

# Sustainable-use marine protected areas to improve human nutrition

**Daniel Viana** (✉ [dviana@hsph.harvard.edu](mailto:dviana@hsph.harvard.edu))

Harvard T.H. Chan School of Public Health

**David Gill**

Duke

**Alex Zvoleff**

Conservation International

**Nils Krueck**

University of Tasmania

**Jessica Zamborain-Mason**

Harvard University <https://orcid.org/0000-0002-4705-0166>

**Christopher Free**

University of California Santa Barbara

**Alon Shepon**

Tel Aviv university <https://orcid.org/0000-0002-4345-8957>

**Michael Mascia**

Conservation International <https://orcid.org/0000-0002-9874-9778>

**Dana Grieco**

Duke University

**Josef Schmidhuber**

Food and Agriculture Organization

**Christopher Golden**

Harvard T.H. Chan School of Public Health <https://orcid.org/0000-0002-2258-7493>

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3 Daniel F. Viana<sup>1,2\*</sup>, David Gill<sup>3</sup>, Alex Zvoleff<sup>2</sup>, Nils C. Krueck<sup>4</sup>, Jessica Zamborain-Mason<sup>1</sup>,  
4 Christopher M. Free<sup>5,6</sup>, Alon Shepon<sup>7</sup>, Michael B. Mascia<sup>3</sup>, Dana Grieco<sup>2</sup>, Josef Schmidhuber<sup>8</sup>,  
5 Christopher D. Golden<sup>1,9,10</sup>

## 6 **Affiliations:**

7 <sup>1</sup> Department of Nutrition, Harvard T.H. Chan School of Public Health, Boston, MA 02115, USA

8 <sup>2</sup> Moore Center for Science, Conservation International, Arlington, VA, USA

9 <sup>3</sup> Duke University Marine Laboratory, Nicholas School of the Environment, Duke University,  
10 Beaufort, NC 28516, USA

11 <sup>4</sup> Institute for Marine and Antarctic Studies (IMAS), University of Tasmania, Hobart, Tasmania,  
12 7001, Australia

13 <sup>5</sup> Bren School of Environmental Science and Management, University of California, Santa  
14 Barbara, Santa Barbara, CA, USA

15 <sup>6</sup> Marine Sciences Institute, University of California, Santa Barbara, CA, USA

16 <sup>7</sup> Department of Environmental Studies, The Porter School of the Environment and Earth  
17 Sciences, Tel Aviv University, Tel Aviv, Israel

18 <sup>8</sup> Division of Markets and Trade, Food and Agriculture Organization, Rome, Italy

19 <sup>9</sup> Department of Environmental Health, Harvard T.H. Chan School of Public Health, Boston, MA  
20 02115, USA

21 <sup>10</sup> Department of Global Health and Population, Harvard T.H. Chan School of Public Health,  
22 Boston, MA 02115, USA

## 23 **Summary**

24 Coral reef fisheries are a vital source of nutrients for thousands of nutritionally vulnerable  
25 coastal communities around the world. Here, we evaluated the potential effects of expanding  
26 sustainable-use marine protected areas (MPAs) to improve the nutrition of coastal  
27 communities. Using information from underwater visual surveys from 2,518 sites located in 53  
28 countries, we developed a Bayesian hierarchical model to estimate the average effect of  
29 existing sustainable-use MPAs reef fish biomass and explored how that may alter fish catch, and  
30 the nutrients supplied to local communities. We then estimated the potential nutritional  
31 benefits of expanding sustainable-use MPAs to all non-MPA coral reefs globally. We found that  
32 existing sustainable use MPAs have on average 15% more biomass than open access reefs.  
33 Translating this into catch, we estimated that expanding sustainable-use MPAs could increase

34 catch potential by 0-20%, which could prevent 0.53-1.95 million cases of inadequate  
35 micronutrient intake globally, a fraction of the people who would continue to be sustained by  
36 this foundation of coastal food systems. Our study estimates the potential nutritional benefits  
37 of expanding sustainable-use MPAs and pinpoints locations with the greatest potential to  
38 reduce inadequate micronutrient intake levels, critical knowledge given the strong international  
39 movement to cover 30% of our oceans with MPAs by 2030.

