

Unaccounted natural vegetation loss in Brazilian Amazon

Cassiano Messias (✉ cassiano.messias@inpe.br)

National Institute for Space Research (INPE) <https://orcid.org/0000-0002-1497-1022>

Claudio Almeida

INPE - National Institute for Space Research

Daniel Silva

World Wide Fund for Nature (WWF)

Luciana Soler

INPE - National Institute for Space Research

Luiz Maurano

National Institute for Space Research (INPE)

Vagner Camilotti

National Institute for Space Research (INPE) <https://orcid.org/0000-0002-5251-3612>

Fábio Alves

Federal University of Western Bahia (UFOB)

Libério Silva

National Institute for Space Research (INPE)

Thiago Lima

National Institute for Space Research (INPE) <https://orcid.org/0000-0002-4968-2036>

Vivian Renó

National Institute for Space Research (INPE) <https://orcid.org/0000-0002-1929-8092>

Deborah Correia-Lima

National Institute for Space Research (INPE) <https://orcid.org/0000-0001-5889-1940>

Amanda Belluzzo

National Institute for Space Research (INPE) <https://orcid.org/0000-0002-5098-1316>

Camila Quadros

National Institute for Space Research (INPE) <https://orcid.org/0000-0002-2619-6596>

Delmina Barradas

National Institute for Space Research (INPE)

Douglas Moraes

National Institute for Space Research (INPE) <https://orcid.org/0000-0003-1488-0363>

Eduardo Marcelino Bastos

National Institute for Space Research (INPE) <https://orcid.org/0009-0000-3423-2799>

Igor Cunha

National Institute for Space Research (INPE) <https://orcid.org/0009-0004-5528-2620>

Jefferson Souza

National Institute for Space Research (INPE) <https://orcid.org/0000-0002-9473-8430>

Lucélia Barros

National Institute for Space Research (INPE) <https://orcid.org/0000-0002-0837-3544>

Luiz Gusmão

National Institute for Space Research (INPE)

Rodrigo de Almeida

National Institute for Space Research (INPE)

Dayane Moraes

National Institute for Space Research (INPE) <https://orcid.org/0009-0003-8617-5054>

Diego Silva

National Institute for Space Research (INPE) <https://orcid.org/0000-0001-7670-4408>

Eduardo Chrispim

National Institute for Space Research (INPE) <https://orcid.org/0000-0002-0830-6288>

Manoel Rodrigues Neto

National Institute for Space Research (INPE) <https://orcid.org/0000-0003-3033-2594>

Marlon Matos

National Institute for Space Research (INPE) <https://orcid.org/0009-0009-6968-3501>

Noeli Moreira

National Institute for Space Research (INPE) <https://orcid.org/0000-0002-5308-8080>

Raíssa Caroline Dos Santos Teixeira

National Institute for Space Research (INPE) <https://orcid.org/0009-0003-0755-7864>

Gabriel Alves

National Institute for Space Research (INPE)

Ana Carolina Andrade

National Institute for Space Research (INPE) <https://orcid.org/0009-0003-9377-3802>

Letícia Perez

National Institute for Space Research (INPE) <https://orcid.org/0000-0002-6784-3964>

Mariane Reis

National Institute for Space Research (INPE)

Bruna Bento

National Institute for Space Research (INPE)

Hugo Castro Filho

National Institute for Space Research (INPE)

Igor Santos

National Institute for Space Research (INPE)

Liliane Araújo

National Institute for Space Research (INPE)

Maira Matias

National Institute for Space Research (INPE)

Murilo Silva

National Institute for Space Research (INPE)

Fábio Pinheiro

National Institute for Space Research (INPE)

André Carvalho

National Institute for Space Research (INPE) <https://orcid.org/0000-0002-7893-6672>

Haron Xaud

Brazilian Agricultural Research Corporation (EMBRAPA Roraima)

Maristela Xaud

Brazilian Agricultural Research Corporation (EMBRAPA Roraima)

Ana Paula Matos

Image Processing and Geoprocessing Laboratory (LAPIG)

Luis Baumann

Image Processing and Geoprocessing Laboratory (LAPIG) <https://orcid.org/0000-0001-5308-9721>

Elaine Silva

Image Processing and Geoprocessing Laboratory (LAPIG)

Laerte Ferreira

Image Processing and Geoprocessing Laboratory (LAPIG)

João Pinto

Instituto Nacional de Pesquisas Espaciais (INPE)

Marcos Adami

Instituto Nacional de Pesquisas Espaciais (INPE) <https://orcid.org/0000-0003-4247-4477>

Article**Keywords:**

Posted Date: November 1st, 2023

DOI: <https://doi.org/10.21203/rs.3.rs-3405875/v1>

License:   This work is licensed under a Creative Commons Attribution 4.0 International License.

[Read Full License](#)

Additional Declarations: There is **NO** Competing Interest.

1 Unaccounted natural vegetation loss in Brazilian Amazon

2

3 *Cassiano Gustavo Messias^{1*}, Cláudio A. de Almeida¹, Daniel E Silva^{1,2}, Luciana Soler¹,*
4 *Luiz E. Maurano¹, Vagner Luis Camilotti¹, Fábio C. Alves^{1,3}, Libério J. da Silva¹,*
5 *Mariane Reis¹, Thiago C. de Lima¹, Vivian Renó¹, Deborah L. C. Lima¹, Amanda P.*
6 *Belluzzo¹, Camila B. Quadros¹, Delmina Carla M. Barradas¹, Douglas Rafael V. de*
7 *Moraes¹, Eduardo Felipe M. Bastos¹, Igor P. Cunha¹, Jefferson J. de Souza¹, Lucélia S.*
8 *de Barros¹, Luiz Henrique A. Gusmão¹, Rodrigo de Almeida¹, Dayane Rafaela V. de*
9 *Moraes¹, Diego M. Silva¹, Eduardo Henrique S. Chripim¹, João Felipe S. K. C. Pinto¹,*
10 *Manoel R. Ribeiro Neto¹, Marlon Henrique H. Matos¹, Noeli Aline P. Moreira¹, Raíssa*
11 *Caroline dos S. Teixeira¹, Gabriel M. R. Alves¹, Ana Carolina Santos¹, Letícia P.*
12 *Perez¹, Bruna Maria P. Bento¹, Hugo C. de Castro Filho¹, Igor S. dos Santos¹, Liliane*
13 *Cristina L. de Araújo¹, Maira Matias¹, Murilo B. da Silva¹, Fábio da C. Pinheiro¹,*
14 *André Carvalho¹, Haron Xaud⁴, Maristela Xaud⁴, Ana Paula Matos⁵, Luis Baumann⁵,*
15 *Elaine Barbosa da Silva⁵, Laerte Guimarães Ferreira⁵, Marcos Adami¹*

16

17 ¹General Coordination for Land Sciences (CGCT), National Institute for Space Research (INPE), Av. dos
18 Astronautas, 1758, Jardim da Granja, São José dos Campos, SP, Brazil, 12227-010

19 ²WWF Brasil CLS 114 Bloco D - 35 - Asa Sul, Brasília, DF, Brasil 70377-540

20 ³Center for Humanities - CEHU, Federal University of Western Bahia, R. da Prainha, n°. 1326, Morada
21 Nobre, Barreiras, BA, Brazil, 47810-047

22 ⁴Brazilian Agricultural Research Corporation (EMBRAPA), Rodovia BR 174, km 8, Boa Vista, RR,
23 Brazil, 69301-970

24 ⁵Image Processing and Geoprocessing Laboratory (LAPIG), Campus Samambaia, Alameda Palmeiras,
25 s/n, Chácaras Califórnia, Goiânia, GO, Brazil, 74001-970

26

27 *Correspondence to: cassiano.messias@inpe.br

28

29 **Abstract**

30 **Deforestation in the Brazilian Amazon has been monitored since 1988 by the**
31 **Brazilian Amazon Satellite Monitoring Program (PRODES Amazonia), and its**
32 **data has been pivotal in guiding environmental public policies in the country.**
33 **While forest formations are officially supported by a monitoring program, a**
34 **significant portion of the Amazon biome (6.6 % or ~280,000 km²) constituted by**
35 **non-forest (NF) phytophysionomies (e.g., savanna, grasslands, flood lands) are**
36 **still unmonitored. To address this information gap, the PRODES NF system was**
37 **built and adapted from the well-established and recognized methodology of**
38 **PRODES Amazonia in Brazil. First findings based on PRODES NF monitoring**
39 **indicate that the Brazilian Amazon lost 10.46% (~30,000 km²) of NF area, mostly**
40 **in the last two decades. The states of Mato Grosso, Roraima and Amapá emerged**
41 **as the primary hotspots of losses, with a growing trend of losses for the last two**
42 **states. Among the phytophysionomies, savannas were the most affected (13.3% of**
43 **their extent). A strong correlation between NF loss and deforestation was revealed**
44 **in the Amazon biome, with no statistical differences in terms of relative area,**
45 **suggesting a continuum of vegetation loss along this biome that does not**
46 **discriminate between forest and non-forest. Finally, PRODES Amazonia and**
47 **PRODES NF together provide relevant official data that sum up a total of**
48 **vegetation loss of ~798.000 km² in the Brazilian Amazon (~19% of the entire**
49 **biome).**

50

51 Forest loss in Brazilian Amazon have been continuously monitored since 1988
52 through the Brazilian Amazon Satellite Monitoring Program (PRODES Amazonia)¹.
53 PRODES data is internationally recognized as a crucial tool to assess and control the
54 extent and rate of deforestation processes, being significant for public policy proposals
55 and enforcement, as well as research on varied topics that include biodiversity, climate
56 change, and human well-being². PRODES Amazonia solely focuses on forest clear-
57 cutting, while companion projects like TerraClass and DETER³ complement it by
58 providing data on land use and land cover changes in the Brazilian Amazon. This includes
59 information on forest regrowth, primary land use classes, and the detection of smaller
60 deforestation or forest degradation events in near real-time throughout the year.
61 Throughout the PRODES Amazonia data series, however, a considerable challenge
62 persisted in addressing the need to map a consistent historical series of natural non-forest
63 vegetation (NF) loss across an area spanning 279,492.08 km², equivalent to 6.6% of the
64 Amazon biome.

