherm, anti-inflammatory.*

43 **1. Introduction**

44 Water is a vital resource for all living creatures. Unfortunately, It is threatened by pollution
45 that is becoming increasingly important, linked to the release of dangerous chemical
46 compounds by different industries: chemicals, textiles, tanneries, agribusiness,
47 pharmaceuticals ... etc.

48 Among these pollutants, many molecules come from pharmaceutical industries. Medicated
49 residues, rejected by industries, but also by hospitals, industrial farms (large consumers of
50 antibiotics) or domestic urine and stool. These residues are detected today at very low
51 concentrations (ng.L^{-1} or $\mu\text{g.L}^{-1}$) in aquatic environment thanks to analytical devices
52 performing. The revelation of the omnipresence of pharmaceutical chemicals (ketoprofen,
53 naproxen, ketotifen and diclofenac...) into the environment has given rise to a new concern,
54 due to their intrinsic biological activity, photodegradation sensitivity and inherent toxicological
55 effects [1]. In addition, Sodium Diclofenac (SD) contains several structural motifs, which are
56 photochemically active and also reactive toward photochemically generated oxidizing agents,
57 such as the hydroxyl radical and singlet oxygen.

58 Many water treatment techniques have been developed in recent years. These techniques
59 include chemical precipitation processes, flocculation, ion exchange, electrolysis, membrane
60 processes; however present dissatisfactory results because of the formation of residual toxic
61 by-products in addition to high treatment cost. Hence, many researchers shifted their interest
62 to adsorption technique due to their efficiency and cheaper cost and also the low cost of
63 biosorbents prepared from vegetal wastes using *Cyclamen persicum* [2], cocoa pod husks [3],
64 olive stones [4], Tea waste [5], cocoa shell [6] and potato peel waste [7] or from animal wastes,
65 such as chicken feathers [8], animal bones [9], *Ensis siliqua* Shells [10] and snail shells [11].
66 In the current work, the used biosorbent is the Calcined Cow Leather (CCL) which presents an
67 abundant slaughterhouse waste especially at Eid al-Adha holiday. Thus, the CCL biosorbent
68 was used to study the adsorption potential of pharmaceutical products. On the other hand, as to
69 describe well the equilibrium adsorption, the linear and nonlinear forms of two-parameter
70 isotherm models are frequently conducted to give the right description and characterization to
71 adsorption mechanism. However, the primary aim of this work is to shed more light on the
72 comparison of these two methods by taking into consideration mathematical and statistical
73 properties of these two methods and compare different statistical criteria to determine the best
74 model fitted to experimental data and isotherm parameters.

75 **2. Materials and methods**

76 **2.1. Material**

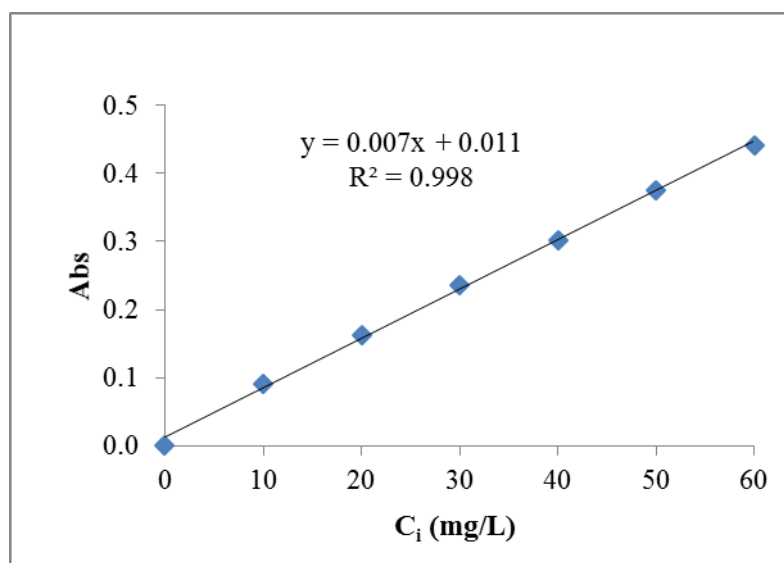
77 Cow leathers used in this work were collected previously from a slaughterhouse's management
78 waste service located in Casablanca city (Morocco), cleaned from blood and other dirt, and
79 salted immediately with common marine salt. Dried in open air for many days for partial
80 removal of water. Then, dried at 70 °C until reach a constant weight. Dried leathers were
81 calcined for 4 hours at 525 °C. The residue ground to fine powder, passed through sieves and
82 particles were collected under 200 µm and rinsed with deionized water until reached pH = 7,

83 then dried for 24 hours at 105 °C. The residual material was kept in a plastic container and
84 preserved in a desiccator for further use and the Calcined Cow Leathers were abbreviated
85 (CCL).

86

87 2.2.Preparation of Sodium Diclofenac (SD) standards

88 The synthetic stock solution (1 g.L⁻¹) used in these studies was prepared by dissolving 100 mg
89 of powdered sodium diclofenac (SD) (99% purity) in 100 mL deionized water. All desired test
90 solutions (0.0-60mg/L) were prepared from appropriate dilution of stock solution. The
91 calibration curve was depicted at this point, by plotting absorbance values vs. Initial
92 concentrations of SD standard solutions to verify Beer-Lambert's law (Figure 1), in order to
93 determine the areas of concentrations. The physic-chemical properties and chemical structure
94 of SD drug are shown in Table 1.



95

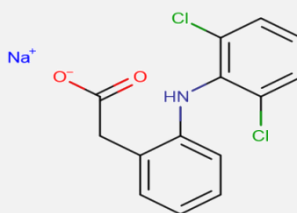
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97

Figure 1: Calibration curve of SD.

Table 1: Physicochemical properties and chemical structure of Sodium Diclofenac.

Sodium Diclofenac[12]	
Chemical formula	C ₁₄ H ₁₀ Cl ₂ NNaO ₂
Systematic name	sodium 2-{2-[(2,6-dichlorophenyl)amino] phenyl} acetate

Chemical structure

Molecular weight	318.1 g/mol
Use	Nonsteroidal anti-inflammatory drug and as an analgesic
Solubility in water (mg/L, 25°C)	0.00482 mg/mL
Log K_{ow} (at pH= 7)	1.37
pK_a (25°C)	4.2
λ_{max}	310 nm
Purity	99%

98

99 2.3.Physical and chemical CCL characterization

100 The biosorption efficiency of the activated biosorbent (CCL) is highly influenced by the
101 physical and chemical properties, for this purpose the prepared biosorbent (CCL) was
102 characterized for its physical characteristics by the BET method using a pore size micromeritics
103 ASAP 2020 apparatus (USA) to determine the specific surface area (BET). SEM analysis (FEI
104 Quanta 200 instrument (USA)) was performed to study the surface morphology of the
105 biosorbent (CCL). Furthermore, the chemical properties of the prepared biosorbent were
106 determined in terms of elemental composition analysis by Energy Dispersive X-rays
107 Spectroscopy (EDXS): X'Pert Pro MPD Panalytical (Netherlands) with Cu anode as the source
108 of X-rays at wavelength $\lambda=1.54 \text{ \AA}$ including ultimate (C, H, N and others) and XRF analysis
109 for CCL and also Fourier Transform Infrared Spectroscopy (FTIR) analysis was carried out in
110 the $450\text{-}4000\text{cm}^{-1}$ to determine the molecular structure of the functional groups and the
111 chemical bonds that exist on the sample of the prepared biosorbent under investigation using
112 an FT/IR-Vertex 70 spectrometer (Germany).