## 40 Main text

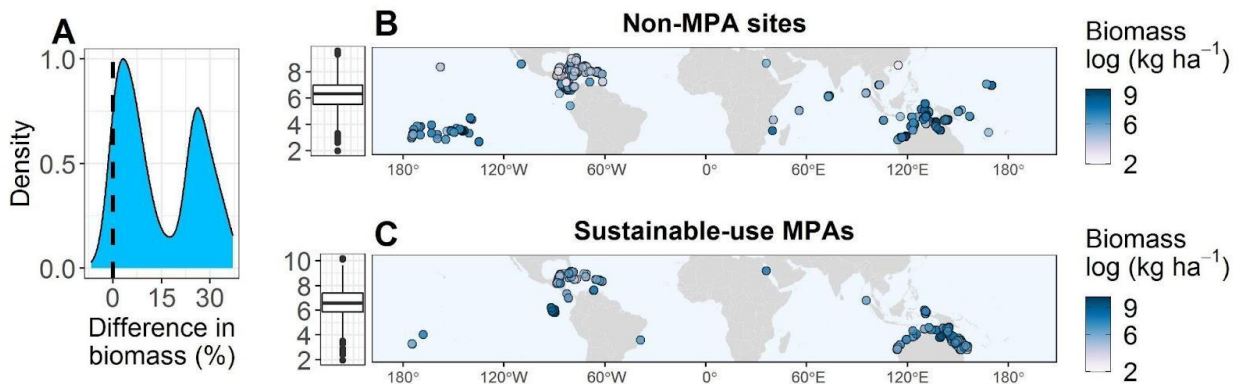
41 Over 2 billion people are unable to access safe, nutritious and sufficient supplies of food, which  
42 threatens human health globally<sup>1</sup>. Coastal residents of tropical developing countries, where  
43 coral reef systems are a vital source of critical micronutrients, vitamins, and fatty acids for  
44 millions of people<sup>2,3</sup>, are particularly vulnerable to nutritional deficiencies<sup>4-6</sup>. Yet, coral reefs  
45 around the world are being severely degraded by pollution, overfishing, and climate change,  
46 imperiling marine biodiversity and the health of millions of people<sup>7</sup>. Policies governing coral  
47 reefs that attempt to address these threats not only shape the future of these ecosystems but  
48 also the health of those who depend upon them.

49 One important management tool rapidly being implemented globally as a measure to  
50 recover coral reef ecosystems and associated fisheries from depletion is marine protected areas  
51 (MPAs)<sup>8</sup>. MPAs are areas of the ocean with specific rules and restrictions designed to protect  
52 marine ecosystems from anthropogenic threats<sup>9,10</sup>. MPAs can be broadly categorized into “no-  
53 take” areas where fishing is prohibited and sustainable-use areas where different forms of  
54 fishing restrictions are in place (referred to here as “sustainable-use MPAs”). Both MPA types  
55 generally have significant positive effects on the biomass of fish within their boundaries  
56 compared to neighboring non-MPA areas<sup>11</sup>. Currently, there are about seventeen thousand  
57 MPAs worldwide, covering about 7.7% of the ocean (MPAtlas, 2021). Despite failing to reach  
58 the globally agreed target of 10% MPA coverage by 2020 (CBD 2011), there is strong  
59 international momentum to set a new target to cover 30% of our oceans with MPAs and other  
60 effective area-based conservation measures (OECMs) by 2030<sup>12-14</sup>.

61 Here, we quantified the potential human nutritional benefits that can arise from  
62 increased access to rehabilitated seafood stocks from sustainable-use MPA implementation.  
63 For this, we compiled information on coral reef fish populations and social and environmental  
64 conditions from 2,518 reef sites in 53 countries, 804 of which were in sustainable-use MPAs, to  
65 estimate the effect of these areas on standing reef fish biomass and the potential nutrient  
66 availability for tropical coastal communities. Our analyses were based on a Bayesian  
67 hierarchical model that estimated the expected standing reef fish biomass under non-MPA and  
68 sustainable-use MPAs conditions, accounting for other social (e.g., human population, market  
69 distance, fisheries governance, human development index) and environmental (e.g.,  
70 productivity, depth, temperature, wave exposure; see Supplement for details) variables. We  
71 then used the model to estimate: 1) the effect of sustainable-use MPAs on standing reef fish  
72 biomass, 2) expected biomass and catch for existing sustainable-use MPAs, and 3) potential  
73 changes in nutritional inadequacies associated with an expansion of sustainable-use MPAs to all  
74 non-MPA reefs.

75 Sustainable-use MPAs increase biomass

76 We estimated the potential net conservation benefits of sustainable-use MPA establishment by  
77 examining the estimated effect size of sustainable-use MPAs on reef fish biomass (Figure 1A).  
78 Globally, we found that sustainable-use MPAs have on average 15% more biomass than non-  
79 MPA sites (Figure 1A), although percentage increases were dependent on associated variability  
80 in physical, environmental, and social conditions in the model and the effectiveness of fisheries  
81 management<sup>15</sup> (Figures 1B and 1C). Locations with effective fisheries management were  
82 expected to have lower effect sizes because the difference between open access and  
83 sustainable-use MPA biomass is likely small. In comparison, locations with low management  
84 effectiveness were expected to have greater biomass gains in sustainable-use MPAs. Other  
85 known drivers of variation in MPA biomass include variability in MPA design, MPA size and age,  
86 and starting conditions, among others<sup>11,16,17</sup>.



87 **Figure 1 - Biomass in sustainable-use Marine Protected Areas (MPAs), highlighting A)**  
88 Distribution of percent differences in biomass of sustainable-use MPAs compared to non-MPA  
89 sites, B) observed biomass in non-MPA sites (n=1,117), and C) observed biomass in sustainable-  
90 use MPA sites (n=804). The distributions of both values are indicated as box plots on the left-  
91 hand side of panels A and B.