65 NF stands for natural vegetation other than strictly forest ecosystems and
66 embraces different types of phytophysionomies. In the Brazilian Amazon, NF occurs as
67 open-like formations such as savannas and grasslands; seasonally flooded areas with
68 sandy soils and sparse trees; ecotones; isolated forest patches with deciduous, semi-
69 deciduous, and even broadleaf characteristics; and natural areas of bare lands⁴ (Fig. S3).
70 These landscape features receive names such as pioneer formations, ecological refuges,
71 *lavrados*, *campinas* and *campinaranas* or white-sand ecosystems⁴. Despite the lack of
72 knowledge about their functioning and ecology⁵⁻⁷, NF ecosystems are important sites for
73 biodiversity conservation with endemic species of different taxa^{5,7-10}.

74 Previous mapping attempts¹¹⁻¹³ have shown the expansion of human activities
75 within NF formations, leading to relatively high percentages (~17%) of accumulated
76 deforestation in NF areas up to 2021 in selected Amazonian municipalities¹³. The overall
77 extent of NF losses in the Brazilian Amazon biome and their specific locations were still
78 unknown, which prevented the assessment of the impacts on Amazonian ecosystems and
79 understanding on the drivers behind this destruction.

80 To tackle this challenge, we have developed PRODES NF, a systematic
81 monitoring system for NF in the Brazilian Amazon. PRODES NF, an adaptation of the
82 PRODES Amazonia forest mapping methodology, utilizes multi-sensor satellite imagery
83 to identify vegetation loss in predominantly open ecosystems. Integrated into the
84 PRODES monitoring system, it ensures continuous mapping of natural vegetation loss

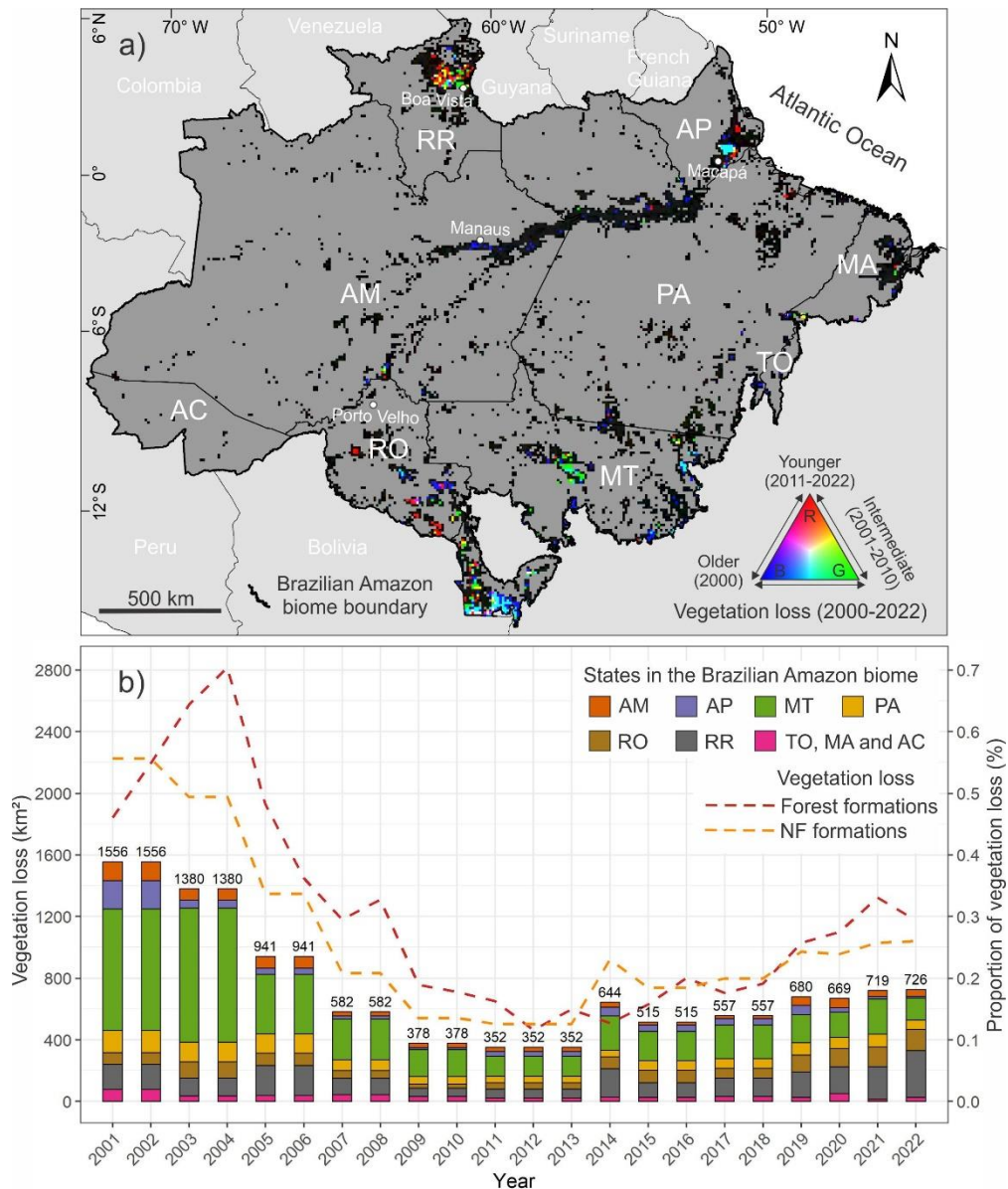
85 throughout the entire Brazilian Amazon biome. This system holds significant potential to
86 support various compliance initiatives, including REDD+, the National Inventory
87 (LULUCF), Trading Forest Certificates (CRA), and corporate commitments to reduce
88 deforestation and ecosystem conversion¹⁴⁻¹⁶. In this study, we present, for the first time
89 using official data, the spatial and temporal distribution of NF vegetation loss (referred to
90 as NF loss) in the Brazilian Amazon biome, concluding with an overview of the total
91 extent of vegetation loss, encompassing both forests and non-forest ecosystems.

92

93 **Non-forest loss in the Brazilian Amazon**

94 **NF loss hotspots**

95 Accumulated NF losses in the Brazilian Amazon biome until 2000 (baseline map)
96 accounted for 12,934.75 km² (4.63% of the total NF area, Fig. 1) and reached 29,247.44
97 km² (10.46%) up to 2022, meaning that more than half of the accumulated loss happened
98 in the last two decades. The prevailing pattern of NF loss up to 2022 unfolded from the
99 southern to the northern regions (Fig. 1a), evidencing three main hotspots of NF loss
100 located in the states of Mato Grosso, Roraima, and Amapá. Substantial and earlier losses
101 (pre-2000) were primarily evident in the southwest sector of Mato Grosso (Fig. 1a, blue
102 color), with additional minor occurrences in isolated zones within the central-eastern of
103 Rondônia. The loss of NF persisted in Mato Grosso during 2001-2010 (cyan) and starts
104 to be visible in the southeastern sector of Amapá. Emerging regions of NF loss (Fig. 1a,
105 green color) appeared in the northern and southwestern sectors of Mato Grosso during
106 this period, alongside sporadic patches within Roraima. More recent instances of NF loss
107 (2011-2022) happened across extensive areas in Roraima (denoted by the red color) and
108 localized patches in Rondônia. In summary, Mato Grosso exhibited a higher degree of
109 established and longstanding NF loss, while Rondônia, Roraima, and Amapá showed an
110 escalating contribution and growing relative values for NF losses.



111
 112 Fig. 1. Spatiotemporal distribution of vegetation loss (NF) in the Brazilian Amazon
 113 biome. a) RGB color composite of vegetation loss through time. Primary colors in the
 114 RGB system represent older (blue), intermediate (green), and younger loss (red). The
 115 secondary colors (cyan, yellow, and magenta) represent continuous loss among the
 116 analyzed period. b) Primary y-axis showing annual increments (km²) of vegetation loss.
 117 The flatlining effect on NF data is because the data are not available annually. Secondary
 118 y-axis showing the percentage of vegetation loss (NF and forest formations) relative to
 119 the area of the Brazilian Amazon biome (NF formations) and Brazilian Legal Amazon
 120 (forest formations). Historical vegetation loss data simplified from PRODES monitoring
 121 program¹⁷.
 122

123 The state of Mato Grosso had the highest absolute lost area (14,469.20 km²), the
124 second in relative terms (32.1%) (Fig. S2). The State ranks second in area deforested in
125 the Brazilian Amazon¹⁷ and holds the highest cultivated area (75.91% of total) and largest
126 cattle herd in the country (35.68%)¹⁸. Concentration of land in large properties contributes
127 to mechanized deforestation, along with the expansion of mechanized agriculture^{19,20}. The
128 transition region with the Pantanal and Cerrado savannas is highly affected by agricultural
129 expansion in the Amazon²¹⁻²³, appearing among the top three hotspots of NF loss (Fig. 1a
130 and Fig. S1). Fig. 1a shows that NF loss in transition areas between biomes is older (up
131 to 2000), with large hotspots appearing in the early 2000s in the central region of Mato
132 Grosso, following the advance of the east-to-west agricultural frontier and the increasing
133 conversion of pastures to soybean fields^{20,24}.