113

114 **2.4.Adsorption procedure**

115 The biosorption experiments were performed in batch system to study the effectiveness of SD
116 removal by taking into consideration the effects of the following parameters: biosorbent
117 dosage (0.05-0.2 g), initial concentration (10- 60 mg/L), contact time (15- 240min),
118 temperature (25, 35, 45 and 55°C), and pH (2- 10). Biosorption kinetics, isotherms and
119 thermodynamic are also studied in this work. For pH Adjustment, Hydrochloric acid (HCl,
120 37%, Sigma Aldrich) and sodium hydroxide solutions (NaOH, 36–38%, Merck) were used and
121 measured using the pH-Meter (METROHM 691). These samples were shaken with a magnetic
122 stirrer (IKA, RT 10 power) at 400 rpm with appropriate time and temperature then filtrated
123 using Fioroni filter paper, as final step the residual supernatants were analyzed using UV-
124 Visible spectrophotometer for all spectral measurements at 310 nm, Nitric acid 69–71% was
125 used with assay for SD [13] and all samples were performed in duplicate. The removal
126 percentage of SD from its solution was calculated as:

127

$$128 \qquad \%R = \frac{C_i - C_f}{C_i} \times 100 \qquad \text{(Eq.1)}$$

129

130 Where C_i is the initial concentration of SD (mg/L) and C_f is the final concentration (mg/L).

131 The metal uptake q (mg/g) was estimated following the equation [14]:

132

$$133 \qquad q_e = \frac{C_i - C_f}{m} \times V \qquad \text{(Eq.2)}$$

134

135 Where m is the quantity of biosorbent (g) and V is the volume of the suspension (L).

136

137 **2.5.Isotherm Modeling**

138 Many mathematical isotherm models have been developed. The most widely used are the
 139 Langmuir and Freundlich models because their mathematical expression is simple, and
 140 they make it possible to correctly represent the isothermal equilibrium adsorption in the
 141 aqueous phase in the majority of cases. Other models are more complex, but they have been
 142 developed more recently to describe the adsorbent-adsorbate interactions. Finally, those
 143 models are allowing us to predict the simultaneous adsorption of several compounds. The
 144 biosorption mechanism of CCL biosorbent on the active substance of Sodium Diclofenac (SD)
 145 in the present study was modeled by the Langmuir, Freundlich, Elovich, Temkin and Dubinin-
 146 Radushkevich (two-parameter) isotherm equations (Table 2).

147 **Table 2:** Two-parameter models studied in this study.

Isotherms	Nonlinear form	Linear form	Plot
Langmuir [15]	$\frac{q_e}{q_m} = \frac{K_L C_e}{1 + K_L C_e}$	$\frac{1}{q_e} = \frac{1}{C_e} \frac{1}{q_m K_L} + \frac{1}{q_m}$	$\frac{1}{q_e}$ vs $\frac{1}{C_e}$
Freundlich [16,17]	$q_e = K C_e^{1/n}$	$\log(q_e) = \log(K_F) + (1/n)\log(C_e)$	$\log(q_e)$ vs. $\log(C_e)$
Temkin [18]	$\frac{q_e}{q_m} = \frac{RT}{\Delta Q} \ln(K_T C_e)$	$q_e = B_T \ln K_T + B_T \ln C_e$ (with $B_T = \frac{q_m RT}{\Delta Q}$)	q_e vs. $\ln C_e$
Elovich [19]	$\frac{q_e}{q_m} = K_E C_e \exp(q_e/q_m)$	$\ln\left(\frac{q_e}{q_m}\right) = \ln(K_E C_e) + \frac{q_e}{q_m}$	$\ln\left(\frac{q_e}{q_m}\right)$ vs q_e
Dubinin– Radushkevich [20]	$q_e = q_m \exp(-\beta \varepsilon^2)$ (with $\varepsilon = RT \ln(1 + \frac{1}{q_e})$)	$\ln(q_e) = \ln(q_m) - \beta \varepsilon^2$	$\ln(q_e)$ vs ε^2

148

149 **2.6. Biosorption kinetics**

150 The biosorption kinetics provides the rate of biosorbate uptake onto the activated biosorbent
 151 within the equilibrium contact time. In this study, two models were implemented to evaluate

152 the rate constant of the biosorption process: The pseudo-first-order and pseudo-second-order
 153 kinetic models gathered in Table 3.

154 **Table 3:** Lagergren Kinetic models pseudo-first -order and pseudo-second-order forms

Kinetic model	Linear form	Plot
Pseudo-first-order [21]	$\text{Ln}(q_t - q_e) = \ln(q_e) - k_1 \cdot t$	$\text{Ln}(q_t - q_e)$ vs. t
Pseudo-second-order [22]	$t/q_t = 1/K_2q_2 + t/q_e$	t/q_t vs. t

155

156 2.7.Error analysis

157 Recently, linear regression has been one of the most viable tools defining the best fitting
 158 relationship quantifying the distribution of adsorbate, mathematically analyzing the adsorption
 159 systems. Non-linear optimization provides method for determining isotherm parameter values
 160 but still requires an error function assessment, in order to evaluate the fit of the isotherm to the
 161 experimental results. Since the choice of error function can affect the parameters derived. In
 162 this study six non-linear error functions (sum squares errors, hybrid fractional error function,
 163 Marquardt's percent standard deviation, average relative error, sum of absolute error and the
 164 coefficient of determination) (Table 4) were examined to evaluate the accuracy of each model
 165 isotherm equation to the experimental data (Table 2).The interpretation method chosen based
 166 on the simple plot ("Excel" software) allowing the linearization and evaluation of the
 167 parameters values for each model. Secondly, an interpretation using nonlinear regression
 168 algorithms using Satatistica version 10 software was tested to study nonlinear forms. Contrary
 169 to the linearization models, nonlinear regression usually involves the minimization or
 170 maximization of error distribution (between the experimental data and the predicted isotherm)
 171 based on its convergence criteria [23].

172 **Table 4:** List of error functions [24-28].

Error Function	Abbreviation	Definition
----------------	--------------	------------

Sum of quadratic gaps	ERRSQ / SSE	$\sum_{i=1}^n (q_{e,calc} - q_{e,mes})_i^2$
Hybrid fractional error function	HYBRID	$\frac{100}{n-p} \sum_{i=1}^n \left[\frac{(q_{e,meas} - q_{e,calc})}{q_{e,meas}} \right]_i$
Marquardt's percent standard deviation	MPSD	$100 \sqrt{\frac{1}{n-p} \sum_{i=1}^n \left(\frac{(q_{e,meas} - q_{e,calc})}{q_{e,meas}} \right)_i^2}$
Average relative error	ARE	$\frac{100}{n} \sum_{i=1}^n \left \frac{(q_{e,meas} - q_{e,calc})}{q_{e,meas}} \right _i$
Sum of the absolute errors	SAE	$\sum_{i=1}^n q_{e,calc} - q_{e,meas} _i$
The Coefficient of determination	R ²	$\frac{(q_{exp} - \overline{q_{cal}})^2}{\sum (q_{exp} - \overline{q_{cal}}) + (q_{exp} - q_{cal})^2}$

173

174 3. Results and discussion

175 In this study, five error functions were examined for isotherm equations and in each case a set
176 of isotherm parameters were determined by minimizing seven well-known error functions to
177 calculate the error deviation between experimental and predicted equilibrium adsorption data
178 of linear and nonlinear methods on the concentration range studied. The error functions studied
179 are mentioned previously on the precedent section. As known, each error function produces a
180 different set of isotherm parameters, due to the difficulty of a direct identification towards the
181 total optimal parameters. However, the parameter "SNE" is used in order to have a better
182 comparison between the sets of parameters for the same isothermal model [29-31].

183 The calculation method for the 'sum of the normalized errors' (SNE) is as follows:

- 184 (a) Select one isotherm and one error function and determine the isotherm parameters
185 that minimize the error function for that isotherm to produce the isotherm parameter set
186 for that error function.

- 187 (b) Determine the values for all the other error functions for that isotherm parameter set.
- 188 (c) Calculate all other parameter sets and all their associated error function values for that
189 isotherm.
- 190 (d) Select each error measure in turn and ratio the value of that error measure for a given
191 parameter set to the largest value of that error from all the parameter sets for that
192 isotherm.
- 193 (e) Sum all these normalized errors for each parameter set.