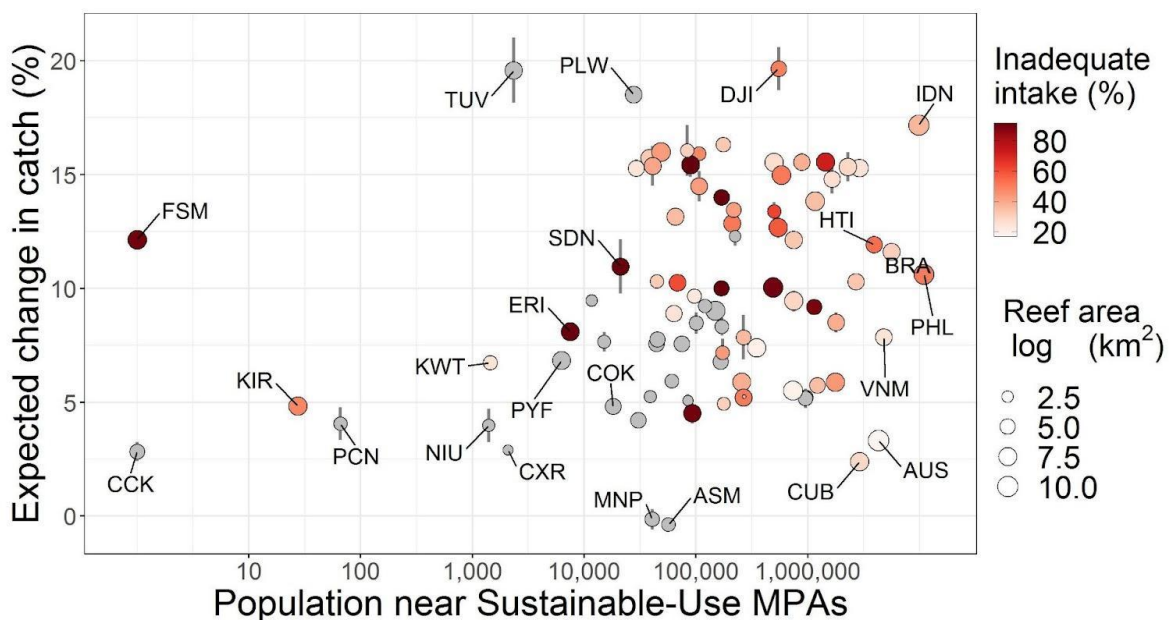
93 Expected catch benefits from existing MPAs

94 We estimated the potential changes in fisheries catch that could result from increases in  
95 biomass in existing sustainable-use MPAs. To do so, we used 804 existing coral reef sustainable-  
96 use MPAs distributed in 32 countries. Overlaying MPA boundary data from MPAtlas<sup>18</sup> with reef  
97 polygons<sup>19</sup>, we found that 37% of all coral reefs in the world are within sustainable-use MPAs,  
98 11% are within no-take MPAs and 51% are non-MPA reefs (Figure S1). We then used our  
99 Bayesian model to predict the potential biomass density in existing sustainable-use MPAs  
100 compared to predicted non-MPA conditions while accounting for local social and environmental  
101 conditions.

102 To estimate the association between standing reef fish biomass and catch within these  
103 coral reef MPAs, we used a simple Graham-Schaefer surplus production model, where the  
104 sustainable harvest rate was dependent on the predicted reef fish biomass and an estimated  
105 population-level intrinsic growth rate for fish of 0.23<sup>20</sup>. We then assessed the expected benefits  
106 in catch from existing sustainable-use MPAs compared to non-MPA conditions and found a

107 mean predicted increase in catch of 12%, ranging from 0 - 20 % (Figure 2). Expected catch  
 108 benefits should be interpreted as the potential sustainable catch benefits derived from existing  
 109 MPAs.

110 Based on the risk of inadequate nutrient supply in countries' overall food systems<sup>6</sup>,  
 111 many existing MPAs are in areas with large coastal populations at high risk of inadequate  
 112 nutritional intake such as Indonesia, Philippines and Haiti and where expected change in catch  
 113 is the highest (Figure 2). Inadequate intake values range from 0% to 100% and should be  
 114 considered as a risk of nutritional inadequacies, with higher values representing larger  
 115 population prevalence predicted to experience inadequate micronutrient intake<sup>21</sup>. We  
 116 calculated inadequate intake by comparing per capita nutrient supply of overall food systems in  
 117 each country against age- and sex- specific nutrient demands (see Methods).



118 **Figure 2 – Expected catch benefits from existing sustainable-use MPAs to support coastal**  
 119 **populations' nutrient intake.** Expected catch effects in all existing sustainable-use MPAs  
 120 relative to expected catch under non-MPA conditions, plotted on a log-scale for country  
 121 populations within 10km of an MPA. Error lines represent uncertainty around estimates of  
 122 change in catch. Inadequate nutrient intake is the average prevalence across key nutrients  
 123 found in aquatic species in 2017 (iron, EPA+DHA, calcium, zinc, and vitamins A and B<sub>12</sub>). Points  
 124 in gray represent countries with no data on prevalence of nutrient intake.