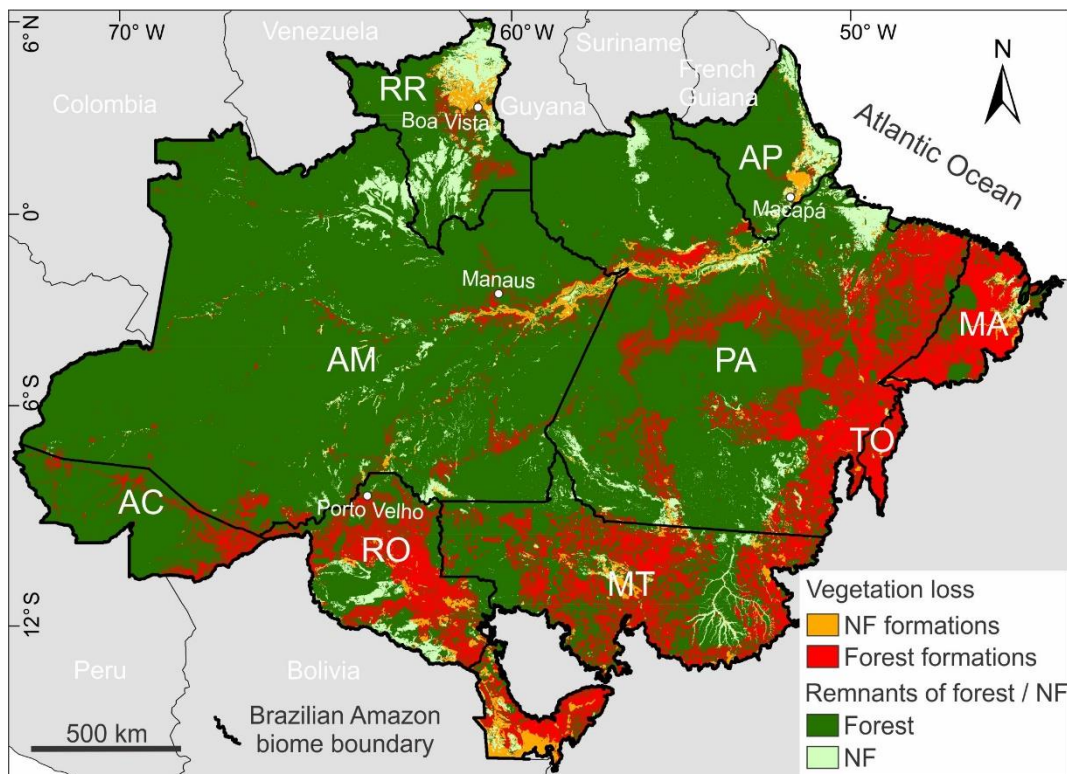
134 The states of Roraima and Amapá have been considered the last agricultural
135 frontiers in the Amazon^{25,26}. In both States, the influx of capital and technology found in
136 old frontier areas are among the main the causes of NF loss¹⁹. The expansion of
137 agricultural production was facilitated by various factors, such as highways, technological
138 innovations in seeds, low land prices, and proximity to the capital^{7,27,28}. State
139 governments have been playing an important role in attracting farmers from other states
140 to the *lavrado* savannas through economic subsidies and flexibility in state environmental
141 legislation²⁷⁻²⁹.

142 Roraima ranks third in total NF loss, with 3,527.70 km² (Fig. 1b). It hosts the
143 Amazon biome's largest continuous savanna area (~43,000 km²)²⁸. NF loss notably
144 intensified from 2001 to 2022, resulting in significant cleared areas. Soybean cultivation
145 saw exponential growth (191% in four years), driven by locally adapted seeds³⁰.
146 Roraima's savanna floodplains also favor rice cultivation³¹. This, along with silviculture
147 expansion, regional road projects, and port infrastructure improvements, heightens the
148 risk of savanna loss^{5,7,27,29}. In 2014, the region around the state capital, Boa Vista,
149 witnessed a significant fourfold increase in agriculture and other land uses, including
150 forestry and urban development. Urban growth reached 22%²⁷.

151 While Amapá initially experienced limited forest and NF loss due to its lack of
152 road connections with other states⁵, upcoming projects like the asphalt paving of BR-210
153 between Boa Vista and Macapá and the establishment of a port at the Amazon River's
154 mouth may heighten pressure on its savannas^{5,9}. Compounding this risk is the fact that
155 only 9.2% of the State's savannas are protected, with just 0.3% falling under strict
156 conservation units³². In contrast, 72% of the State's primarily forested lands are protected.

157 The protection gap between savannas and forests implies that land clearing leakage^{33,34}
158 into the savannas might be the reason behind the significantly larger area lost (7.3%)
159 compared to deforestation (2.8%)¹⁷, making Amapá the only state with this discrepancy.

160 Despite not having extensive, continuous areas of NF loss, Rondônia ranks as the
161 third state with the highest relative losses (11.4%). This has resulted in a loss of 2,656.70
162 km², making it the fourth-largest state in terms of area (Fig. S2a). Over the past decade,
163 there has been a consistent increase in the proportion of NF loss (Fig. S2b). While small
164 hotspots were observed as early as 2000, new focal points have emerged in the last decade
165 (Fig. 1a and Fig. S1). These recent changes may be attributed primarily to the conversion
166 of land to pasture, with a growing portion being allocated for soybean production^{11,35,36}.
167 The central region has older conversion areas, aligning with significant deforestation rates
168 experienced in the 1990s, especially near the BR-364 highway^{35,37}. It's important to note
169 that most NF areas in Rondônia are located within protected areas and are sparingly
170 distributed across the landscape³⁸. This fragmented distribution could explain the
171 presence of isolated hotspots of NF loss (Fig. S1) and the relatively modest ~10% loss in
172 a state that has already experienced a substantial 46.4% reduction in its forest cover³⁹.



174 Fig. 2. Geographical distribution of vegetation loss (NF and forest formations) in the
175 Brazilian Amazon biome, with the location of remaining areas of natural NF / forest. NF
176 data are from 2000 to 2022. Vegetation loss data in forest formations are simplified from

177 the PRODES monitoring program based on the cumulative mask until 2007 and from
178 2008 to 2022 (<http://terrabrasilis.dpi.inpe.br/downloads/>).

179

180 **Temporal trend of NF loss**

181 Examining yearly NF loss spanning from 2001 to 2022 (Fig. 1b) unveils three
182 distinctive: (1) between 2001 and 2008, losses exhibited a noteworthy decline, ranging
183 from 1,555.60 to 581.50 km²; (2) between 2009 and 2013, NF loss was relatively stable,
184 not surpassing 400 km² per year; (3) from 2014 to 2022, yearly NF loss increased,
185 oscillating between 515.00 and 726.50 km². Notably, the highest losses were recorded in
186 2001-2002 (1,555.60 km²), whereas the lowest values occurred during 2011-2014 (351.90
187 km²; Fig. 1b). Proportionally to their extents, deforestation (here used only to discriminate
188 forest loss mapped by PRODES Amazonia) almost always showed higher values than
189 that of NF loss during 2003-2013, with an inverted behavior during 2014-2022. In 2014,
190 NF relative loss was almost twice as high as the deforestation (~0.23% and 0.13%,
191 respectively; Fig. 1b). Nonetheless, the relative differences between deforestation and NF
192 loss yielded no significant statistical differences (Student t-test: $t = 0.768$; $df = 40.511$;
193 $p = 0.4464$), while both processes exhibited a robust positive correlation (Spearman's
194 rank correlation: $r = 0.87$; $p < 0.0001$) (Fig. 1b), suggesting that the extent of vegetation
195 loss in forested and non-forested areas did not differ in relative terms and show a
196 continuum at the biome scale.

197 The strong correlation between deforestation and NF loss suggests a shared
198 response to common factors. The observed decline in deforestation in the Brazilian
199 Amazon, particularly after 2004 (Fig. 1b), can be largely attributed to environmental
200 policies implemented by the Brazilian government in response to high deforestation rates
201 in the Brazilian Amazon. These policies, including: the Plan for Prevention and Control
202 of Deforestation in the Legal Amazon (PPCDAM); the Amazon Protected Areas Program
203 (ARPA); the prioritization of Amazonian municipalities for preventing, monitoring, and
204 controlling illegal deforestation; the Cattle Agreement; and the soy moratorium, have
205 played a significant role in reducing deforestation rates in the region^{24,40-45}. However, it
206 remains unclear to what extent NF loss during the same period responded to these
207 policies, as they were specifically designed and enforced for forested areas only. On the
208 other hand, it is conceivable that the decline in commodity prices during this period,
209 which has been shown to reduce new land clearings in the Amazon⁴⁶, could have played

210 a role in the observed decline in NF loss.