194 The smallest sum normalized error (SNE) is considered to be optimal for that isotherm
195 provided. Modeling of equilibrium isotherms according to the linear and non-linear models of
196 Langmuir, Freundlich, Temkin, Elovich and D-R was performed and the values of the five
197 error functions of the two-parameter isotherms investigated in this study are listed in Tables 5
198 and 6. Based on the R^2 coefficient values, the analysis of non-linear model of Langmuir is the
199 best fitted for SD using experimental data, since this model has the highest correlation
200 coefficient, the lowest values of errors functions and SNE value as well. The analysis of the
201 error function results confirms the model determined based on the R^2 correlation coefficient
202 values for the nonlinear regression, but disagrees with the model determined by the linear
203 regression, which confirms by comparing and studying all the resulted values of error functions
204 and taking into consideration the inherent distortion resulting from the linearization of non-
205 linear forms, that the parameters of the non-linear form cannot be solved from the linear
206 equation, so the nonlinear regression is more precise, efficient and gives exact results than
207 linear regression for the modeling of the adsorption isotherms. The highest R^2 and the lowest
208 SSE, HYBRID, MPSD, ARE, SAE and SNE are in bold.

209 **Table 5:** The linear and nonlinear isotherm models constants for SD biosorption onto CCL biosorbent.

Diclofenac

Langmuir

Linear form	$K_L = 0.152 \text{ L/mg}$	$q_m = 18.140 \text{ mg/g}$	$R_L = 0.397$
Nonlinear form	$K_L = 0.150 \text{ L/mg}$	$q_m = 18.678 \text{ mg/g}$	$R_L = 0.401$
Freundlich			
Linear form	$K_F = 2.270 \text{ mg}^{(1-n)} \text{ Ln g}^{-1}$		$n = 1.192$
Nonlinear form	$K_F = 2.495 \text{ mg}^{(1-n)} \text{ Ln g}^{-1}$		$n = 1.346$
Temkin			
Linear form	$K_T = 2.135 \text{ L/mg}$		$B_T = 3.203$
Nonlinear form	$K_T = 2.146 \text{ L/mg}$		$B_T = 1.388$
Elovich			
Linear form	$K_E = 0.135 \text{ L/mg}$		$q_m = 19.231 \text{ mg/g}$
Nonlinear form	$K_E = 0.150 \text{ L/mg}$		$q_m = 13.194 \text{ mg/g}$
Dubinin–Radushkevich			
Linear form	$q_m = 6.573 \text{ mg/g}$	$\beta = 1.10^{-6}$	$E = 707.11 \text{ kJ/mol}$
Nonlinear form	$q_m = 8.373 \text{ mg/g}$	$\beta = 2.10^{-6}$	$E = 500.501 \text{ kJ/mol}$

210 **Table 6:** Error functions values of the two-parameter isotherm models for the linear and nonlinear
211 forms of diclofenac (SD) biosorption on CCL

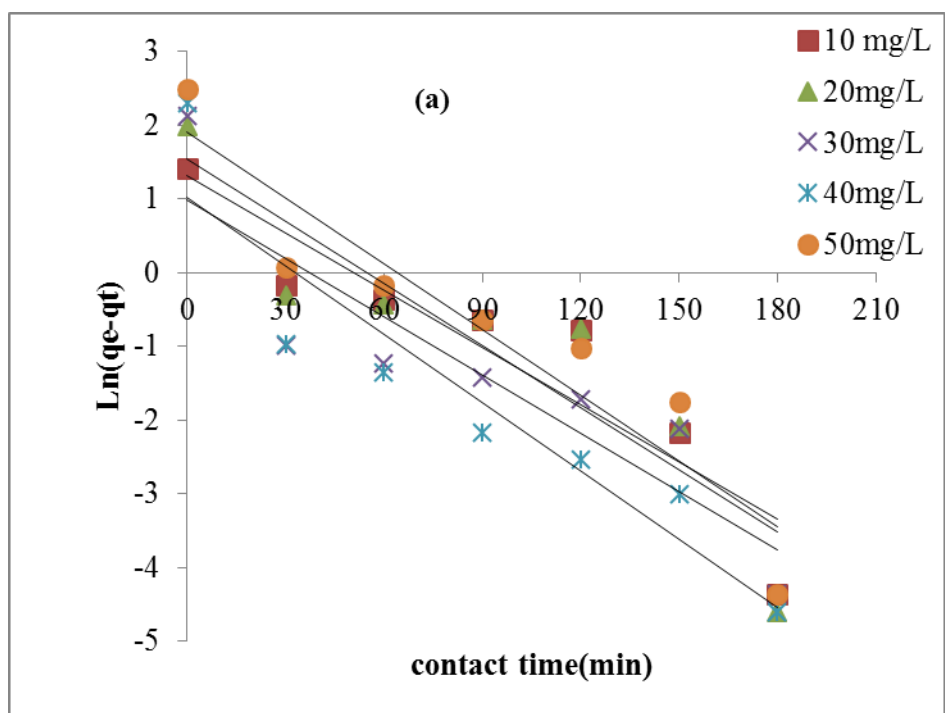
	Model form	R²	SSE	HYBRID	ARE	SAE	MPSD	SNE
Langmuir	Nonlinear	0.998	0.085	1.505	2.478	0.511	9.460	0.280
	Linear	0.773	17.620	154.620	28.899	9.333	75.571	5.000
Freundlich	Nonlinear	0.994	0.250	3.913	7.379	0.963	14.940	0.596
	Linear	0.990	0.552	3.839	7.210	1.415	10.375	0.595
Temkin	Nonlinear	0.996	0.203	1.864	5.582	0.964	8.762	0.436
	Linear	0.996	0.193	1.749	5.369	0.934	8.451	0.420
Elovich	Nonlinear	0.997	0.151	2.611	5.800	0.676	12.413	1.064
	Linear	0.993	0.376	2.823	6.517	1.238	9.345	0.521
R-D	Nonlinear	0.932	3.202	25.142	19.880	3.825	27.367	1.804
	Linear	0.881	3.964	24.631	15.702	3.648	23.484	1.629

212

213 Biosorption kinetics

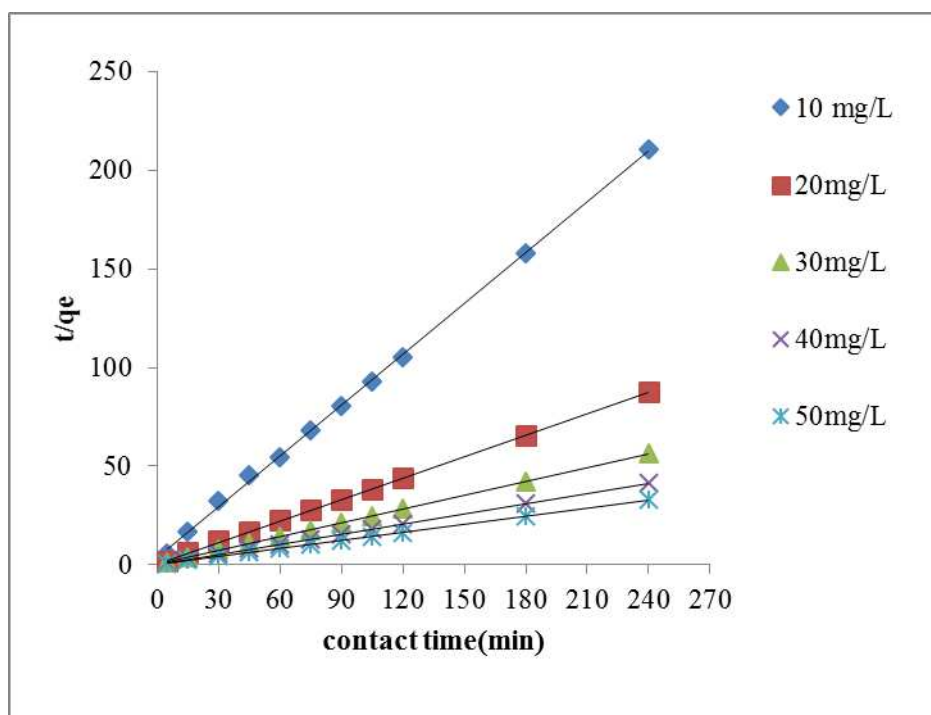
214 Figures 2 illustrates the plot of $\ln(q_e - qt)$ versus time (t) for the pseudo-first-order of SD on
215 CCL at different concentrations. The values of $K_{ads,1}$ can be predicted and $q_{e,cal}$ can be also
216 determined from the intercept and the slope when $t=0$. However, Figure 3 shows the plot of
217 t/qt against t (minutes) which gives $1/q_{e(cal)}$ as the slope and $1/k_2 q_e^2$ as the intercept, where k_2
218 ($\text{g} \cdot \text{mg}^{-1} \cdot \text{min}^{-1}$) is the rate constant of the second-order biosorption.

