

# Stand structural diversity and species with leaf nitrogen conservative strategy drive aboveground carbon storage in tropical old-growth forests

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## Research

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# Abstract

**Background:** Tropical old-growth forest ecosystems are essential for global carbon regulation. Even there are mounting evidences for the significance of species and functional composition, stand structure and elevation gradients on aboveground carbon storage, the relative strengths of these drivers and whether elevation effects via biotic factors are not clear. Furthermore, the mechanisms (the mass-ratio hypothesis or niche complementarity hypothesis) are still poorly understood.

**Methods:** We analyzed aboveground carbon storage, species diversity, stand structural diversity, community-weighted mean (CWM) of functional traits and functional diversity (FDvar) using data from 56 old-growth forest communities with different elevation gradients in Dawei mountain of southwestern China. Multiple regression models were used to test the relative importance of the predictor variables and structural equation model was used to explore the direct and indirect influences on the aboveground carbon storage.

**Results:** Our optimal multiple regression model show aboveground carbon storage is mostly affected by diameter at breast height (DBH) diversity, followed by FDvar of dry matter concentration in mature leaves and CWM nitrogen concentration in young leaves. The final structural equation model indicates elevation indirectly affected aboveground carbon storage via DBH diversity. The stand structural diversity, but not species diversity or functional diversity, enhanced aboveground carbon storage.

**Conclusions:** Our results indicate mass-ratio and niche complementarity effect promote aboveground carbon storage simultaneously. The complex stand structure and species with leaf nitrogen conservative strategy were the crucial drivers of aboveground carbon storage in tropical old-growth forests.

## Background

Tropical forests are essential in global carbon (C) regulation (Saatchi et al. 2011). There are increasing evidence demonstrated the relationships between ecosystem functioning and biodiversity in natural forests (Lohbeck et al. 2016; Yuan et al. 2018a; Li et al. 2019). Although many studies supported the species diversity enhanced the aboveground C storage (Poorter et al. 2015; Li et al. 2019), other studies disagree (Cavard et al. 2010; Ruiz-Jaen et al. 2011; Fotis et al. 2018). The inconsistent results state that the relationships between ecosystem functioning and biodiversity needs to be further clarified (Fotis et al. 2018; Li et al. 2019), especially the multivariate mechanisms simultaneously affecting ecosystem functioning and biodiversity (Li et al. 2019).

The mass ratio hypothesis (Grime et al. 1998) and niche complementarity hypothesis (Tilman et al. 1997) are the two main mechanisms how species diversity would influence ecosystem functioning. The niche complementarity hypothesis demonstrates functional diversity (FDvar) and species diversity could increase the efficiency of resource utilization, thereby increasing C storage (Díaz et al. 2011). The mass ratio hypothesis predicts the main species in the community determined ecosystem functions and can be examined by the correlations between ecosystem functioning and community-weighted mean (CWM)

functional traits (Díaz et al. 2007). And, the species that enhanced ecosystem functioning may be differently dominated by conservative or acquisitive traits (Conti et al. 2013; Prado-Junior et al. 2016; Ali et al. 2017). There are studies supporting the niche complementarity hypothesis (Morin et al. 2011; Zhang et al. 2012; Cavanaugh et al. 2014) and mass ratio hypothesis (Lin et al. 2016; Fotis et al. 2018), which were found not exclusive to each other (Li et al. 2019).

Stand structural diversity was also demonstrated to represent facilitation and niche differentiation drive the ecosystem functions in recent years (Zhang and Chen. 2015; Poorter et al 2015; Yuan et al. 2018a). Multilayered stand structures promote more efficient use of light via better crown space allocation (Yachi and Loreau. 2007). Meanwhile functional diversity and species diversity could affect the ecosystem functioning through stand structural diversity indirectly (Chiang et al. 2016; Ali et al. 2016; Dănescu et al. 2016;). Thus, it is crucial to consider species diversity, stand structural diversity, functional dominance and functional diversity simultaneously to test the mechanisms and the drivers' relative importance (Li et al. 2019).