211 In line with the Amazon's deforestation trend, NF loss increased from 2013
212 onward (Fig. 1b). From 2013 to 2022, both NF loss and deforestation exhibited a rate of
213 $0.21 \text{ km}^2 \cdot \text{y}^{-1}$ (relative to their vegetation extents). Several factors contributed to this rise,
214 including the growing value of soybeans, which led to the conversion of pasture areas in
215 states like Mato Grosso with better logistical infrastructure for soybean production, and
216 the shift of pastures towards the active deforestation frontier^{24,47}. Additionally, the
217 increase in the global demand for meat resulted in the expansion of the cattle herd in the
218 Amazon¹⁸. The discussed environmental policies were identified as drivers of
219 deforestation leakage^{33,34} over NF formations by different studies^{5,9,22,28,48,49}. Similar
220 events were observed in the biome towards the neighbor Cerrado^{22,50,51}.

221 Political decisions have played a significant role in driving natural vegetation loss
222 escalation. Changes to the Forest Code in 2012, including amnesty for pre-2008
223 deforestation, and other bills aimed at easing environmental licensing, likely incentivized
224 deforestation in anticipation of further legislative changes due to political pressure from
225 the ruralist caucus in the National Congress⁵²⁻⁵⁴. The political influence on deforestation
226 intensified in the Amazon during the period of 2018-2021, likely influenced by the
227 incentives and discourse by then-president Jair Bolsonaro, coupled with the weakening
228 of command-and-control measures for deforestation in the Brazilian Amazon under his
229 government. Consequently, there was a rapid surge in deforestation and conversion,
230 human-induced fires, and illegal mining activities in various Brazilian biomes, well
231 documented elsewhere^{53,55-58}. In addition, NF formations, even in the Amazon Biome,
232 have less protection under the Brazilian Forest Code. A key limitation lies in the lack of
233 differentiation among the various Amazonian phytophysiognomies, which instead results
234 in their protection being generalized across the entire Brazilian Legal Amazon region.
235 For instance, while private properties are required by law to preserve 80% of forest
236 ecosystems in the biome, open ecosystems are less protected, ranging from 35% in
237 *cerrado* areas to 20% in “general” grasslands⁵⁹.

238

239 **Losses by phytophysiognomies**

240 Cross-referencing spatial data on suppression with the official map of Brazilian
241 vegetation coverage⁶⁰ allowed us to estimate the losses related to different NF
242 phytophysiognomies (Table 1). The ecotones (see Fig. S3 for phytophysiognomies

243 distribution) had the largest suppressed area (12,388.40 km²). They correspond to
 244 mixtures of different vegetation types (e.g., contact of savanna-ombrophilous forest,
 245 ombrophilous forest-deciduous forest; see Table S2), where their separation is limited
 246 through image interpretation^{4,6}. The loss of forest formation within NF areas was
 247 significant. During the analyzed period (2000-2022), 5,732.92 km² were cleared, and
 248 when considering ecotone areas (Table S2) with other NF formations, this number rises
 249 to 17,619.43 km², an average of 590.44 km² (± 619.94 km²) lost annually. By normalizing
 250 forest loss within the NF mask by its respective area and applying the same approach to
 251 deforestation within the PRODES Amazonia mask¹⁷ from 2001 to 2022, both processes
 252 displayed a loss rate of 0.04 km².year⁻¹ per km² of forest. This value signifies an ongoing
 253 pattern of vegetation loss extending across forested and NF areas, indicating a consistent
 254 continuum of loss.

255

256 Table 1. Phytophysiognomies suppression in the Brazilian Amazon inside the non-forest mask.
 257 See Table S2 for details regarding the ecotones. Differences in the mapping scales of vegetation
 258 and suppression data account for variations in the total suppression values found in this analysis.

NF phytophysiognomies	Total area (km ²)	NF Mask (%)	Lost area (km ²)	Lost area (%)	Lost area inside NF mask (%)
Ecotones (see table S2)	58,022.82	20.77	12,388.40	21.35	42.28
Savanna	77,428.90	27.72	10,246.55	13.23	34.97
Dense Ombrophilous Forest	44,358.87	15.88	2,987.14	6.73	10.19
Evergreen Seasonal Forest	4,838.88	1.73	1,238.30	25.59	4.23
Open Ombrophilous Forest	10,561.81	3.78	1,193.76	11.30	4.07
Pioneer formations	31,474.98	11.27	546.51	1.74	1.87
Semi-deciduous Seasonal Forest	2,351.38	0.84	309.63	13.17	1.06
Campinarana (white-sand vegetation)	37,040.58	13.26	171.73	0.46	0.59
Savanna-steppe	7,059.11	2.53	6.96	0.10	0.02
Deciduous Seasonal Forest	1,824.67	0.65	4.09	0.22	0.01

259

260 Savannas

261 Savannas, which represent approximately 30% - the largest proportion - of the NF
 262 mask, experienced the largest losses (excluding ecotones; Table 1). In the Amazon, these
 263 formations are more susceptible to clearing compared to dense forested areas²⁷.
 264 Considering the combined area of savannas and savanna-steppe, the loss amounted to
 265 10,253.51 km², accounting for approximately 35% of the NF losses and 13.3% of their
 266 total extent within the NF mask. If the contact areas of savannas with other formations

267 (ecotones; Table S2) are also included, losses would reach 21,550.58 km². The three main
268 hotspots of NF loss are located precisely within this ecosystem, specifically in the states
269 of Mato Grosso, Roraima, and Amapá (Fig. 1a and Fig. S1).

270 The main factor driving losses in these states has been linked to the expansion of
271 soybean cultivation, as discussed above. Even when not suppressed, the savannas may be
272 affected by fire and grazing^{27,61}. The threats primarily arise from neglect and biases
273 towards these ecosystems, often seen as successional stages of forests with low
274 biodiversity and ecological significance⁶²⁻⁶⁴. However, the Amazonian savannas are old
275 formations⁶⁵ and constitute vegetation islands (Fig. S3) with distinct characteristics from
276 the neighbor Brazilian Cerrado^{7,66}. In the states of Roraima and Amapá, the savannas
277 show a great heterogeneity and embraces different phytophysionomies (e.g., the
278 *lavrados*) with a biodiversity that is still poorly understood and home to endemic species
279 that are threatened in extinction due to limited protection within conservation units^{5,7,28,67}.
280 Some studies have highlighted their conservation importance due to their richness, rarity,
281 and endemic species, as well as species adapted to these ecosystems^{5,9}, low protection
282 (~12% within strictly protected areas⁹), and limited research on their ecology and
283 biodiversity^{5,7,32}.

284

285 Pioneer formations

286 Pioneer formations within the NF mask have experienced a 1.75% reduction in
287 their extent. These formations consist of pioneering vegetation elements such as grasses,
288 bryophytes, therophytes, cryptophytes, among others, which undergo a continuous
289 succession due to the seasonal instability of the inundated terrain, occurring
290 predominantly in lacustrine and alluvial soils⁴. They constitute a significant portion of the
291 vegetation in the floodplains along the major Amazonian rivers (Fig. S3). These
292 floodplains are periodically inundated and have historically been inhabited by local
293 riverine populations who have relied on the fertile soils for agriculture, extractive
294 activities, and fishing^{68,69}. Deforestation in these areas is primarily driven by agriculture
295 and livestock⁷⁰. The floodplains of the Solimões/Amazonas River have experienced
296 moderate NF loss, with higher intensity during the first period (Fig. 1a and Fig. S1),
297 particularly concentrated around the city of Manaus and in the Lower Amazon region
298 between the states of Amazonas and Pará. The central section of the Amazon River
299 channel (ranging from 56°W to 55°W) has been recognized as one of the regions most

300 affected by vegetation loss, encompassing both recent and historical losses, even though
301 certain areas might still maintain up to 70% of their forest cover^{70,71}.

302

303 Campinaranas (white-sand vegetation)

304 The campinaranas, also known as white-sand ecosystems, and the savanna-
305 steppes were the least affected by NF loss (0.46% and 0.10%, respectively; Table 1).
306 While the latter has the smallest total area within the NF mask among the open
307 phytophysionomies, the campinaranas rank second in area, behind only the savannas.
308 They encompass a gradient of grassland to forest formations, showing a broad geographic
309 distribution (Fig. S3), and occur as predominant formations in large patches of hundreds
310 of square kilometers in the Negro River basin (ottobasin, level 2) and in border regions
311 with Colombia, as well as exhibiting an island-like distribution pattern throughout the
312 Amazon embedded in other habitats^{4,6,10}. Their greater geographical distribution in the
313 northwest of the basin, a region with the lowest anthropogenic pressure in the Brazilian
314 Amazon^{72,73}, may explain the low levels of losses in this formation. Furthermore, the
315 nutrient-poor, acidic, and water-limited sandy soils^{6,10,74} may limit their suitability for
316 agricultural use.

317 The white-sand ecosystems have lower species richness and diversity of fauna and
318 flora compared to adjacent upland forests, but these are adapted to specific soil conditions
319 and include endemic species⁷⁵⁻⁷⁹, making them of utmost biological importance and
320 strategically significant for conservation^{6,10}. However, like other NF ecosystems in the
321 Amazon, knowledge of biodiversity and even the functioning of these ecosystems are still
322 limited, especially in those island-like areas with patches smaller than 1 km² and
323 immersed within forested areas, which are difficult to map using satellite imagery⁶. Due
324 to the nature of their soils, they are fragile systems and highly sensitive to anthropogenic
325 disturbances⁸⁰, and their low protection (< 1 km² within conservation units)⁶ makes them
326 extremely vulnerable. Although of low agricultural interest, sand extraction poses a
327 significant threat of white-sand ecosystems in the state of Pará⁸¹.