219 Figure 2 shows the nonlinearity of the plot of $\ln(q_e - q_t)$ as a function of time (t). This indicates
220 that the biosorption kinetics of diclofenac is not controlled by the first model. Whereas, Figure
221 3 indicates that physical adsorption plays an important role in SD adsorption onto CCL.
222 Overall, the pseudo-second-order kinetic model fits the experimental data better than the
223 pseudo-first-order kinetic model as it correlates the experimental data well for all
224 concentrations for the selected active substance (SD) under investigation.



225

226 **Figure 2:** Linearized plots for Pseudo-first-order kinetics for biosorption of sodium diclofenac (SD)
227 onto CCL biosorbent.



228
229 **Figure 3:** Linearized plots for Pseudo-second-order kinetics for biosorption of sodium diclofenac
230 (SD) onto CCL biosorbent.

231 It is observed that the regression lines are almost superimposed by the experimental data. All
232 of the kinetic parameters determined from these figures are expressed in Table 7, which
233 indicate that the values of the correlation coefficient (R^2) for the linear model of the pseudo-
234 first-order are far from unity. The values of $q_{e,exp}$ are incomparable with those of $q_{e,cal}$, which
235 confirm that the system is not well fitted to the pseudo-first-order equation which is inadequate
236 to obtain good adsorption kinetics. Accordingly, the pseudo-second-order equation was
237 applied. Therefore, all the R^2 values obtained from the pseudo-second-order model are closer
238 to unity, indicating that the biosorption of SD on the activated biosorbent studied fits the model
239 well. Moreover, the experimental $q_{e,exp}(mg/g)$ values agreed satisfactorily with the calculated
240 values $q_{e,cal}(mg/g)$. The pseudo-second-order kinetic model is based on the assumption that
241 chemical adsorption is the rate controlling step. It can predict the behavior over the whole range
242 of contact period [32-34]. This suggests that the overall rate of the adsorption process studied

243 here is controlled by chemisorption which proceeds by the exchange or sharing of valance
 244 electrons between the biosorbate and biosorbent.

245 **Table 7:** Lagergren pseudo-first-order and pseudo-second-order kinetics biosorption parameters for SD
 246 biosorption onto CCL at 25 °C.

247

	C_e (mg/L)	Pseudo-First order				Pseudo-Second order		
		$q_{e,exp}$ (mg/g)	$q_{e,cal}$ (mg/g)	$K_{ads,1}$ (min^{-1})	R^2	$q_{e,cal}$ (mg/g)	$K_{ads,2}$ (min^{-1})	R^2
Diclofenac	10mg/L	1.143	0.311	0.019	0.792	1.168	0.188	0.999
	20mg/L	2.738	0.475	0.020	0.742	2.770	0.184	0.999
	30mg/L	4.238	0.516	0.022	0.717	4.292	0.142	0.999
	40mg/L	5.786	0.807	0.022	0.784	5.848	0.102	0.999
	50mg/L	7.310	0.855	0.025	0.841	7.407	0.077	0.999

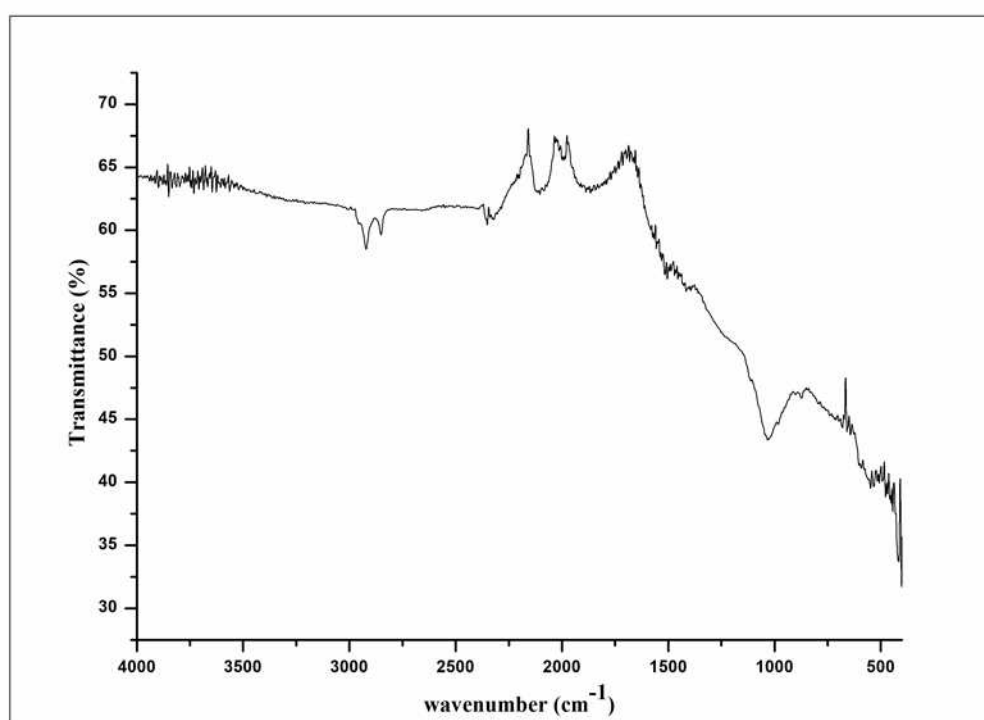
248

249 **Physical and chemical characterization**

250 The chemical composition is determined by X-ray fluorescence. The elemental analysis of CCL
 251 shows an abundant presence of oxygen (65.5%) and carbon (21.5%) followed up by small
 252 amounts of other elements such as calcium (2.71%), sodium (2.59%), phosphorus (2.01),
 253 chlorine (1.71), magnesium (1.32%), silicon (1.08%), sulfur (0.728%), aluminum (0.302%),
 254 iron (0.252%), potassium (0.142%), zinc (0.029%), strontium (0.019%), titanium (0.015%),
 255 iodine (0.009%), bromine (0.005%) and copper (0.004%).

256 The activated carbon (CCL) was submitted to IR spectroscopy as depicted in figure 4. The two
 257 closest and acute bands around 2800-2900 cm^{-1} appear in the spectrum correspond to
 258 asymmetric and symmetric C-H stretching vibrations of aliphatic groups, -CH₃ and -CH₂. The
 259 closest peaks at 2300-2400 cm^{-1} are found on the spectrum of the activated carbon prepared

260 from calcined cow leather which shows the presence of carboxylic acid, -COOH functional
261 group. The overlapped bands at the 1500-1650 cm^{-1} region are due to C=C stretching vibrations
262 in sp^2 hybridized carbons in polyaromatics rings. The strong band at 1000-1150 cm^{-1} region is
263 a very typical absorption for activated carbons in general and has been ascribed to either a C-
264 O stretching vibration or to a Si-O stretching vibration mineral matter contained in the carbon
265 (silicates), which may be due to inorganic silicates in the calcined cow leather sample. The
266 peaks detected at 1000-1100 cm^{-1} in the spectrum of the CCL biosorbent reveals the presence
267 of C-O-C stretching vibrations of esters, ether or phenol groups whereas the weak to medium
268 peaks located at 500-800 cm^{-1} are assigned for C-H out-of plane bending of benzene derivatives,
269 O-H stretching vibrations of C-O-H band. The results obtained agreed with the previous
270 research where C-H out-of plane bending vibration for benzene derivatives were found on the
271 surface of various activated carbons [35], whereas O-H functional group has been found on the
272 surface of most of the activated carbons, including the commercial grade activated carbons
273 [36].