Abiotic factors were also the important drivers of aboveground C storage through determining plant survival and growth (Sullivan et al. 2017; Li et al. 2019). Microclimate impacts species abundances and distribution (Murphy et al. 2015), which in turn affects both biological and physical stand attributes (Fahey et al. 2015). Thus, abiotic factors could indirectly affect C storage by affecting biotic factors (Fotis et al. 2018; Li et al. 2019). Elevation, considered as a comprehensive factor reflecting climate, negatively affected the ecosystem functioning (Fotis et al. 2018). Meanwhile, Cavanaugh (2014) reported elevation was negatively correlated with species diversity and functional diversity, however, no effect was found on C storage.

In this study, we address three questions using 56 plots in tropical natural forests in southwestern China. First, how biotic (species diversity, stand structural diversity, functional diversity and functional composition) and abiotic (elevation gradients) factors drive aboveground C storage? Second, how these predictor variables relatively drive aboveground C storage. Third, how the predictor variables directly or indirectly influence aboveground C storage. Thus, we hypothesize (1) species diversity, stand structural diversity and functional diversity can enhance aboveground C storage simultaneously (2) CWM functional traits affected aboveground C storage positively or negatively; (3) elevation negatively affects aboveground C storage; elevation can also indirectly affect aboveground C storage via biotic factors.

## Material And Methods

### Site description and plots design

Our study was conducted in Dawei Mountain area (22°35'-23°07' N, 103°20'-104°03' E), located in Yunnan Province, Southwestern China. The annual rainfall is 1700–1900 mm. The annual average temperature is 22.6 °C. The coldest month (January) is 15.2 °C and the hottest month (July) is 27.7 °C. There is no frost in the whole year. The elevation range between 225 m and 2365 m for the highest peak of Dajian Mountain. The forest types change significantly from the bottom to the top of mountain.

In order to ensure the comparability of forest communities in different elevation gradients, only old-growth forests far away from disturbance were selected. A vegetation and soil comprehensive investigation was conducted before plot selection. With the help of local forestry department, four elevation gradients, containing 800 m, 1200 m, 1600 m and 2000 m, respectively were selected and fourteen 20 m × 20 m plots were set with a distance more than 100 m for each plot in each elevation gradient. Total 56 plots were selected. Between 2017–2018, all the individual trees were identified in the plot or lab through the collected specimen to species level (Li et al. 2019). Details of species composition are listed in Annex 1 of Additional file 1.

#### Quantification of aboveground C storage

We measured the diameter at breast height (DBH) of all the trees with DBH higher than 5 cm in each plot. The telescopic pole was used to measure tree height lower than 18 m and the clinometer was used for the height measurement for the other trees. The allometric equation, which was according to height, DBH and wood density, was used to calculate the aboveground biomass of individuals whose DBH were higher than 5 cm (Chave et al. 2014). Aboveground C storage was calculated by multiplying aboveground biomass by 0.5 (Chave et al. 2005). Wood density for most species were obtained from field investigation and other species' wood density were used the wood density of average family or average wood density of corresponding plot (Cavanaugh et al. 2014). The global allometric equation was calculated as follows:

$$AGB = 0.0673 \times (\rho \times DBH^2 \times H)^{0.976} \quad (1)$$

where  $\rho$  represents the wood density ( $\text{g cm}^{-3}$ ),  $H$  represents the height (m), DBH represents diameter at breast height (cm).

#### The quantification of stand structural and species diversity

Species and stand structural (height and DBH) diversity were quantified by Shannon-Wiener biodiversity index (Ali et al. 2016). Recommended by Ali et al (2016), different DBH classes (8, 6, 4, and 2 cm) and height classes (5, 4, 3, and 2 m) were calculated. The proportions of individual species, height class and DBH class was represented by the relative basal area (Finegan et al. 2015; Ali et al. 2016). Because the different discrete height and DBH diversity classes may predict aboveground C storage differently, relationships between aboveground C storage and each class of stand structural diversity were elevated and the lowest AIC values were used to select the classes used for the models in Annex 2 of Additional file 1 (Yuan et al. 2018b).

#### Functional traits

Leaf nitrogen concentration (LNC), leaf phosphorus concentration (LPC), N:P ratio, specific leaf area (SLA), leaf dry matter content (LDMC) and leaf thickness (LT), which were crucial for plant survival and growth (Wright et al. 2010; Finegan et al. 2015; Ali et al. 2017), were measured of all the species in the

plots in both young and mature leaves, on account of the central plant trade-offs strongly correlating with leaf longevity (Aerts et al. 2000; Ali et al. 2017). All the leaf functional traits were measured based on standard measurement methods (Pérez-Harguindeguy et al. 2013).