127 Expanding sustainable-use MPAs to reduce nutritional inadequacies

128 We predicted the potential nutritional gains of expanding sustainable-use MPAs to all non-MPA  
 129 reefs (total area of 70,003 km<sup>2</sup>, Figure 3A), by (i) calculating the estimated total number of  
 130 people nutritionally supported by existing and future MPAs, and (ii) calculating the potential  
 131 change in the prevalence of inadequate nutrient intake with MPA expansion. First, we defined  
 132 nutritional support from MPAs as the provision of at least 5% of aquatic animal source food

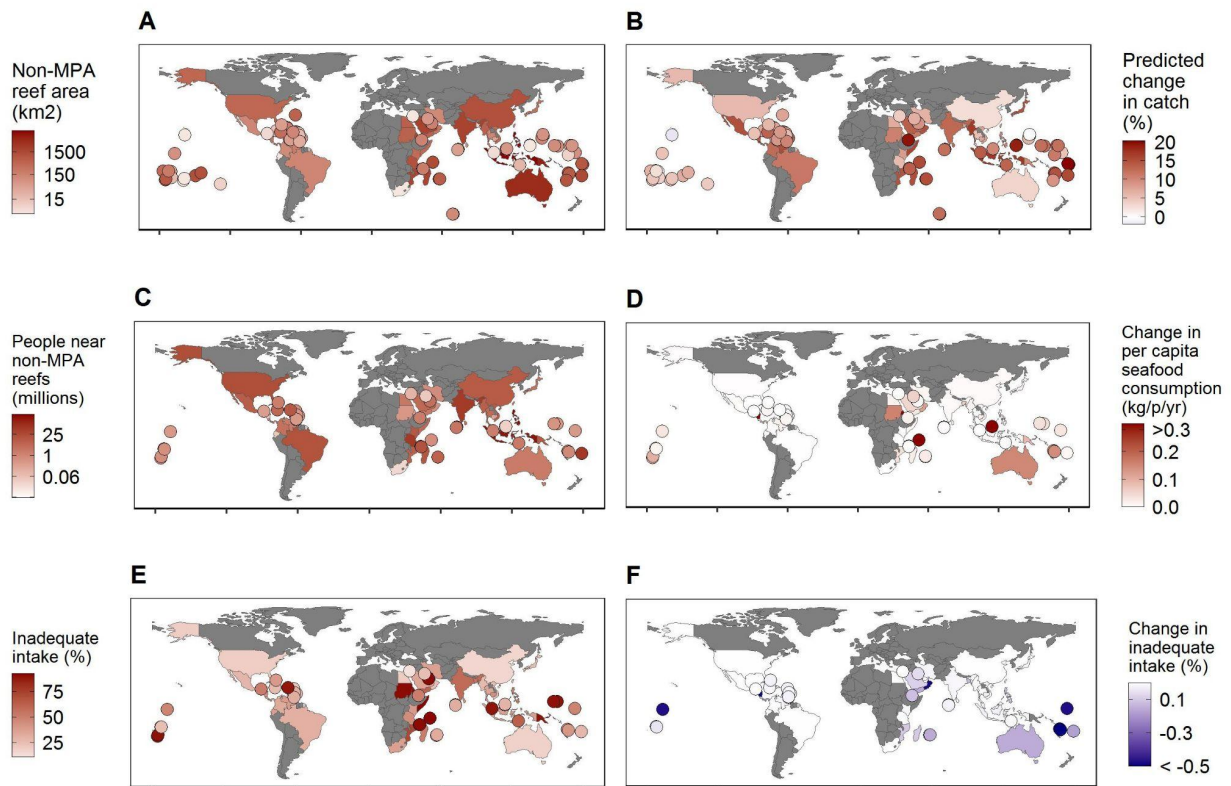
133 intake from coral reef catch (see Figure S8 for sensitivity analysis). We then summed the  
134 population near reefs above this threshold to estimate the total number of people supported  
135 by MPAs.

136 Next, we calculated the potential change in nutrient supply attributed to changes in  
137 catch from MPA expansion (Figure 3B). To account for geographical differences in reef catch  
138 composition from environmental conditions and human behavior, we used the Sea Around Us  
139 (SAU) database<sup>22</sup> to assign species catch composition proportions based on their reported data  
140 from 2017. This portfolio of species caught were then cross-referenced with the Aquatic Food  
141 Composition Database<sup>6</sup> to assign nutrient composition to each species, and to estimate overall  
142 nutrient supply from the potential catch. Differences in nutrient supply attributed to predicted  
143 changes in catch were then added to the overall diets of populations living within a 10km buffer  
144 around reefs (see supplement for sensitivity analysis) to predict the potential changes in  
145 inadequate nutrient intake from sustainable-use MPA establishment. In addition, we calculated  
146 the total number of people nutritionally supported by MPAs.

147 We found that expanding sustainable-use MPAs to non-MPA locations could increase  
148 catch in many nutritionally vulnerable countries (inadequate intake higher than 25%), with  
149 potential positive impacts on human nutrition and health. On average, catch could increase by  
150 12% when considering all countries or 15 % (from 2-20%) if we only consider nutritionally  
151 vulnerable countries. Globally, the expansion of sustainable-use MPAs could lead to reductions  
152 in inadequate intake across all assessed nutrients for 0.53 - 1.95 million individuals (reduction  
153 of 0.2-0.9 million vitamin B<sub>12</sub>, 0.1-0.5 million calcium, 0.07-0.2 million iron, 0.06-0.2 million  
154 vitamin A, 0.05-0.15 million omega-3 long-chain polyunsaturated fatty acids (specifically  
155 DHA+EPA), and 0.02-0.08 million zinc inadequate intakes). Beyond preventing nutritional  
156 deficiencies, sustainable-use MPAs could also maintain the support of nutritional needs for 2.8  
157 to 30 million people by substantially contributing to overall aquatic animal-source food intake.  
158 Countries such as Madagascar, Mozambique, Kiribati, Yemen, and Solomon Islands have the  
159 highest potential reductions of inadequate intake following implementation of effective  
160 sustainable-use MPAs (Figure 3F). Other countries, including Seychelles and Sudan, have high  
161 potential changes in per capita seafood consumption but cannot be modeled in terms of  
162 inadequate intake because of a lack of catch or baseline nutrient supply data.