275

Figure 4: IR spectrum of biosorbent (CCL)

276

The XRD analysis was carried out to determine the structure of the prepared activated carbon.

277

The absence of a sharp peak reveals a predominantly amorphous structure for CCL, which

278

present an important property for an excellent biosorbent [37]. As presented in figure 5, the

279

intensity of the diffraction peaks mainly observed in the range $2\theta = 20-80^\circ$ is

280

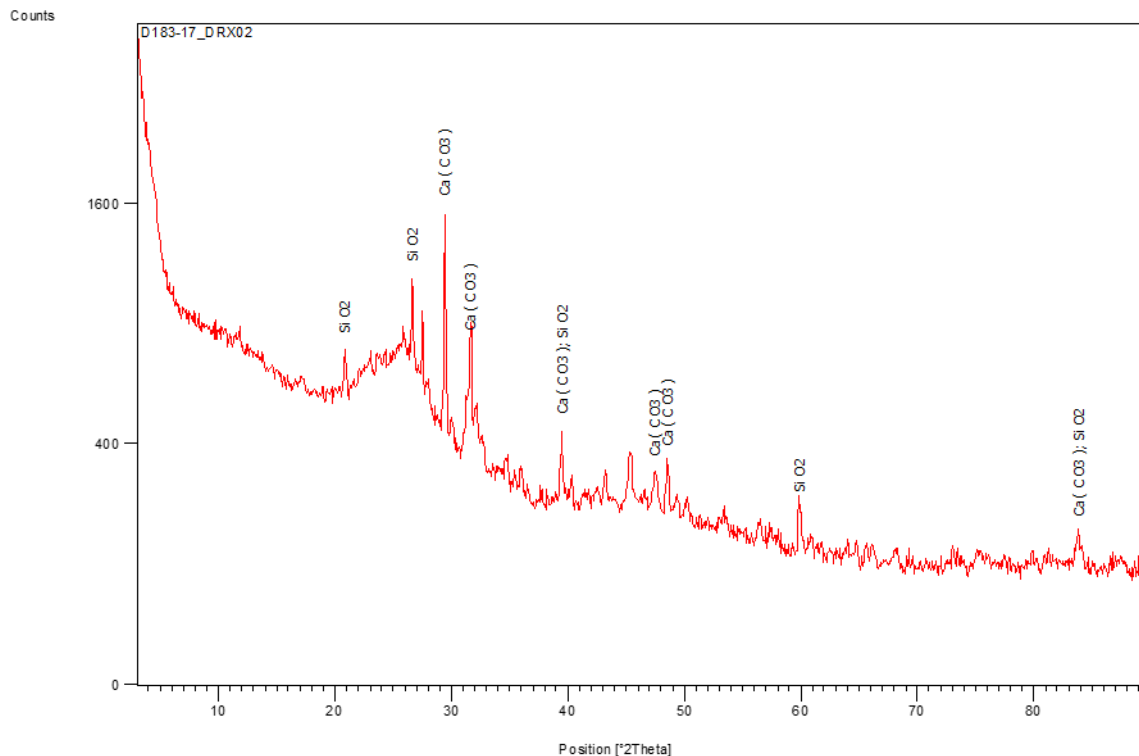
20.8673° , 26.6418° , 27.4364° , 29.4410° , 31.5664° , 39.4772° , 45.3652° , 47.5004° , 48.5601° , 59.8923° ,

281

and 83.8580° , which confirm the presence of two compounds :calcite (CaCO_3) and silicate

282

(SiO_2).



283

284

Figure 5: Powder XRD Pattern of the CCL biosorbent.

285

Scanning electron microscopy was performed to measure the morphological features of the

286

CCL biosorbent. From the obtained SEM micrographs, the change in surface texture and pore

287

development is clearly visible, whereas the biosorbent has a large distribution of grains of

288

different sizes. As shown by figure 6, the pores available on the surface of the CCL biosorbent

289 are visible, well pronounced with distinct and smooth pore walls, due to grinding and several
290 times washing with distilled water. However, the structure looks like amorphous silica after
291 carbonization, which is a highly reactive species and can form surface complexes with the
292 cations under investigation.

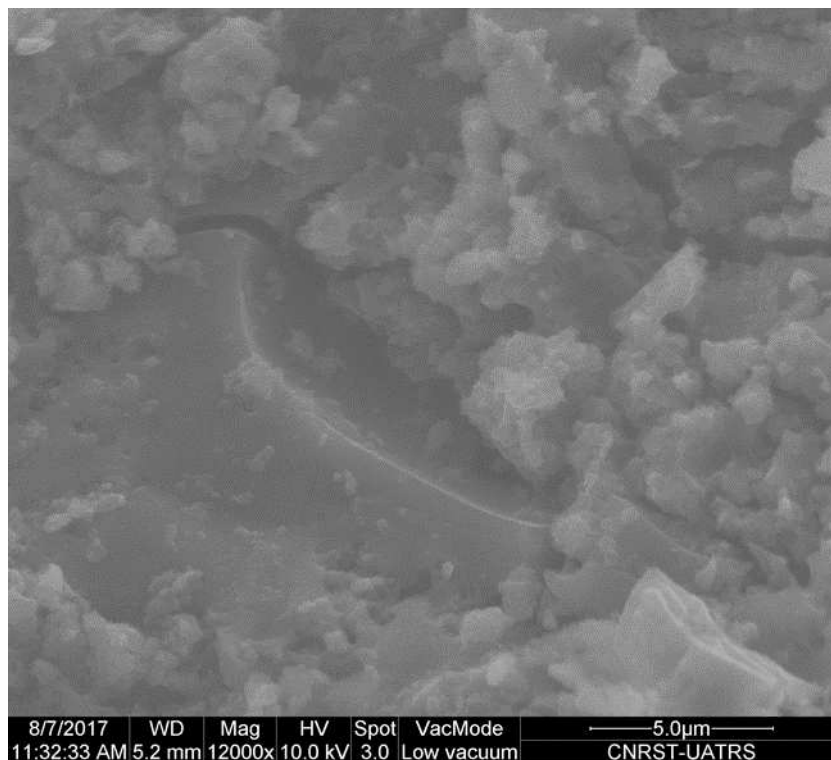


Figure 6: SEM micrograph of CCL biosorbent.

295 The specific surface area of CCL was determined by Brunauer-Emmett-Teller (BET)
296 multipoint technique by adsorption-desorption isotherms of nitrogen at its liquid temperature
297 (77K) and was found $S_p = 15 \text{ m}^2/\text{g}$. The specific surface area (BET) is outstanding by
298 comparing it with that of other biosorbents derived from a natural origin.

299 **Conclusion**

300 In this work, the prime objective is the valorization of the slaughterhouse wastes and the use
301 of those raw natural materials for the preparation of low cost biosorbents with good textural
302 and structural properties. Therefore, it was crucial for this study to identify informations
303 concerning the steps of adsorbent preparation, the textural and structural characteristics of the

304 prepared biosorbent as well as those relating to their adsorption capacities of various inorganic
305 pollutants. For this purpose, the biosorbent was prepared from calcined cow leather (CCL) and
306 have been used successfully as an adsorbing agent for the removal of active pharmaceutical
307 ingredient (Sodium Diclofenac) from aqueous solutions. After activation, we carried out
308 characterization tests by physico-chemical methods (FX, DRX, IRTF and SEM) and we found
309 that CCL is essentially composed of oxygen and carbon which gives it a structure of calcite
310 (CaCO_3) and silicate (SiO_2). With the increase of concentration from 10mg/L to 50mg/L, the
311 removal efficiency increases from 68.57% to 87.71% for diclofenac sodium (SD) reaching the
312 equilibrium at 20mg/L. The kinetic model fits the experimental data of the pseudo-second-
313 order better than the pseudo-first-order model, as it correlates the experimental data well for
314 all concentrations for the anti-inflammatory under investigation. Modelling of equilibrium
315 isotherms according to the linear and non-linear models of Langmuir, Freundlich, Temkin,
316 Elovich and D-R was performed. Based on the R^2 coefficient values, the analysis of non-linear
317 and linear forms for SD shows that the non-linear form of Langmuir model has a better R^2
318 correlation coefficient, while the linear model of Temkin has a high R^2 correlation coefficient
319 for the adsorption of SD. The analysis of the error function results confirms the model
320 determined based on the R^2 correlation coefficient values for the nonlinear regression, but
321 disagrees with the model determined by the linear regression, which confirms that the nonlinear
322 regression is more accurate and gives exact results than the linear regression for the modeling
323 of the adsorption isotherms.