The calculation of CWM trait values in each plot were based on the following formula (Conti and Díaz., 2013; Ali et al., 2017):

$$\text{CWM}(\text{trait}_x) = \sum_{i=1}^s p_i \times x_i \quad (2)$$

where CWM (trait<sub>x</sub>) represents the CWM X trait, s represents the species number in each plot, p<sub>i</sub> represents i th species' relative abundance in the plot and x<sub>i</sub> represents the i th species' trait value.

Functional diversity was calculated as follows (Conti and Díaz. 2013):

$$\text{FDvar} = 2/\pi \arctan(5V) \text{ and } V = \sum_{i=1}^s p_i (\ln x_i - \ln \bar{x})^2 \quad (3)$$

where  $p_i = a_i / \sum_{i=1}^s a_i$ , a<sub>i</sub> represents i th species' relative abundance in the plot, x<sub>i</sub> represents the i th

species' trait value,  $\ln \bar{x} = \sum_{i=1}^s p_i \ln x_i$ . The variation range of FDvar values is 0 to 1.

## Statistical analyses

One-way ANOVA and least square difference (LSD) multiple comparison tests were used to test the difference in aboveground C storage and abiotic factors among different elevations (Annex 3 of Additional file 1). Pearson correlation coefficients were employed to investigate the relationships among aboveground C storage, species diversity, functional dominances, functional diversity, stand structural diversity and elevation in Annex 1 of Additional file 2 (Li et al. 2019). Shapiro-Wilk test was used to examine the normality of all the data (Zhang et al. 2012). Non-normal continuous variables were natural-logarithm-transformed to improve normality and linearity before the data statistics (Zhang et al. 2012; Li et al. 2019).

Simple and multiple linear regressions were employed to explore the relationships between aboveground C storage and explanatory variables (Li et al. 2019). The combination of 28 indices yields total 268435455 models, which were beyond the capacity of R (Ali et al. 2017). To select the optimal subset of predictor variables of the aboveground C storage, ordinary least squares multiple regression analyses was first conducted with all 12 CWM indices, resulting 4095 possible models. Then we used the other 15 biotic indices (FDvar, species diversity and stand structural diversity) to conduct multiple regression analyses, which resulted 32367 possible models. Finally, we combined the elevation with the two former best subsets of predictors to conduct multiple analyses and the 8 predict variables resulted 255 possible models. All possible models were evaluated with corrected Akaike information criterion (AICc) and the

best-fit regression model were selected based on the lowest AICc values (Li et al. 2019; Wen et al. 2019). Multicollinearity was diagnosed by variance inflation factor (VIF) and all VIF values were lower than 10 in the best-fit regression model, which suggested our results were not affected by collinearity among predictor variables (Graham 2003). Detailed statistics of all models are provided in Annex 2 of Additional file 2, respectively. The 'glmulti' package in R 3.6.1 was used to select models (Calcagno 2013).

Structural equation model (SEM), which was widely used to explore the complex relationships between ecosystem functioning and predictor variables (van der Sande et al. 2017), were conducted to invest how the biotic and abiotic factors affect aboveground C storage directly or indirectly. The variables retained in the best-fit regression model were used to construct SEM (Ali et al. 2017; Wen et al. 2019). The comparative fit index (CFI), goodness-of-fit index (GFI), root mean square error of approximation (RMSEA), Chi-square ( $\chi^2$ ) test and AIC were employed to test the fitness of SEM (Zhang and Chen. 2015). The SEM were implemented using the AMOS 21.0 software.

## Results

### Correlations between aboveground C storage and influencing factors

Aboveground C storage was most strongly affected by DBH ( $R^2 = 0.413$ ,  $P < 0.001$ ) and height ( $R^2 = 0.207$ ,  $P < 0.001$ ) diversity (Figure. 1). The CWM nitrogen concentration of young leaves negatively affected aboveground C storage ( $R^2 = 0.08$ ,  $P = 0.034$ ). Aboveground C storage decreased with FDvar of LDMC and SLA in both young and mature leaves. The FDvar of phosphorus concentration in young leaves ( $R^2 = 0.072$ ,  $P = 0.045$ ) and elevation gradients ( $R^2 = 0.067$ ,  $P = 0.03$ ) also negatively affected aboveground C storage.