328

329 **Final remarks**

330 The strong correlation we identified between NF loss and deforestation (Fig. 1b)
331 suggests a shared response to prevalent factors that contribute to the Amazonian
332 vegetation loss. Conversion to agriculture and pasture has long been recognized as the

333 primary driver of forest loss in the Brazilian Amazon⁸² and, as showed here, it is also
334 believed to be the main cause of NF loss, although further investigation is needed. The
335 development of road and port infrastructure in states like Roraima and Amapá, as
336 discussed above, has facilitated the expansion of agricultural lands, reminiscent of past
337 frontiers in the Amazon^{83,84}. These states have become emerging agricultural frontiers,
338 driven by the cultivation of crops adapted to the region's soil and climate conditions.
339 Additionally, public policies such as the beef agreement and soy moratorium, which only
340 restrict forest clearing, may have inadvertently led to the encroachment of cattle and soy
341 production into NF ecosystems that were not previously monitored by the PRODES
342 system^{5,9,22,28,48,49}.

343 The limited protection of Amazon NF phytophysionomies is one of the common
344 factors observed in different states where large areas of NF have been lost, accentuated
345 by the low protection in the Brazilian Forest Code, as discussed before. Regarding
346 Amazonian savannas, only 12.3% are within conservation units, reaching 58% if
347 Indigenous Lands are considered⁹. However, a more detailed analysis of the
348 representativeness of different types of NF within protected areas is still lacking, as the
349 protection of each NF phytophysionomies is uneven^{29,32,62,63}. Additionally, political
350 pressures and approval delays/denial of ecological-economic zoning plans have posed a
351 frequent threat to these ecosystems, particularly in states where agricultural expansion is
352 the main proposal for economic development^{5,7,11,28}.

353 The join analysis of PRODES Amazonia and PRODES NF has provided the first-
354 ever official quantification of total vegetation loss in the Brazilian Amazon at the biome
355 level. Taking into account the total NF loss (29,247.44 km²) and the total deforestation
356 (768,930.88 km²)¹⁷, overall, the Brazilian Amazon biome has experienced a total loss of
357 798,178.32 km², representing 18.93% of its original vegetation cover (Fig. 2). The lower
358 proportional loss of NF compared to forests may reflect their island-like spatial
359 distribution, mostly located outside the *deforestation arc* (Fig. S3) and lower agriculture
360 suitability. However, the increasing trend of NF loss observed in the last decade, with
361 rates following deforestation, and the growing threat from the expansion of agribusiness
362 activities over NF formations, combined with inadequate protection and undervaluation,
363 indicates a scenario of significant losses, especially for small island-like NF formations.
364 A more detailed analysis currently underway for each vegetation type may provide further
365 insights into the threats.

366 We emphasize the recommendations proposed by researchers who have examined
367 the risk faced by NF formations^{5,7,9,32,63}. Establishing protected areas, whether for
368 sustainable use or integral protection, both of which have strong evidence of effectively
369 safeguarding local biodiversity and reducing forest loss⁸⁵⁻⁸⁷, represents a critical measure
370 for strategically planning the conservation of these ecosystems. However, it will require
371 an effort to map the heterogeneity within these ecosystems to effectively protect distinct
372 phytogeographic units of these vegetation types. In this regard, the limited knowledge
373 about the biodiversity and social importance of these vegetation types needs to be
374 overcome to obtain minimal informed decisions regarding the protection and sustainable
375 use of these ecosystems through the suitable, responsible, and participatory ecological-
376 economic zoning.

377 As assigned by Overbeck et al.⁶³, savannah conservation often requires different
378 strategies than that of forests (e.g., prescribed fire, grazing). Protecting and sustainably
379 managing these ecosystems will require research to harness their contributions to
380 biodiversity, ecosystem services, and existing uses as already occurring, for example, in
381 the Pampa grasslands⁸⁸. The implementation of monitoring programs, such as the ongoing
382 DETER (Near Real-Time Deforestation Detection System)³ in NF areas and the
383 continuation of the annual deforestation mapping by PRODES NF are essential actions
384 to enable greater control through enforcement agencies and responsible trade of
385 agricultural products from legally cleared areas.

386 In this context, extending the soybean moratorium and cattle agreement, or
387 implementing any sectoral or national mechanism with a robust monitoring and
388 verification system to promote deforestation-free agricommodity supply chains,
389 represents significant measures to curb the progression of NF loss⁴⁵. The introduction of
390 PRODES NF provides essential data (i) to allows the monitoring of socioenvironmental
391 compliance of production sites and the use of traceability systems to achieve more
392 sustainable production, and (ii) to plan effective actions to control and combat the
393 conversion of natural vegetation in non-forest areas of the Amazon.

394

395 **Acknowledgements**

396 We would like to thanks to the National Council of Technological and Scientific
397 Development (CNPq) under project number 444418/2018-0 namely "Monitoring
398 Brazilian Biomes by Satellite - Building New Capacities" and supported by INPE. We

399 also thank the institutional support from INPE. Special thanks to the Brazilian
400 Association of Vegetable Oil Industries (ABIOVE) for providing financial support for
401 fieldwork. We extend our appreciation to Embrapa Roraima for their partnership,
402 fieldwork insights, planning and several car trips. We thank David Galbraith (University
403 of Leeds) for his comments and insights on the initial version of the manuscript. Lastly,
404 we acknowledge the experts Adriano Venturieri (Embrapa Eastern Amazon), Andréa
405 dos Santos Coelho (SEMAS Pará), Evelyn Moraes Novo (INPE) and Tassio Koiti
406 Igawa (Embrapa Eastern Amazon) for their valuable contributions.

407 **Author contributions** CGM, CAA, DES, LS, and LEM were involved in the
408 coordination of the project and in the article development; VLC, FCA, LJS, and MR
409 contributed to article development, data analysis, and figures; TCL, VR, DLCL, APB,
410 DCMB, and DRVM served as auditors; CBQ to MBS comprised the team of
411 interpretation experts; FCP and AC were responsible for project database maintenance;
412 HX, MX acted as external consultants and participated in fieldwork; APM, LB, EBS,
413 and LGF formed the external validation team for the results; MA analyzed the
414 validation results, conducted a thorough review, and made substantial contributions to
415 the writing of the article.

416 **Data availability** The annual NF loss data in the Brazilian Amazon biome for the
417 period from 2000 to 2022 can be downloaded at
418 <http://terrabrasilis.dpi.inpe.br/downloads/>.

419 **Author Information** Reprints and permissions information is available at
420 www.nature.com/reprints and permissions. The authors declare no competing financial
421 interests. Correspondence and requests for materials should be addressed to CGM
422 (cassiano.messias@inpe.br) or CAA (claudio.almeida@inpe.br).

423

424 **Additional Information** Supplementary Information is available for this paper.

425

426 **References**

- 427 1 Almeida, C. A. *et al.* Methodology for Forest Monitoring used in PRODES and DETER
428 projects - 2nd edition (updated). (INPE, São José dos Campos, 2022).
- 429 2 INPE. *PRODES Amazonia citation track*,
430 <[http://www.obt.inpe.br/OBT/assuntos/programas/amazonia/prodes/citacoes-ao-](http://www.obt.inpe.br/OBT/assuntos/programas/amazonia/prodes/citacoes-ao-prodes)
431 [prodes](http://www.obt.inpe.br/OBT/assuntos/programas/amazonia/prodes/citacoes-ao-prodes)> (2023).