324

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330 The data used to support the findings of this study are available from the corresponding author
331 upon request.

332 **Competing interests**

333 The authors declare they have no competing interests.

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337 **Authors' contributions**

338 **H. HIYANE:** Writing - original draft, Software, Investigation. **S. BENKADDOUR:**
339 Investigation. **R. SLIMANI:** Writing - review & editing. **B. CAGNON:** review & editing,
340 Supervision. **K. KARROUCHI:** review & editing, Supervision. **M. EL HADDAD:** Writing -
341 review & editing. **Y.ACHOUR:** Writing - review & editing, Supervision. **S. EL ANTRI:**
342 Writing - review & editing, Supervision. **S. LAZAR:** Writing - review & editing, Supervision.

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Figures

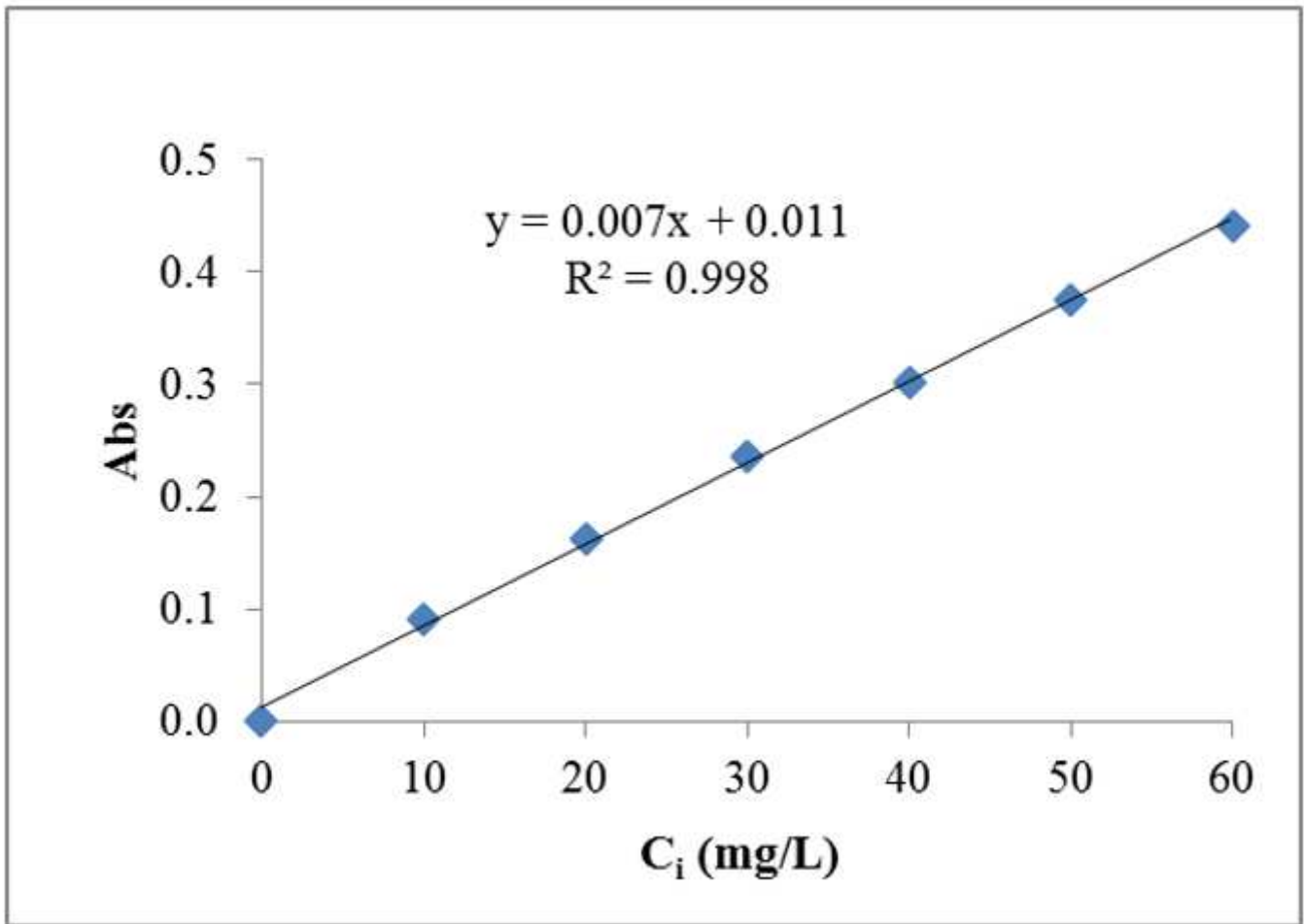


Figure 1

Calibration curve of SD.

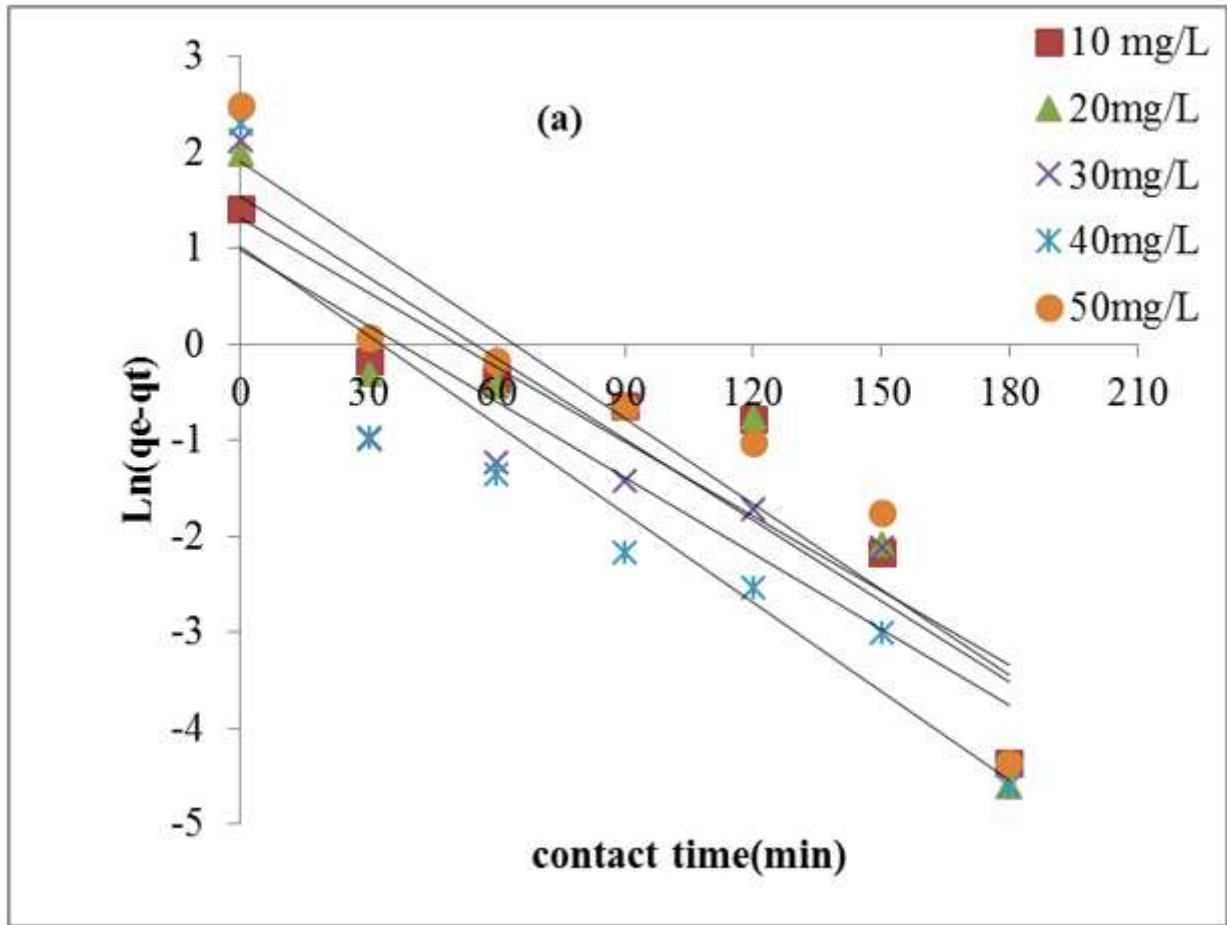


Figure 2

Linearized plots for Pseudo-first-order kinetics for biosorption of sodium diclofenac (SD) onto CCL biosorbent.