### The relative importance of all predictor variables for aboveground C storage

Only 4 predictor variables were retained in the optimal multiple regression model, account for 57.21% of the variation in aboveground C storage (Table 1). DBH diversity still most importantly affected aboveground C storage ( $P < 0.001$ ,  $\beta = 0.47$ ). The FDvar of leaf dry matter content in mature leaves ( $P = 0.009$ ,  $\beta = -0.27$ ) and CWM nitrogen concentrations of young leaves ( $P = 0.011$ ,  $\beta = -0.25$ ) both negatively affected aboveground C storage significantly. The height diversity didn't significantly affect the aboveground C storage in the final multiple regression model ( $P = 0.081$ ,  $\beta = -0.18$ ).

Table 1  
The optimal model resulted from a sequence of regression analyses of aboveground carbon storage.

| Model and predictor  | Coeff. | Beta  | <i>t</i> | <i>P</i>          | <i>R</i> <sup>2</sup> | AICc   |
|--|--------|-------|----------|-------------------|-----------------------|--------|
| <b>Effects of community-weighted mean (CWM)</b>  |        |       |          |                   |                       |        |
| <i>Model</i> <sup>1</sup>  |        |       |          | <b>&lt; 0.001</b> | 0.28                  | 105.52 |
| Constant   | 9.20   | 0     | 7.21     | <b>&lt; 0.001</b> |                       |        |
| CWM SLA-Y  | -0.01  | -0.39 | -2.33    | <b>0.024</b>      |                       |        |
| CWM LT-Y   | -2.66  | -0.46 | -3.21    | <b>0.002</b>      |                       |        |
| CWM LNC-Y  | -0.02  | -0.34 | -2.70    | <b>0.009</b>      |                       |        |
| CWM LDMC-M   | -5.35  | -0.44 | -0.28    | <b>0.007</b>      |                       |        |
| <b>Effects of functional diversity (FDvar), species diversity and stand structural diversity</b> |        |       |          |                   |                       |        |
| <i>Model</i> <sup>2</sup>  |        |       |          | <b>&lt; 0.001</b> | 0.51                  | 80.80  |
| Constant   | 3.35   | 0     | 14.36    | <b>&lt; 0.001</b> |                       |        |
| FDvar LDMC-M   | -1.17  | -0.27 | -2.59    | <b>0.012</b>      |                       |        |
| DBH Diversity  | 1.85   | 0.49  | 4.24     | <b>&lt; 0.001</b> |                       |        |
| Height Diversity   | 0.53   | 0.17  | 1.60     | 0.116             |                       |        |
| <b>Joint effect of all predict variables</b>   |        |       |          |                   |                       |        |
| <i>Model</i> <sup>3</sup>  |        |       |          | <b>&lt; 0.001</b> | 0.57                  | 76.18  |
| Constant   | 4.14   | 0     | 11.15    | <b>&lt; 0.001</b> |                       |        |
| CWM LNC-Y  | -0.02  | -0.25 | -2.65    | <b>0.011</b>      |                       |        |
| FDvar LDMC-M   | -1.16  | -0.27 | -2.73    | <b>0.009</b>      |                       |        |
| DBH Diversity  | 1.78   | 0.47  | 4.30     | <b>&lt; 0.001</b> |                       |        |
| Height Diversity   | 0.56   | 0.18  | 1.78     | 0.081             |                       |        |

Note: The regression coefficient (Coeff.), standardized regression coefficient (Beta), t-test and *P*-value are given for each predictor variable. For the model, the coefficient of determination (*R*<sup>2</sup>), *P*-value and Akaike Information Criterion (AICc) are also given. *P* values < 0.05 are given in bold.

The direct and indirect impacts on aboveground C storage

The optimal SEM accounted for 53, 13, 5, 2 and 1% of the variation in aboveground C storage, DBH diversity, height diversity, FDvar of LDMC in mature leaves and CWM nitrogen concentration in young

leaves (Figure. 2). The DBH diversity directly enhanced aboveground C storage significantly ( $P = 0.002$ ,  $\beta = 0.48$ ), whereas CWM young leaves' nitrogen concentration ( $P = 0.009$ ,  $\beta = -0.27$ ) and FDvar of mature leaves' dry matter content ( $P = 0.034$ ,  $\beta = -0.27$ ) both had significant negative direct effect (Figure. 2, Table 2). Elevation negatively affected DBH diversity directly ( $P < 0.001$ ,  $\beta = -0.37$ ) and didn't affect aboveground C storage directly ( $P = 0.906$ ,  $\beta = -0.02$ ). However, elevation had a significant negative indirect effect via DBH diversity ( $P = 0.002$ ,  $\beta = -0.18$ ). The total effect of elevation was significant ( $P = 0.044$ ,  $\beta = -0.30$ ).