163 Four major factors drive the extent of potential nutritional impacts of sustainable-use  
164 MPA expansion in our study, (i) non-MPA reef area (Figure 3A), (ii) population size near non-  
165 MPA coral reefs (Figure 3C), (iii) the prevalence of inadequate intake within coastal  
166 communities (Figure 3E), and (iv) the efficacy of fisheries management. First, the larger the area  
167 of non-MPA reef (Figure 4A), the larger the potential for MPA expansion to provide nutritional  
168 benefits. Second, the size of the local population around reefs (Figure 4C) determines the per  
169 capita consumption estimate for reef-caught seafood (Figure 3D). While large local populations  
170 will lead to low per capita impacts, but potentially high numbers of people impacted, small  
171 populations can have high per capita impacts. Third, increasing catches in coastal communities  
172 with high levels of inadequate intake (Figure 4E) has the greatest potential to decrease  
173 nutritional risks. On the other hand, increasing catch in communities that already have  
174 adequate nutrient intake will have minimal impact on nutritional status. Lastly, the data  
175 supported the addition of an interaction between sustainable-use MPA and the national  
176 efficacy of fisheries management in the model (Figure S2), reflecting those potential changes in

177 catch following MPA establishment will be different across countries depending on their  
 178 efficacy of fisheries management (Mora et al 2012). This reflects the fact that locations with  
 179 high fisheries management efficacy, which may have high biomass outside of MPAs, have little  
 180 potential for MPAs to provide net biomass increases. Because of uncertainty around all these  
 181 factors, the absolute number of people impacted by MPA expansion is also uncertain (from 0.4  
 182 to 1.8 million people; See Figure S8 for sensitivity analysis). Yet, our results suggest that MPAs  
 183 can benefit human nutrition, and the general geographical patterns appear robust to a large  
 184 range of conditions.



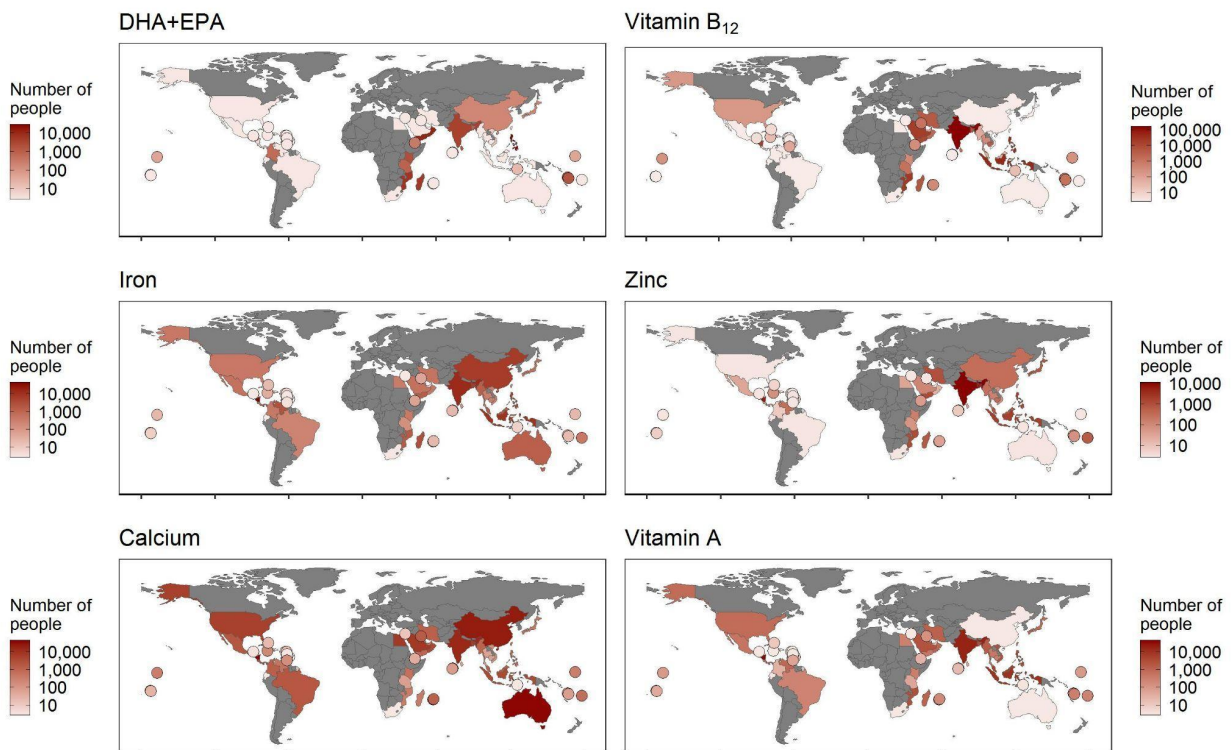
185  
 186  
 187 **Figure 3 – Nutritional impacts of expanding sustainable-use MPAs into unprotected reefs**  
 188 (A) non-MPA reef area within each country based on the overlap between MPA and reef areas;  
 189 (B) predicted change in catch in a hypothetical sustainable-use MPA expansion into non-MPA  
 190 reefs; (C) total number of people within a 10km buffer around non-MPA reefs; (D) predicted  
 191 change in per capita consumption of seafood for the population around non-MPA reefs; (E)  
 192 prevalence of present inadequate intake across countries (from Golden et al 2021); and (F)  
 193 predicted average change in inadequate intake for coral reef coastal populations across all  
 194 assessed nutrients (iron, EPA+DHA, calcium, zinc, and vitamins A and B<sub>12</sub>). Countries smaller  
 195 than 25,000 km<sup>2</sup> are illustrated as points.