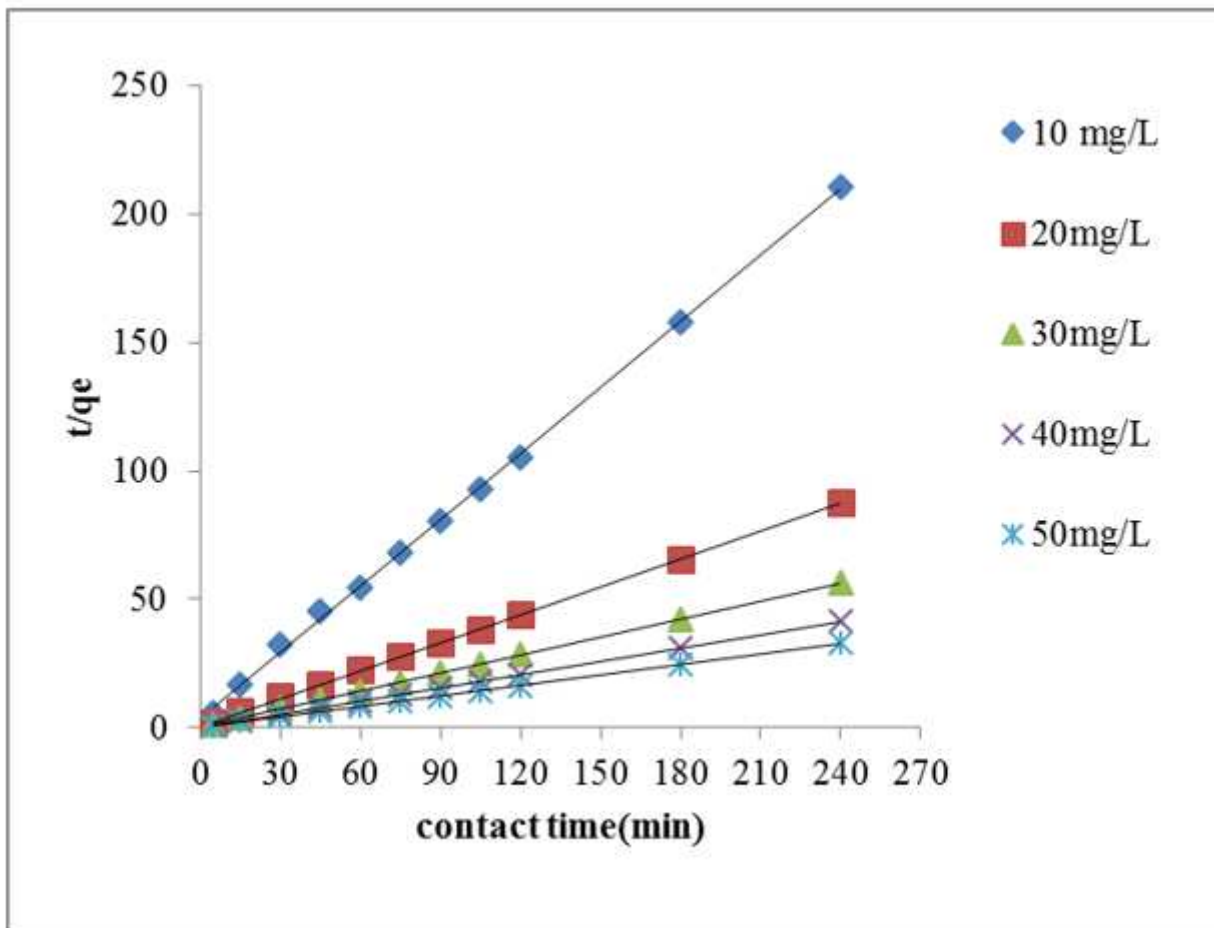


Figure 3

Linearized plots for Pseudo-second-order kinetics for biosorption of sodium diclofenac (SD) onto CCL biosorbent.

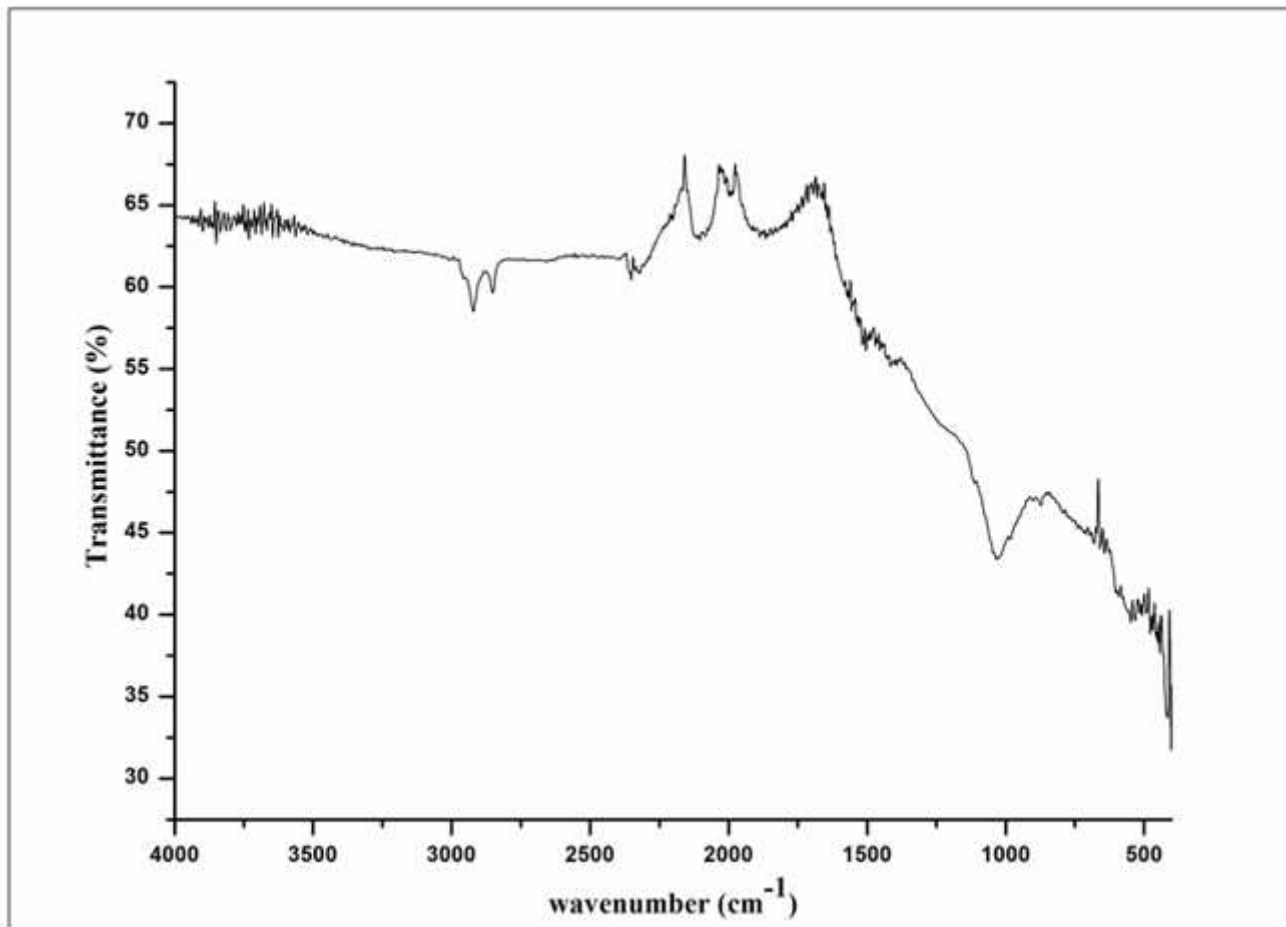


Figure 4

IR spectrum of biosorbent (CCL)