Table 2  
The direct, indirect, and total standardized effects on aboveground carbon storage based on structural equation model.

| Predictor        | Pathway to aboveground carbon storage | Effect | P-value |
|------------------|---------------------------------------|--------|---------|
| Elevation        | Direct effect                         | -0.02  | 0.906   |
|                  | Indirect effect via DBH diversity     | -0.18  | 0.002   |
|                  | Indirect effect via height diversity  | -0.04  | 0.057   |
|                  | Indirect effect via CWM LNC-Y         | -0.02  | 0.412   |
|                  | Indirect effect via FDvar LDMC-M      | -0.04  | 0.245   |
|                  | Total effect                          | -0.30  | 0.044   |
| DBH Diversity    | Direct effect                         | 0.48   | < 0.001 |
| Height Diversity | Direct effect                         | 0.19   | 0.057   |
| CWM LNC-Y        | Direct effect                         | -0.27  | 0.003   |
| FDvar LDMC-M     | Direct effect                         | -0.27  | 0.004   |

Note: The indirect effect of elevation was calculated by multiplying the standardized effects of all paths on one route, from elevation to mediator, and then to aboveground carbon storage, while the total effect was calculated by adding standardized direct and indirect effects.

## Discussion

Biodiversity promoting ecosystem functioning has long been debated during last two decades (Forrester and Bauhus. 2016). Inconsistent with the hypothesis, no correlation between aboveground C storage and species diversity was found, which may be caused by a potent effect of the dominant productive species (Cardinale et al. 2011; Tobner et al. 2016; Fotis et al. 2018) or competition exclusion (Grime. 1973; Grace et al. 2016; Ali et al. 2016). In natural forests, species richness may achieve saturation and previous studies have reported ecological niches of the most species were overlapping, which may reduce the influence of species diversity on ecosystem functioning (Silver et al. 1996; Hooper et al. 2005; Ruiz-Jaen et al. 2011). As expected, both the DBH and height diversity promoted ecosystem functioning, consistent



with recent researches (Dănescu et al. 2016; Ali et al. 2016; Forrester and Bauhus. 2016; Fotis et al. 2018). These may result from multilayered forest structure enhancing the capture of water, light, and soil nutrients (Forrester. 2014; Poorter et al. 2015). No correlation was found between stand structure diversity and species diversity (Annex 1 of Additional file 2), inconsistent with the hypothesis and previous results (Zhang and Chen. 2015; Ali et al. 2016). The variations among and within species could result in the differences of stand structure (Clark 2010; Zhang et al. 2012) and this result potentially stressed the importance of variations within species.

Unexpected, significant negative associations between FDvar of leaf functional traits and aboveground C storage were found, which was not consistent with the niche complementarity hypothesis and some previous results (Ali et al. 2016). Functional diversity is positively related to species diversity (Annex 1 of Additional file 2) and low trait diversity maybe caused by the large dominated species (Van Gelder et al. 2006). The results provide a potential solution to stand structure (Zhang and Chen 2015; Dănescu et al. 2016; Forrester and Bauhus 2016; Ali et al. 2016; Fotis et al. 2018) as the niche complementarity mechanism to maintenance ecosystem functioning, but not the species diversity (Cavard et al. 2010; Ruiz-Jaen et al. 2011; Ali et al. 2016) or the functional diversity (Conti et al. 2013; Cavanaugh et al. 2014; Finegan et al. 2015; Fotis et al. 2018).

The CMM nitrogen concentration of young leaves negatively related to aboveground C storage, consistent with our hypothesis and previous studies (Finegan et al. 2015; Ali et al. 2017). Our result indicates mass ratio hypothesis is essential for promoting ecosystem functioning (Finegan et al. 2015) and plant's leaf economic spectrum correlated with the aboveground C storage (Garnier et al. 2004). The high aboveground C storage was accompanied by low CWM leaf nitrogen concentration, which indicate "slow" forests result in high aboveground C storage accumulation in old-growth forests (Quesada et al. 2012). Conservative resource use syndromes were associated with low C losses through mortality (Wright et al. 2010) or decomposition (Wright et al. 2004; Conti and Díaz 2013;). In old-growth communities, the adaptations that forests were dominated by conservative species might result in less biomass turnover and more aboveground C storage (Ruiz-Jaen et al. 2011; Conti and Díaz 2013).