- 432 3 Diniz, C. G. *et al.* DETER-B: The New Amazon Near Real-Time Deforestation
433 Detection System. *IEEE Journal of Selected Topics in Applied Earth Observations and*
434 *Remote Sensing* **8**, 3619-3628, <https://doi.org/10.1109/JSTARS.2015.2437075> (2015).
- 435 4 IBGE. *Manual técnico da vegetação brasileira*. 2 edn, (IBGE, 2012).
- 436 5 Mustin, K. *et al.* Biodiversity, threats and conservation challenges in the Cerrado of
437 Amapá, an Amazonian savanna. *Nature Conservation* **22**, 107-127,
438 <https://doi.org/10.3897/natureconservation.22.13823> (2017).
- 439 6 Adeney, J. M., Christensen, N. L., Vicentini, A. & Cohn-Haft, M. White-sand
440 ecosystems in Amazonia. *Biotropica* **48**, 7-23, <https://doi.org/10.1111/btp.12293>
441 (2016).
- 442 7 Barbosa, R. I., Campos, C., Pinto, F. & Fearnside, P. M. The "Lavrados" of Roraima:
443 biodiversity and conservation of Brazil's Amazonian savannas. *Functional Ecosystems*
444 *and Communities* **1**, 29-41 (2007).
- 445 8 Araujo, M. A. M., da Rocha, A. E. S., Miranda, I. S. & Barbosa, R. I. Hydro-edaphic
446 conditions defining richness and species composition in savanna areas of the northern
447 Brazilian Amazonia. *Biodiversity Data Journal*, e13829,
448 <https://doi.org/10.3897/BDJ.5.e13829> (2017).
- 449 9 de Carvalho, W. D. & Mustin, K. The highly threatened and little known Amazonian
450 savannahs. *Nature Ecology & Evolution* **1**, 100, [https://doi.org/10.1038/s41559-017-](https://doi.org/10.1038/s41559-017-0100-0100)
451 [0100](https://doi.org/10.1038/s41559-017-0100-0100) (2017).
- 452 10 Fine, P. V. A. & Bruna, E. M. Neotropical white-sand forests: origins, ecology and
453 conservation of a unique rain forest environment. *Biotropica* **48**, 5-6,
454 <https://doi.org/10.1111/btp.12305> (2016).
- 455 11 Santos, L. B. *et al.* Proposta metodológica para mapeamento das áreas de não-floresta
456 presentes no projeto de monitoramento de áreas desflorestadas da Amazônia Legal
457 Brasileira. *Research, Society and Development* **11**, e20411425794,
458 <https://doi.org/10.33448/rsd-v11i4.25794> (2022).
- 459 12 Sano, E. E. *et al.* in *XVIII Simpósio Brasileiro de Sensoriamento Remoto* 59488
460 (INPE, Santos, 2017).
- 461 13 Almeida, C. A. *et al.* Mapping natural non-forest vegetation removal in the Brazilian
462 Amazon – a pilot project. *Int. Arch. Photogramm. Remote Sens. Spatial Inf. Sci.* **XLIII-**
463 **B3-2022**, 1341–1348, <https://doi.org/10.5194/isprs-archives-XLIII-B3-2022-1341-2022>
464 (2022).
- 465 14 van der Hoff, R. & Rajão, R. The politics of environmental market instruments:
466 coalition building and knowledge filtering in the regulation of forest certificates trading
467 in Brazil. *Land Use Policy* **96**, 104666,
468 <https://doi.org/10.1016/j.landusepol.2020.104666> (2020).
- 469 15 Bernasconi, P., Blumentrath, S., Barton, D. N., Rusch, G. M. & Romeiro, A. R.
470 Constraining Forest Certificate's Market to Improve Cost-Effectiveness of Biodiversity
471 Conservation in Sao Paulo State, Brazil. *PLoS One* **11**, e0164850,
472 <https://doi.org/10.1371/journal.pone.0164850> (2016).
- 473 16 Soares-Filho, B. *et al.* Brazil's market for trading forest certificates. *PLoS One* **11**,
474 e0152311, <https://doi.org/10.1371/journal.pone.0152311> (2016).
- 475 17 INPE. Terrabrasilis - Plataforma de dados geográficos. (2023).
- 476 18 IBGE. *Sistema IBGE de Recuperação Automática - SIDRA*,
477 <<https://sidra.ibge.gov.br/home/pimpfgr/nordeste>> (2023).
- 478 19 Alves, D. S. in *Amazonia and Global Change* (eds Michael Keller, Mercedes
479 Bustamante, John Gash, & Pedro Silva Dias) 11-23 (2009).
- 480 20 Morton, D. C. *et al.* Cropland expansion changes deforestation dynamics in the
481 southern Brazilian Amazon. *Proc. Natl. Acad. Sci. USA* **103**, 14637-14641,
482 <https://doi.org/10.1073/pnas.0606377103> (2006).
- 483 21 Bonini, I. *et al.* Collapse of ecosystem carbon stocks due to forest conversion to
484 soybean plantations at the Amazon-Cerrado transition. *For. Ecol. Manage.* **414**, 64-73,
485 <https://doi.org/10.1016/j.foreco.2018.01.038> (2018).

- 486 22 Magalhães, I. B. *et al.* Brazilian Cerrado and soy moratorium: effects on biome
487 preservation and consequences on grain production. *Land Use Policy* **99**, 105030,
488 <https://doi.org/10.1016/j.landusepol.2020.105030> (2020).
- 489 23 Bezerra, F. G. S. *et al.* New land-use change scenarios for Brazil: Refining global SSPs
490 with a regional spatially-explicit allocation model. *PLoS One* **17**, e0256052,
491 <https://doi.org/10.1371/journal.pone.0256052> (2022).
- 492 24 Macedo, M. N. *et al.* Decoupling of deforestation and soy production in the southern
493 Amazon during the late 2000s. *PNAS* **109**, 1341-1346,
494 <https://doi.org/10.1073/pnas.1111374109> (2012).
- 495 25 Staevie, P. M. Um balanço das discussões sobre os impactos do agronegócio sobre a
496 Amazônia brasileira. *Revista NERA* **21**, 98-112 (2018).
- 497 26 Silva, E. in *Revista Globo Rural* Vol. 371 28-33 (Editora Globo, Rio de Janeiro, RJ,
498 2016).
- 499 27 da Silva, G. F. N. & Oliveira, I. J. Reconfiguration of the landscape in the Amazonian
500 savannas. *Mercator* **17**, e17028, <https://doi.org/10.4215/rm2018.e317028> (2018).
- 501 28 Barbosa, R. I. & Campos, C. Detection and geographical distribution of clearing areas
502 in the savannas ('lavrado') of Roraima using Google Earth web tool. *Journal of*
503 *Geography and Regional Planning* **4**, 122-136 (2011).
- 504 29 Yokomizo, G. K. I. & Costa, L. D. N. O uso do cerrado amapaense e os recursos
505 vegetais. *DRd - Desenvolvimento Regional em Debate* **6**, 164-177,
506 <https://doi.org/10.24302/drd.v6i3.1122> (2016).
- 507 30 Rodrigues, C. in *gl* (2023).
- 508 31 Cordeiro, A. C. C., Suhre, E., Medeiros, R. D. & Vilarinho, A. A. Sistemas de cultivo e
509 manejo de água na produção de diferentes genótipos de arroz em várzea, no estado de
510 Roraima. *Pesquisa Agropecuária Tropical* **40**, 362-369 (2010).
- 511 32 Hilário, R. R. *et al.* The fate of an Amazonian savanna: government land-use planning
512 endangers sustainable development in Amapá, the most protected Brazilian state.
513 *Tropical Conservation Science* **10**, 1-8, <https://doi.org/10.1177/1940082917735416>
514 (2017).
- 515 33 Pfaff, A. & Robalino, J. Spillovers from conservation programs. *Annual Review of*
516 *Resource Economics* **9**, 299-315, [https://doi.org/10.1146/annurev-resource-100516-](https://doi.org/10.1146/annurev-resource-100516-053543)
517 [053543](https://doi.org/10.1146/annurev-resource-100516-053543) (2017).
- 518 34 Meyfroidt, P. *et al.* Focus on leakage and spillovers: informing land-use governance in a
519 tele-coupled world. *Environmental Research Letters* **15**, 090202,
520 <https://doi.org/10.1088/1748-9326/ab7397> (2020).
- 521 35 Alves, D. S., Escada, M. I. S., Pereira, J. L. G. & de Albuquerque Linhares, C. Land use
522 intensification and abandonment in Rondônia, Brazilian Amazônia. *Int. J. Remote Sens.*
523 **24**, 899-903, <https://doi.org/10.1080/0143116021000015807> (2010).
- 524 36 Costa, O. B., Matricardi, E. A. T., Pedlowski, M. A., Cochrane, M. A. & Fernandes, L.
525 C. Spatiotemporal mapping of soybean plantations in Rondônia, Western Brazilian
526 Amazon. *Acta Amazon.* **47**, 29-38, <https://doi.org/10.1590/1809-4392201601544>
527 (2017).
- 528 37 Ferraz, S. F. B., Vettorazzi, C. A., Theobald, D. M. & Ballester, M. V. R. Landscape
529 dynamics of Amazonian deforestation between 1984 and 2002 in central Rondônia,
530 Brazil: assessment and future scenarios. *For. Ecol. Manage.* **204**, 69-85,
531 <https://doi.org/10.1016/j.foreco.2004.07.073> (2005).
- 532 38 Rosa, M. C. *O que está por trás das áreas de “não-floresta” do projeto PRODES-INPE*
533 *no estado de Rondônia?* Bachelor thesis, Universidade de Brasília, (2017).
- 534 39 INPE. *PRODES Amazônia - Monitoramento do Desmatamento da Floresta Amazônica*
535 *Brasileira por Satélite*,
536 <<http://www.obt.inpe.br/OBT/assuntos/programas/amazonia/prodes>> (2023).
- 537 40 Assunção, J. & Rocha, R. Getting greener by going black: the effect of blacklisting
538 municipalities on Amazon deforestation. *Environment and Development Economics* **24**,
539 115-137, <https://doi.org/10.1017/S1355770X18000499> (2019).

540 41 Nepstad, D. *et al.* Slowing Amazon deforestation through public policy and
541 interventions in beef and soy supply chains. *Science* **344**, 1118-1123,
542 <https://doi.org/10.1126/science.1248525> (2014).

543 42 West, T. A. P. & Fearnside, P. M. Brazil's conservation reform and the reduction of
544 deforestation in Amazonia. *Land Use Policy* **100**, 105072,
545 <https://doi.org/10.1016/j.landusepol.2020.105072> (2021).

546 43 Silva, A. L. & Bueno, M. A. F. The Amazon Protected Areas Program (ARPA):
547 participation, local development, and governance in the Brazilian Amazon.
548 *Biodiversidade Brasileira* **7**, 122-137, <https://doi.org/10.37002/biobrasil.v%vi%i.641>
549 (2017).