196 **Nutritional targeting of vulnerable populations with marine conservation**

197 Because each country has different inadequacies among nutrients, sustainable-MPAs  
 198 can be strategically implemented to reduce specific nutritional inadequacies in coastal



199 populations (Figure 4). For example, strategically creating sustainable-use MPAs in Yemen and  
 200 Madagascar has the potential to reduce inadequate intake risks of omega-3 long-chain  
 201 polyunsaturated fatty acids (DHA+EPA). Increased intake of DHA+EPA can promote brain and  
 202 eye health and is associated with reduced risk of heart disease<sup>23</sup>. India and Bangladesh could  
 203 particularly benefit from increased supply of vitamin B<sub>12</sub>, where inadequacy is more than twice  
 204 the global average. Vitamin B<sub>12</sub> deficiency is associated with increased risk of heart disease and  
 205 cognitive decline<sup>24</sup>. Mozambique and Cambodia could benefit from increased supply of iron,  
 206 which is particularly important for healthy brain development and growth in children<sup>25</sup>, and can  
 207 prevent maternal mortality<sup>26</sup>. Sustainable-use MPAs in Nicaragua and Madagascar have the  
 208 potential to reduce inadequate intake of zinc, which supports immunity and is particularly  
 209 important for children and pregnant women<sup>27</sup>. In contrast, Kuwait and Indonesia could benefit  
 210 mostly from increased supply of calcium, which supports bone health and blood pressure<sup>28</sup>.  
 211 Lastly, Oman and Kiribati could benefit from increased vitamin A supply, which supports eye  
 212 health and cell growth<sup>29</sup>.  
 213



214  
 215  
 216 **Figure 4 - Potential nutritional benefits from sustainable-use MPA expansion.** Total reduction  
 217 in inadequate intake of key nutrients from coral reef sources for coastal human populations  
 218 from expansion of sustainable-use MPAs relative to non-MPA conditions. Countries smaller  
 219 than 25,000 km<sup>2</sup> are illustrated as points.

## 220 Discussion

221 Our global analysis suggests that effective sustainable-use MPAs have the potential to  
222 increase biomass and support nutritional security through increased catches. These benefits are  
223 on top of the traditional MPA impacts on livelihoods and marine biodiversity (in essence, co-  
224 benefits). We predict that expansion of sustainable-use MPAs to non-MPA reefs can have  
225 significant nutritional benefits, both sustaining the nutrition of vulnerable populations and  
226 decreasing the prevalence of inadequate intake of vital nutrients. For coastal communities that  
227 rely on coral reef resources for key nutrient intake, increasing sustainable supply of nutritious  
228 food can have significant positive impacts on their health and well-being. Many factors will  
229 affect the magnitude of impacts on a local scale. For example, we show that the quantity of  
230 seafood communities will catch (harvest rate) and how many people will consume that seafood  
231 (per capita consumption) are particularly important. Other local conditions such as reef area,  
232 prevalence of inadequate intake, and the efficacy of fisheries management will also play an  
233 important role.

234 Our model predictions are based on the current performance of sustainable-use MPAs.  
235 To restrict catch, these areas use an array of fisheries management tools to support reef  
236 management, including gear restrictions, access rights, size limits, temporal closures, bag limits,  
237 and more<sup>8</sup>. Some of these sustainable-use MPAs may be zoned for multiple uses, potentially  
238 containing small no-take areas within them, which may also benefit fished areas through  
239 spillover<sup>30,31</sup>. Because our results depend on actual performance of sustainable-use MPAs, if  
240 management of these areas improves (through more investment, management capacity,  
241 planning, community participation, etc.), potential biomass and consequential nutritional gains  
242 could be even greater. Understanding how close MPAs are to their maximum sustainable yield  
243 can provide a better estimate of their true potential to provide nutritional benefits. In addition,  
244 for fish stocks below maximum sustainable yield, recovery will likely require a short-term  
245 decrease in catch to obtain long-term gains<sup>32,33</sup>. Therefore, it is important to consider potential  
246 short-term nutritional costs to achieve predicted long-term nutritional benefits.