As expected, the aboveground C storage decreased significantly with the elevation gradient increasing, which was consistent with previous results (Fotis et al. 2018). The DBH diversity, CWM SLA and LPC decreased and CWM LT increased with the elevation increasing (Annex 4 of Additional file 1), which indicated the elevation had significant filtering effect (Grime. 2006). These leaf conservative resource use syndromes and low stand structural diversity at higher elevations may result from adaptations of plants to the harsh environment (Read et al. 2014). As hypothesized, the elevation gradient can also indirectly affect the aboveground C storage via the stand structural diversity, which was consistent with previous studies where climate factors indirectly affected aboveground C storage via forest compositions (van der Sande et al. 2017; Li et al. 2019). Unexpected, we didn't find elevation indirectly affect aboveground C storage via or species diversity or leaf functional traits.

We found the stand structure diversity was the most important driver of aboveground C storage, followed by CWM LNC and FDvar of leaf traits. This result indicates niche complementarity may have a more important effect on promoting ecosystem functioning than mass-ratio effect. Both consideration the niche complementarity and mass-ratio effect simultaneously can better explain the mechanisms of ecosystem function (Table 1). However, it must be noted that some other functional traits, such as plant diameters (Cavanaugh et al. 2014; Finegan et al. 2015), wood traits (Ali et al. 2017) and leaf anatomical traits (He et al. 2016) affecting ecosystem functioning were not measured in this study. And dynamic of other ecosystem functioning (e.g. aboveground productivity) maybe not consistent with aboveground C storage (Chisholm et al. 2013). Meanwhile, ecosystem functioning and biotic factors changed with succession (Lohebeck et al. 2013; Ali et al. 2016; Ali et al. 2017) and other abiotic factors, such as soil nutrients (Zhang and Chen. 2015; Yuan et al. 2018a) and climate factors (Li et al. 2019). Thus, we encourage more studies investigating the relationships between ecosystem functioning and biodiversity taking into consideration the predictor variables simultaneously in communities of different stages of succession.

## Conclusions

The data from 56 old-growth forest plots indicate both niche complementarity and mass-ratio effect are essential in promoting ecosystem functioning. The stand structural diversity strongly enhanced the aboveground C storage, but not the species diversity or functional diversity. The species dominated by leaf nitrogen conservative strategy drive the aboveground C storage. Elevation has a significant filtering effect on DBH diversity and CWM leaf functional traits. The elevation indirectly affected the aboveground C storage via DBH diversity. Our results indicated complex stand structure and species with leaf nitrogen conservative use strategy can maximize aboveground C storage in tropical old-growth forests.

## Abbreviations

AGC: aboveground carbon storage; CWM: community-weighted mean; FDvar: functional diversity; LDMC: leaf dry matter content; SLA: specific leaf area; LT: leaf thickness; LNC: leaf nitrogen concentration; LPC: leaf phosphorus concentration; N:P: N:P ratio; Y: young leaves; M: mature leaves; DBH: diameter at breast height.

## Declarations

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### Authors' contributions

JZ, GW and YL collected and analyzed the data and were major contributors in writing the manuscript. XW, DP, XY and AH made substantial contributions to the conception and design, the interpretation of data and writing of the manuscript. All authors read and approved the final manuscript.

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## **Availability of data and materials**

The datasets analyzed during the current study are available from the corresponding author upon request.

## **Ethics approval and consent to participate**

Not applicable.

## **Consent for publication**

Not applicable.

## **Competing interests**

The authors declare that they have no competing interests.

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## Supplementary Information

### Additional File 1

**Annex 1** Species composition in old-growth forests

**Annex 2** General linear relationship between aboveground C storage and stand structural (tree diameter at breast height (DBH) and height) diversity of different classes across the plots. The bold discrete classes were selected.

**Annex 3** Community characteristics across tropical mature forests in China. Mean and range (in parentheses). Lowercase letters represent significant pairwise differences between different elevation gradients ( $P < 0.05$ ).