550 44 Soares-Filho, B. S. *et al.* Contribution of the Amazon protected areas program to forest
551 conservation. *Biol. Conserv.* **279**, 109928, <https://doi.org/10.1016/j.biocon.2023.109928>
552 (2023).

553 45 Heilmayr, R., Rausch, L. L., Munger, J. & Gibbs, H. K. Brazil's Amazon soy
554 moratorium reduced deforestation. *Nature Food* **1**, 801-810,
555 <https://doi.org/10.1038/s43016-020-00194-5> (2020).

556 46 Assunção, J., Gandour, C. C. & Rocha, R. in *Climate Policy Initiative Working Paper*
557 (PUC Rio, Rio de Janeiro, RJ, 2012).

558 47 Song, X. P. *et al.* Massive soybean expansion in South America since 2000 and
559 implications for conservation. *Nature Sustainability* **4**, 784-792,
560 <https://doi.org/10.1038/s41893-021-00729-z> (2021).

561 48 Richards, P., Arima, E., VanWey, L., Cohn, A. & Bhattarai, N. Are Brazil's deforesters
562 avoiding detection? *Conservation Letters* **10**, 470-476,
563 <https://doi.org/10.1111%2Fconl.12310> (2017).

564 49 Soterroni, A. C. *et al.* Expanding the Soy Moratorium to Brazil's Cerrado. *Science*
565 *Advances* **5**, eaav7336, <https://doi.org/10.1126/sciadv.aav7336> (2019).

566 50 Moffette, F. & Gibbs, H. K. Agricultural Displacement and Deforestation Leakage in
567 the Brazilian Legal Amazon. *Land Economics* **97**, 155-179,
568 <https://doi.org/10.3368/wple.97.1.040219-0045R> (2021).

569 51 Kuschnig, N., Cuaresma, J. C., Krisztin, T. & Giljum, S. Spatial spillover effects from
570 agriculture drive deforestation in Mato Grosso, Brazil. *Science Reports* **11**, 21804,
571 <https://doi.org/10.1038/s41598-021-00861-y> (2021).

572 52 Ferrante, L. & Fearnside, P. M. Brazil's new president and 'ruralists' threaten
573 Amazonia's environment, traditional peoples and the global climate. *Environ. Conserv.*
574 **46**, 261-263, <https://doi.org/10.1017/S0376892919000213> (2019).

575 53 Pereira, E. J. A. L., Ribeiro, L. C. S., Freitas, L. F. S. & Pereira, H. B. B. Brazilian
576 policy and agribusiness damage the Amazon rainforest. *Land Use Policy* **92**, 104491,
577 <https://doi.org/10.1016/j.landusepol.2020.104491> (2020).

578 54 Rajão, R. *et al.* The rotten apples of Brazil's agribusiness. *Science* **369**, 246-248,
579 <https://doi.org/10.1126/science.aba6646> (2020).

580 55 Pelicice, F. M. & Castello, L. A political tsunami hits Amazon conservation. *Aquat.*
581 *Conserv.: Mar. Freshwat. Ecosyst.* **31**, 1221-1229, <https://doi.org/10.1002/aqc.3565>
582 (2021).

583 56 Ferrante, L. & Fearnside, P. M. Brazil's political upset threatens Amazonia. *Science*
584 **371**, 898, <https://doi.org/10.1126/science.abg9786> (2021).

585 57 Vale, M. M. *et al.* The COVID-19 pandemic as an opportunity to weaken
586 environmental protection in Brazil. *Biol. Conserv.* **255**, 108994,
587 <https://doi.org/10.1016/j.biocon.2021.108994> (2021).

588 58 Ramos, A. The Amazon under Bolsonaro. *Aisthesis* **70**, 287-310,
589 <https://doi.org/10.7764/aisth.70.13> (2021).

590 59 Brasil. *Lei nº 12.651, de 25 de maio de 2012* (Publicado no D.O.U. de 28 mai. 2012,
591 2012).

592 60 IBGE. *Vegetação 1:250.000*, <<https://www.ibge.gov.br/geociencias/informacoes-ambientais/vegetacao/22453-cartas-1-250-000.html?=&t=acesso-ao-produto>> (2022).

- 594 61 Lima, J. M. *et al.* Influência do regime de queimadas sobre a riqueza e composição
595 florística de uma savana isolada na Amazônia - PELD Oeste do Pará. *Oecologia*
596 *Australis* **24**, 301-316, <https://doi.org/10.4257/oeco.2020.2402.06> (2020).
- 597 62 Overbeck, G. E. *et al.* Placing Brazil's grasslands and savannas on the map of science
598 and conservation. *Perspect. Plant Ecol. Evol. Syst.* **56**, 125687,
599 <https://doi.org/10.1016/j.ppees.2022.125687> (2022).
- 600 63 Overbeck, G. E. *et al.* Conservation in Brazil needs to include non-forest ecosystems.
601 *Divers. Distrib.* **21**, 1455-1460, <https://doi.org/10.1111/ddi.12380> (2015).
- 602 64 Bond, W. J. & Parr, C. L. Beyond the forest edge: ecology, diversity and conservation
603 of the grassy biomes. *Biol. Conserv.* **143**, 2395-2404,
604 <https://doi.org/10.1016/j.biocon.2009.12.012> (2010).
- 605 65 Mayle, F. E., Langstroth, R. P., Fisher, R. A. & Meir, P. Long-term forest-savannah
606 dynamics in the Bolivian Amazon: implications for conservation. *Philos. Trans. R. Soc.*
607 *London, Ser. B: Biol. Sci.* **362**, 291-307, <https://doi.org/10.1098/rstb.2006.1987> (2007).
- 608 66 Prance, G. T. Islands in Amazonia. *Philosophical Transactions of the Royal Society of*
609 *London. Series B: Biological Sciences* **351**, 823-833,
610 <https://doi.org/10.1098/rstb.1996.0077> (1996).
- 611 67 Barbosa, R. I. & Miranda, I. S. in *Savanas de Roraima - Etnoecologia, Biodiversidade e*
612 *Potencialidades Agrossilvipastoris* (eds R. I. Barbosa, H. A. M. Xaud, & J. M. Costa e
613 Souza) (FEMACT, 2004).
- 614 68 Pinedo-Vasquez, M., Ruffino, M. L., Padoch, C. & Brondízio, E. S. (Springer,
615 Dordrecht, 2011).
- 616 69 Junk, W. J. (Springer, Berlin, 1997).
- 617 70 Renó, V. F., Novo, E. M. L. M., Suemitsu, C., Rennó, C. D. & Silva, T. S. F.
618 Assessment of deforestation in the Lower Amazon floodplain using historical Landsat
619 MSS/TM imagery. *Remote Sens. Environ.* **115**, 3446-3456,
620 <https://doi.org/10.1016/j.rse.2011.08.008> (2011).
- 621 71 Renó, V. & Novo, E. Forest depletion gradient along the Amazon floodplain. *Ecol.*
622 *Indicators* **98**, 409-419, <https://doi.org/10.1016/j.ecolind.2018.11.019> (2019).
- 623 72 Barreto, P. *et al.* *Human pressure on the Brazilian Amazon forests.* (Imazon, Global
624 Forest Watch, WRI, 2006).
- 625 73 Numata, I. & Cochrane, M. A. Forest Fragmentation and Its Potential Implications in
626 the Brazilian Amazon between 2001 and 2010. *Open Journal of Forestry* **02**, 265-271,
627 <https://doi.org/10.4236/ojf.2012.24033> (2012).
- 628 74 Luizão, F. J., Luizão, R. C. C. & Proctor, J. Soil acidity and nutrient deficiency in
629 central Amazonian heath forest soils. *Plant Ecol.* **192**, 209-224,
630 <https://doi.org/10.1007/s11258-007-9317-6> (2007).
- 631 75 Fine, P. V. A. & Baraloto, C. Habitat endemism in white-sand forests: insights into the
632 mechanisms of lineage diversification and community assembly of the Neotropical
633 flora. *Biotropica* **48**, 24-33, <https://doi.org/10.1111/btp.12301> (2016).
- 634 76 Matos, M. V. *et al.* Comparative phylogeography of two bird species, *Tachyphonus*
635 *phoenicius* (Thraupidae) and *Polytmus theresiae* (Trochilidae), specialized in
636 Amazonian white-sand vegetation. *Biotropica* **48**, 110-120,
637 <https://doi.org/10.1111/btp.12292> (2016).
- 638 77 Alonso, J. Á., Metz, M. R. & Fine, P. V. A. Habitat specialization by birds in Western
639 Amazonian white-sand forests. *Biotropica* **45**, 365-372,
640 <https://doi.org/10.1111/btp.12020> (2013).
- 641 78 Guevara, J. E. *et al.* Low phylogenetic beta diversity and geographic neo-endemism in
642 Amazonian white-sand forests. *Biotropica* **48**, 34-46, <https://doi.org/10.1111/btp.12298>
643 (2016).
- 644 79 Anderson, A. B. White-sand vegetation of Brazilian Amazonia. *Biotropica* **13**, 199-210,
645 <https://doi.org/10.2307/2388125> (1981).
- 646 80 Ramalho, W. P. *et al.* Impacto do assoreamento sobre a diversidade de peixes em
647 igarapés de um complexo vegetacional de campinarana no noroeste do Acre, Brasil.
648 *Neotrop. Biol. Conserv.* **9**, 105-114 (2014).