247 We consider sustainable-use MPAs as all conserved areas included in the MPAtlas where  
248 fishing is permitted<sup>9</sup>. In addition to state-managed areas, it encompasses many types of area-  
249 based management strategies such as Locally Managed Marine Areas (LMMAs), Environmental  
250 Protection Areas (EPAs), and Marine Conservation Areas (MCAs). In many cases, these and  
251 other effective area-based conservation measures (OECMs) empower local stakeholders and  
252 incentivize resource management, while considering local characteristics of the fishery and  
253 cultural traditions of coastal communities<sup>34</sup>. Several studies have shown that when  
254 communities are empowered and have secure rights to a fishery, there is greater incentive for  
255 successful fisheries management yielding nutritional benefits<sup>35-37</sup>. However, more research is  
256 needed to evaluate how different types of sustainable-use MPAs affect biomass, catch and  
257 human nutrition. Various other management measures (e.g., individual quotas) could also  
258 improve nutrient supply but are not commonly implemented in sustainable-use MPAs in  
259 tropical coastal areas.

260 Expansion of sustainable-use MPAs depends on strong local, national and global  
261 commitments and investments. Today, only a small fraction of resources from governments,  
262 regional development banks, and multilateral funding agencies are directed to strengthening

263 governance of small-scale fisheries<sup>38</sup>. The lack of adequate capacity to manage MPAs has led to  
264 the creation of many “paper parks”, which are designated areas that are not effectively  
265 implemented<sup>9</sup>, thus having limited potential to provide environmental, economic, and  
266 nutritional benefits. In addition, implementation of sustainable-use MPAs are usually not part  
267 of the food and nutrition agenda of coral reef countries. Our results suggest that populations  
268 who depend on reef systems for nutrition would benefit from funding that is directed to  
269 sustainable marine resource management.

270 Climate change impacts on reef ecosystems creates an uncertain future for millions of  
271 people who depend on reef fisheries for nutrition and livelihood. Without actions (such as  
272 sustainable-use MPAs) to ensure sustainability of catch into the future, there can be significant  
273 loss of nutritional benefits in the coming years, with important implications for public health.  
274 Therefore, it is not only the prevention of inadequate intake and the supporting benefits of  
275 MPAs that can be important, but also the buffering against the risk of future loss in climate-  
276 vulnerable systems. Without conservation action in the present, the risk of future negative  
277 nutritional impacts is inevitable in the long-term.

278 To achieve all United Nation Sustainable Development Goals (SDGs), it is critical to find  
279 synergistic strategies whereby coastal marine governance can achieve human health and  
280 conservation outcomes, ensuring both sustainable populations of diverse species and the flow  
281 of vital nutrients to vulnerable populations. While SDG goals of “zero hunger” and “good health  
282 and well-being” are particularly dependent on food availability to vulnerable communities, the  
283 goal of “life below water” is particularly dependent on ocean conservation. Here, we show how  
284 meeting the “life below water” goal can help support “zero hunger” and other goals. It is  
285 essential to analyze tradeoffs between both goals and find win-win solutions that benefit both  
286 people and the planet.

## 287 Methods

### 288 Reef fish underwater surveys

289 We compiled information of coral reef transects from the Reef Life Survey (RLS) and Atlantic  
290 and Gulf Rapid Reef Assessment (AGRRA) databases. Both databases are based on underwater  
291 fish counts by size class within a belt transect. We then calculated individual biomass by using  
292 length-weight relationships published for all species on Fishbase (Froese and Pauly 2021) and  
293 then multiplying individual biomass by the total number of fish within each size class. The final  
294 compiled database contained 16,365 surveys from 2,518 tropical coral reef sites (i.e., within  
295 23.5 latitude degrees) distributed across 53 countries. Data collections were conducted from  
296 1997 to 2020. Where data from multiple years were available for a single site, we included only  
297 the most recent year. To estimate the “fishable biomass” we retained only fish larger than  
298 10cm<sup>39</sup>. Because underwater fish counts do not accurately capture biomass of large schools of  
299 pelagic fish (e.g. Scombrids, Sphyraenids) or large transient fish, we removed all shark and ray  
300 species<sup>39,40</sup>. In addition, because of data constraints we only consider reef fish catch as nutrient  
301 source, however, invertebrate species and aquatic plants can also be an important source of  
302 nutrients in many low-income countries.

303 We divided all survey sites into three basic categories: non-MPA areas, sustainable-use  
304 MPAs and no-take MPAs. Non-MPA areas are all sites outside of marine protected areas, which  
305 can be subject to regional or national-level policies (whether enforced or not). In general, these  
306 areas are not managed through additional area-based regulations. Sustainable-use MPA sites  
307 are all sites within an area-based management system that allows fishing within its borders,  
308 including areas such as multiple-use MPAs, Locally Managed Marine Areas (LMMAs) or  
309 Environmental Protection Areas (EPAs). No-take MPAs, in contrast, is a term used here to  
310 describe areas where no forms of fishing are allowed (also known as fully protected, marine  
311 refugia, etc.). These areas may contain other economic activities such as tourism or can be  
312 strictly for research and conservation. In total, we had 1,117 non-MPA sites, 804 sustainable-  
313 use MPA sites, and 500 no-take MPA sites.