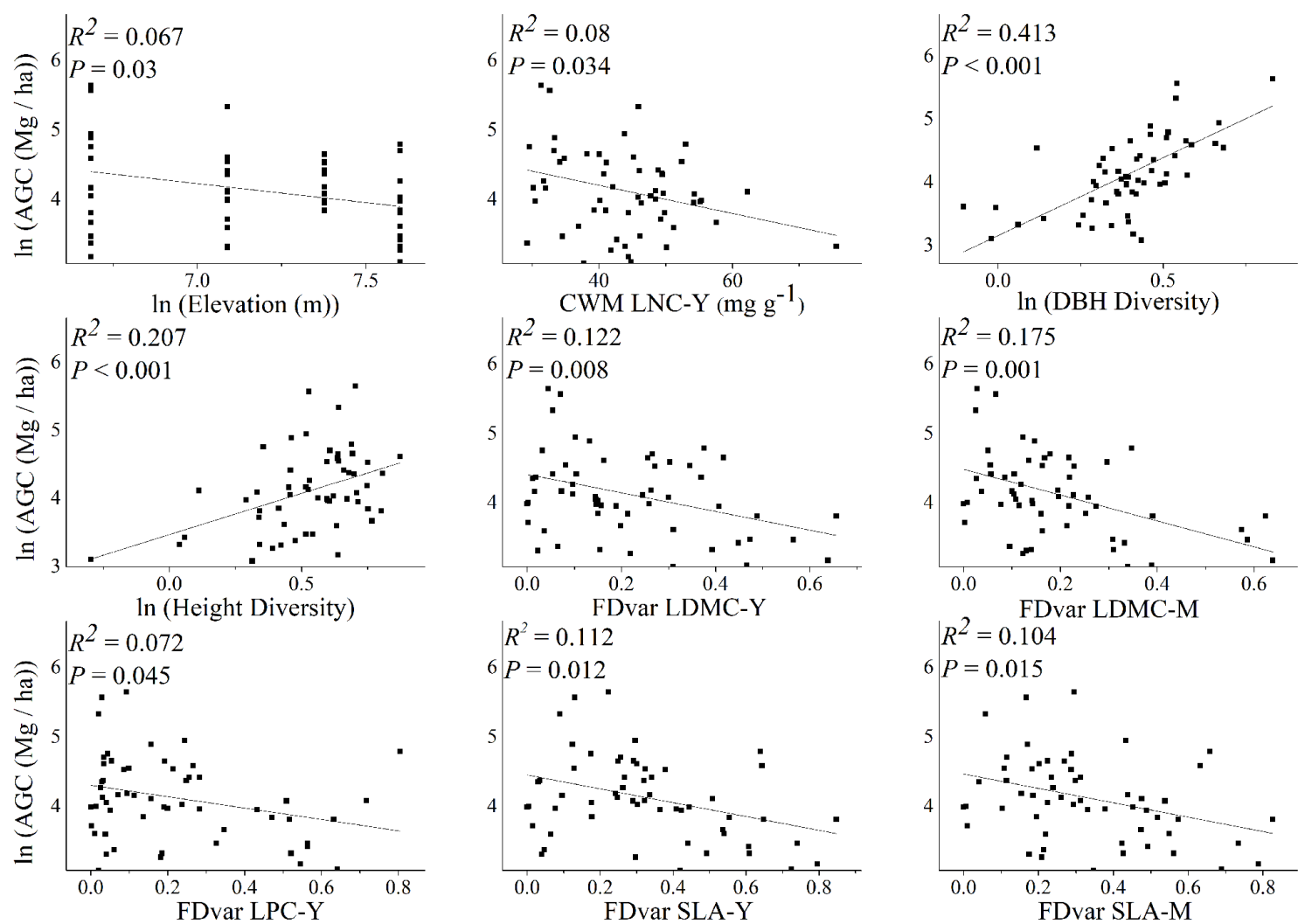
**Annex 4** Bivariate relationships between elevation (dependent) and exogenous (independent) variables (n = 56). Only significant relationships are shown here ( $P < 0.05$ ).

**Additional File 2**

**Annex 1** Pearson correlation coefficients for pairwise comparisons between aboveground carbon storage and species diversity, stand structural diversity, functional diversity, functional compositions and elevation gradients, as well as for pairwise correlations between all independent variables.