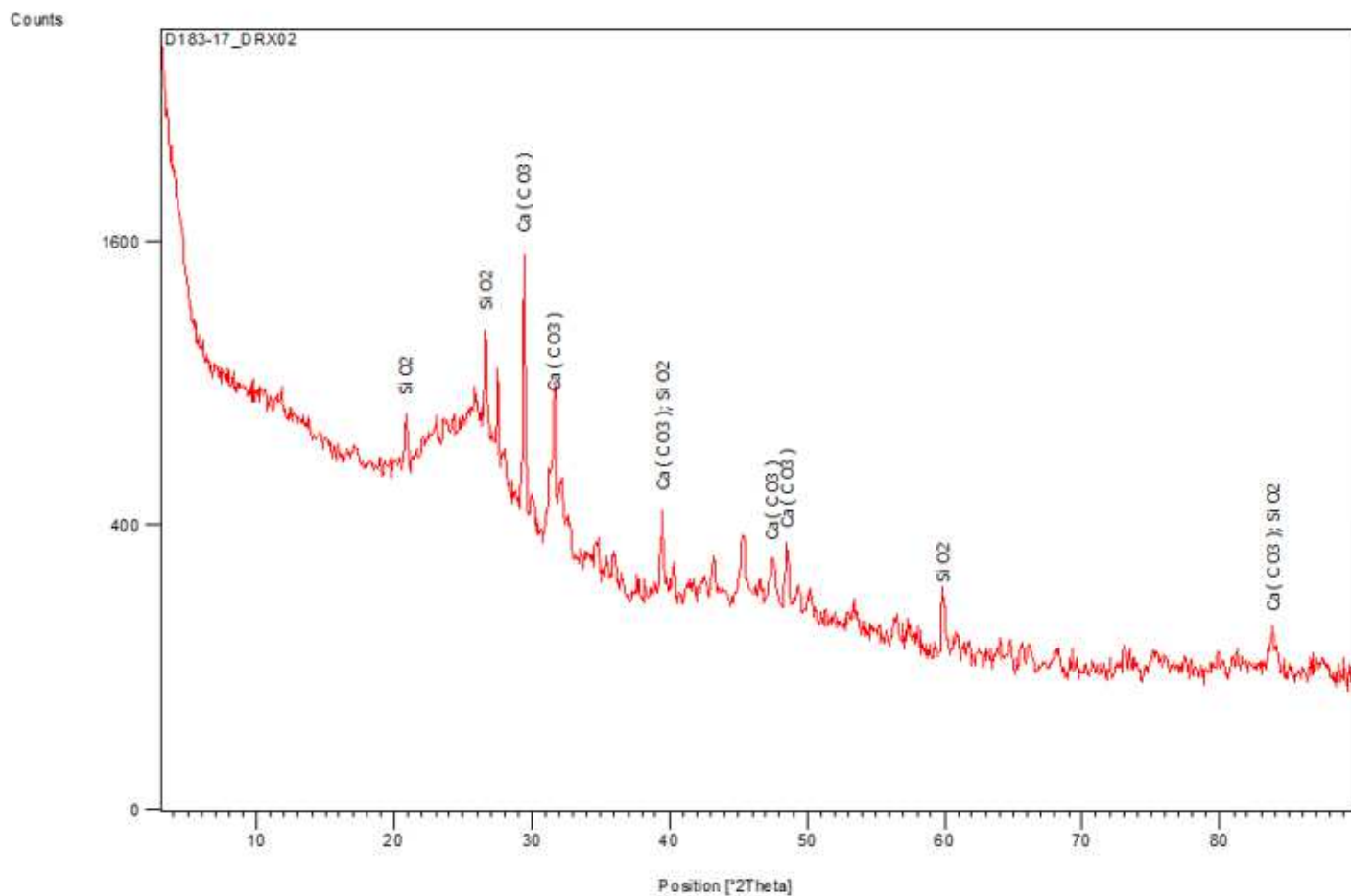


Figure 5

Powder XRD Pattern of the CCL biosorbent.

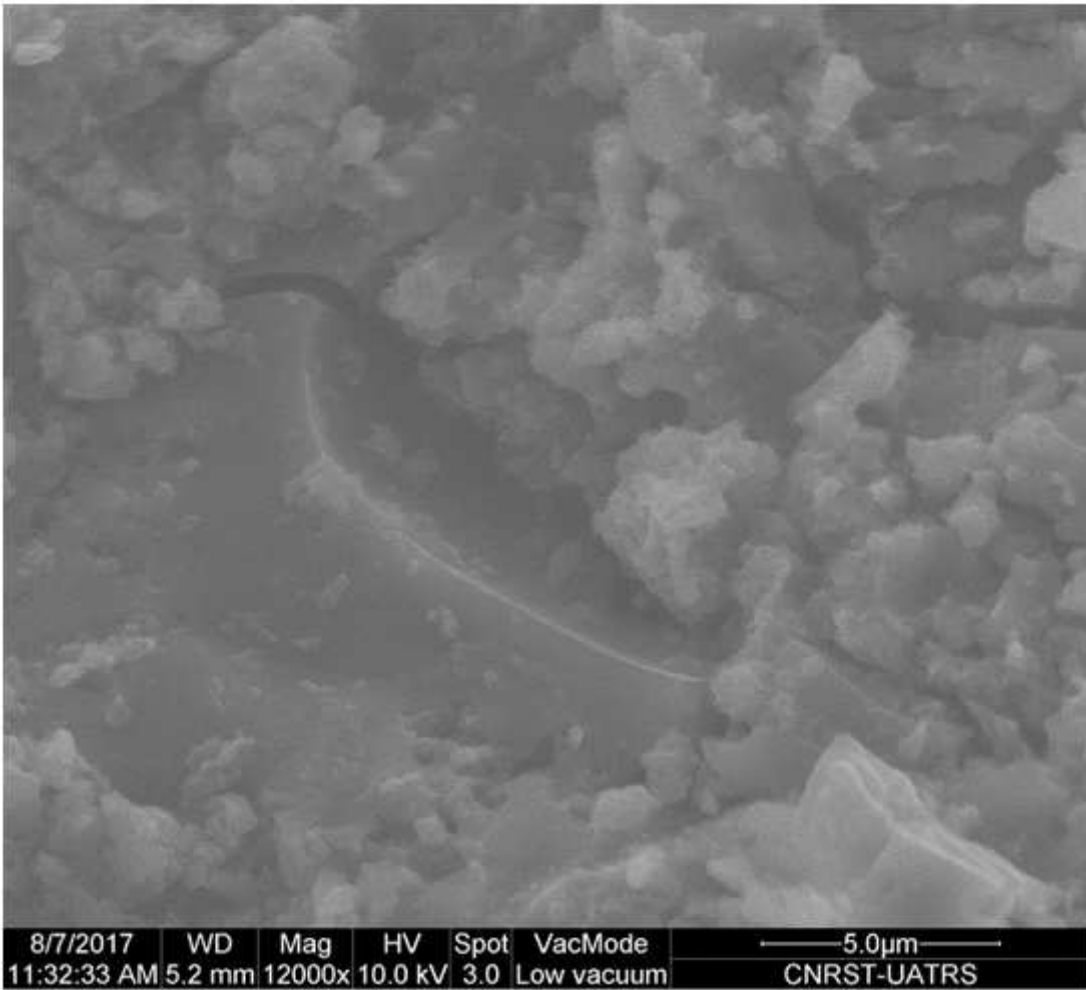


Figure 6

SEM micrograph of CCL biosorbent.