- 649 81 Ferreira, L. V., Chaves, P. P., Cunha, D. A., Rosário, A. S. & Parolin, P. A extração
650 ilegal de areia como causa de desaparecimento de campinas e campinaranas no estado
651 do Pará, Brasil. *Pesquisas, Botânica* **64**, 157-173 (2013).
- 652 82 Berenguer, E. *et al.* in *Amazon Assessment Report 2021* (eds C. Nobre *et al.*) Ch. 19,
653 19.11-19.41 (United Nations Sustainable Development Solutions Network, 2021).
- 654 83 Moran, E. F. *Developing the Amazon*. 292 (Indiana University Press, 1981).
- 655 84 Wood, C. H. & Porro, R. (University Press of Florida, Gainesville, 2002).
- 656 85 Nepstad, D. *et al.* Inhibition of Amazon deforestation and fire by parks and Indigenous
657 Lands. *Conserv. Biol.* **20**, 65-73, <https://doi.org/10.1111/j.1523-1739.2006.00351.x>
658 (2006).
- 659 86 Nolte, C., Agrawal, A., Silviu, K. M. & Soares-Filho, B. S. Governance regime and
660 location influence avoided deforestation success of protected areas in the Brazilian
661 Amazon. *PNAS* **110**, 4956-4961, <https://doi.org/10.1073/pnas.1214786110> (2013).
- 662 87 Pfaff, A., Robalino, J., Herrera, D. & Sandoval, C. Protected areas' impacts on Brazilian
663 Amazon deforestation: examining conservation-development interactions to inform
664 planning. *PLoS One* **10**, e0129460, <https://doi.org/10.1371/journal.pone.0129460>
665 (2015).
- 666 88 Baggio, R., Overbeck, G. E., Durigan, G. & Pillar, V. D. To graze or not to graze: a
667 core question for conservation and sustainable use of grassy ecosystems in Brazil.
668 *Perspectives in Ecology and Conservation* **19**, 256-266,
669 <https://doi.org/10.1016/j.pecon.2021.06.002> (2021).

670

671

672 **Methods**

673 We used a multi-sensor satellite imagery approach with digital image processing
674 and visual image interpretation techniques to map NF vegetation loss within the Brazilian
675 Amazon biome for the period from 2000 to 2022. Images from the Landsat series (MSS,
676 TM, ETM+, OLI sensors) were used to create the baseline map for 2000 and track changes
677 until 2014. From 2016 onwards, Sentinel-2A and 2B images (MSI sensor with a 20-10 m
678 spatial resolution) were utilized due to their improved temporal resolution, allowing for
679 images with less cloud coverage (see Table S2 for details). The number of images per
680 year ranged from 182 to 210 for Landsat and from 546 to 885 for Sentinel (Table S2).
681 Images from the sensor MSS (12 images) were used as auxiliaries to better identify
682 changes in the floodplains of the Amazon and Solimões rivers for the 2000 base map.

683 The images were processed using scripts in the Google Earth Engine (GEE) cloud
684 computing platform⁸⁹. Selection filters were applied to obtain cloud-free or minimally
685 cloud-covered images. PRODES methodology established an optimal mapping period for
686 each Landsat orbit/point during the dry season, taking into account the region's extensive
687 longitudinal and latitudinal range¹. Areas with persistent cloud cover were classified as
688 unobserved. Additional preprocessing steps included resampling the red spectral band of
689 Sentinel-2 from 10 m to 20 m resolution for compatibility with other bands, creating
690

691 mosaics of Sentinel-2 images based on the Landsat grids, and enhancing image contrast
692 through histogram manipulation.

693

694 **NF loss mapping**

695 NF loss was mapped using the incremental approach in which NF loss of a given
696 year is mapped using a cumulative or exclusion mask of NF loss from previous years,
697 similar to the PRODES approach¹. This procedure ensures that only newly cleared areas
698 are mapped each year, preventing duplicate mapping of the same area. The year of 2000
699 was defined as the basemap to create the mask of NF loss. The inclusion of older satellite
700 images (1980s and 1990s) from the Landsat MSS and TM sensors was necessary to create
701 this mask.

702 Mapping of NF loss occurred biennially from 2000 to 2018 and annually starting
703 in 2019, based on the availability of human resources and higher temporal resolution
704 satellite images. The sole exception was in 2012, when mapping for 2013 took place
705 instead. This deviation was due to the failure of the Landsat-5/TM sensor in November
706 2011, which disrupted the continuous monitoring of the Earth's surface. Monitoring only
707 resumed in February 2013 with the launch of Landsat-8. For the two-year interval
708 (biennium), we distributed half of the total increment evenly to each year as an
709 approximation for the NF loss amounts in the unmapped year.

710 NF vegetation was mapped by visual image interpretation techniques using
711 specific elements and interpretation keys to better distinguish between preserved and
712 suppressed vegetation patches. The elements color, shade, texture, shape, and context
713 were used for this purpose. Together with PRODES NF team of interpreters, experts with
714 knowledge in Amazonian NF vegetation also helped to define standard visual
715 interpretation keys (Table S1). NF mapping was performed in the TerraAmazon
716 software⁹⁰ at maximum (1:125,000) and minimum (1:75,000) scales based on a same false
717 color composite: shortwave infrared (R), near infrared or near infrared narrow (G) and
718 red (B). For the mapping of NF loss in the floodplains of some large Amazonian rivers, a
719 mask of water bodies was used to avoid mapping of anthropogenic land use on riverbanks
720 and seasonally dry lakes.

721

722 **Post-processing and audit of NF data**

723 The NF data underwent post-mapping operations to ensure data quality. An
724 independent audit team was established to check and correct any eventual errors (e.g.,

725 omission, commission and topology) of NF polygons, using freely available high
726 resolution satellite images when necessary. A spatial filter was applied to remove
727 polygons smaller than one hectare (1 ha), which was the minimum area mapped in this
728 study.

729

730 **Accuracy assessment**

731 PRODES NF accuracy assessment was implemented using a stratified random sampling
732 by class^{91,92}. We sampled 2,100 points (Table S4) and its validation was carried out on
733 the Temporal Visual Inspection (TVI) platform, an open-source tool that simplifies the
734 visualization of points in Landsat images for long time series and allows to perform
735 analyzes quickly, practically and with simultaneous supervision⁹³. A time series from
736 2000 to 2022 was used, with an image in false color composition (RGB / NIR, SWIR,
737 RED) and the mosaic in the composition (RGB / SWIR, NIR, RED). The overall map
738 accuracy was 0.96 and the natural class have 0.99 of user's and producer's accuracy
739 (Tables S4, S5, S6).

740

741 **Spatiotemporal analysis of NF data**

742 We synthesized the spatiotemporal distribution of NF data in the Brazilian
743 Amazon through maps and graphs, depicting annual state-wise increments and
744 percentages of NF loss as well as the remaining natural NF vegetation. NF maps included
745 the percentage of NF loss for two periods, until 2000 and between 2001 and 2022, based
746 on grids of 10 x 10 km according to previous tests. This strategy was used in order to
747 better represent the spatial distribution of NF loss data in the study area. The grid
748 approach was also derived for three periods (up to 2000, 2002-2010 and 2013-2022)
749 aiming to produce a RGB color composite showing the temporal dynamics of NF loss.
750 The cumulative NF data through time was also compared to deforestation in the Brazilian
751 Amazon. Deforestation data was downloaded from the PRODES digital collections¹⁷.

752 Phytophysiognomy loss was assessed by overlaying NF loss polygons with
753 official Brazilian vegetation mapping data at a 1:250,000 scale⁶⁰. The calculation
754 presented in Table 1 utilized the Legend_1 attribute from the vegetation data, while the
755 aggregated information of the ecotone features (described in the nm_contat legend) was
756 included in Table S3.

757

758 **References**

- 759 1 Almeida, C. A. *et al.* Methodology for Forest Monitoring used in PRODES and DETER
760 projects - 2nd edition (updated). (INPE, São José dos Campos, 2022).
- 761 17 INPE. Terrabrasilis - Plataforma de dados geográficos. (2023).
- 762 60 IBGE. *Vegetação 1:250.000*, <<https://www.ibge.gov.br/geociencias/informacoes-ambientais/vegetacao/22453-cartas-1-250-000.html?=&t=acesso-ao-produto>> (2022).
- 763 89 Gorelick, N. *et al.* Google Earth Engine: Planetary-scale geospatial analysis for
764 everyone. *Remote Sens. Environ.* **202**, 18-27, <https://doi.org/10.1016/j.rse.2017.06.031>
765 (2017).
- 766 90 TerraAmazon, v. 7.3.2 (2023).
- 767 91 Stehman, S. V. Impact of sample size allocation when using stratified random sampling
768 to estimate accuracy and area of land-cover change. *Remote Sensing Letters* **3**, 111-120,
769 <https://doi.org/10.1080/01431161.2010.541950> (2012).
- 770 92 Olofsson, P. *et al.* Good practices for estimating area and assessing accuracy of land
771 change. *Remote Sens. Environ.* **148**, 42-57, <https://doi.org/10.1016/j.rse.2014.02.015>
772 (2014).
- 773 93 Nogueira, S. H. M., Parente, L. L. & Ferreira, L. G. in *XXVII Congresso Brasileiro de*
774 *Cartografia e XXVI Expositocarta*. 624-628.
- 775

Supplementary Files

This is a list of supplementary files associated with this preprint. Click to download.

- [Supplementaryinformation.pdf](#)