**Annex 2** Results for the Model Selection procedure, based on Akaike Information Criterion (AICc).

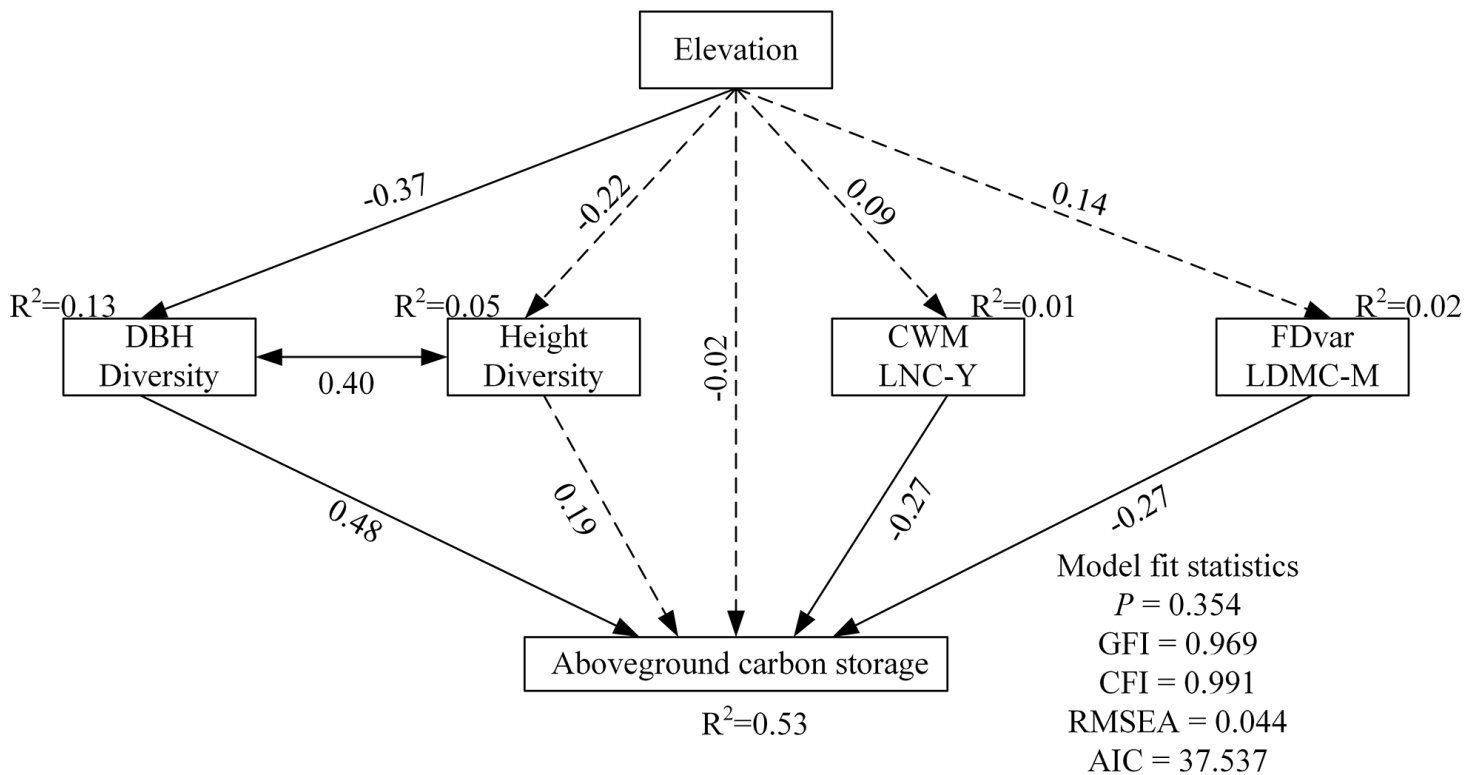
**Figures**



**Figure 1**

Significant bivariate relationships between aboveground carbon storage (AGC) and predictor variables (n = 56).





**Figure 2**

The optimal structural equation model relating aboveground carbon storage to predictor variables. Model-fit statistics and standardized regression coefficient of each path are shown in the figure. Significant routes and non-significant routes at the level of 0.05 were represented by solid arrows and dashed arrows, respectively. The total variation of the explained variable was showed as  $R^2$ , which is explained by all the explanatory variables.

## Supplementary Files

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