### 314 Spatial analysis

315 For each coral reef polygon, we calculated the total reef area that is within a sustainable-use  
316 MPA, no-take MPA, and non-MPA. To do this, we intersected all coral reef polygons<sup>19</sup> with MPA  
317 polygons from the MPAtlas<sup>18</sup>. Within the MPAtlas database, each MPA was divided into  
318 sustainable-use or no-take MPAs, allowing the calculation of the percentage of reefs falling  
319 within each category.

320 Next, we calculated the population around existing MPAs and non-MPA reefs by  
321 intersecting all reef and MPA polygons with the raster of the gridded population of the world in  
322 2019 (CIESIN, 2019). We calculated the population within 5, 10, 20, 25 and 30 kilometers  
323 buffers around reefs and MPAs. To avoid double counting coastal populations, all overlapping  
324 polygon buffers were aggregated. We used the `sf` package<sup>41</sup> in R statistical software<sup>42</sup> to  
325 perform all spatial analysis.

### 326 Predicting fish biomass

327 We used a Bayesian model to predict fish biomass (above 10cm) in every coral reef around the  
328 world. For all coral reef polygons, we predicted the biomass of reef fish per unit area (kg/ha)  
329 under two alternative conditions (non-MPA and sustainable-use MPAs) while accounting for  
330 each site's own environmental and social covariates (Figure S2). Site covariates considered in  
331 this analysis included chlorophyll concentration, sea surface temperature mean, sea surface  
332 temperature range, nitrate concentration, wave exposure, reef area, shore distance, human  
333 population, market distance, human development index, and fisheries management  
334 effectiveness<sup>39,43</sup>. To account for variability in MPA effectiveness across countries (due to  
335 differences in management, staff capacity, state of the reef prior to MPA establishment, etc.),  
336 we also considered an interaction term between presence of sustainable-use MPAs and  
337 fisheries management effectiveness across nations<sup>15</sup>. In addition, we set ecoregions as a  
338 random effect to account for the spatial structure of the data. Collinearity among covariates  
339 was examined based on bivariate correlations and variance inflation factors, which led to the  
340 exclusion of both environmental variables (pH, salinity, primary productivity, and minimum sea  
341 surface temperature) and social variables (land cover, fisher density, and government  
342 effectiveness).

343 We used the **brms** package<sup>44</sup> to construct the model in R statistical software<sup>42</sup>. Models  
344 were run using the Hamiltonian Monte Carlo algorithm for 10000 iterations and 4 chains.  
345 Posterior estimates were informed by the data alone (weakly informed priors). Convergence  
346 was monitored by examining posterior chains and stability and checking if the scale reduction  
347 factor was close to 1. Next, we tested a null model with intercepts only and a full model that  
348 included all covariates. We compared both models through leave-one-out cross validation  
349 information criteria (LOOIC), ensuring that our full model performed better than the null model  
350 ( $\text{elpd\_diff} = -95.7$ ). In addition, we used LOOIC to test if the model with interaction performed  
351 better than the model without the interaction term between MPA and fisheries management  
352 effectiveness. To examine model fit and homoscedasticity, we checked residuals against fitted  
353 values and conducted posterior predictive checks (Figure S3). In addition, we evaluated the  
354 goodness-of-fit of the model using leave-one-out cross-validation ( $\text{loo\_r2} = 0.41$ ). When  
355 predicting biomass in reef polygons, we assumed a model with a random intercept since not all  
356 ecoregions with reef polygons are represented in our data.

357 Biomass predictions per unit area (kg/ha) were then multiplied by the area in each reef  
358 polygon to estimate the total reef fish biomass on each reef. Implicitly, we thereby assumed  
359 equal productivity across each reef polygon. Reef polygons range from about 100 square  
360 meters to 9.8 thousand square kilometers, with a median of 6.3 square kilometers. We  
361 acknowledge that biomass estimates are affected by (i) our reliance on biomass and social and  
362 environmental conditions for reefs within our dataset which may or may not be representative  
363 of all reef systems, (ii) potential spatial and temporal imprecision, (iii) other factors not  
364 accounted in the model could also drive biomass, (iv) social and environmental conditions can  
365 vary over smaller scales than reef polygons considered here, however, data collections enabling  
366 to account for local patchiness in productivity would be extremely challenging if not impossible.

### 367 Predicting potential changes in catch due to MPA establishment and operation

368 The potential change in catch from sustainable-use MPAs was estimated by comparing  
369 predicted biomass under non-MPA and MPA conditions. To estimate catch from biomass, we  
370 used a simple surplus production model (Shafer, 1954). This model assumes that the harvest  
371 rate that produces maximum sustainable yield ( $F_{\text{MSY}}$ ) is half of the intrinsic population growth  
372 rate ( $r$ ) of the species. Therefore, species that grow and reproduce faster can sustain higher  
373 levels of harvest than slow growing species. Population-level intrinsic growth rates were  
374 derived from McClanahan and Graham 2015<sup>45</sup> ( $r = 0.23$ ). Harvest rate will also depend on the  
375 standing biomass in each site relative to the assumed biomass that maximizes yield ( $B_{\text{MSY}}$ ), to  
376 reflect the fact that sites with lower biomass should have relatively higher harvest rates than  
377 sites with higher biomass. As a proxy for  $B_{\text{MMSY}}$  we used the 90th biomass quantile of predicted  
378 biomass for sustainable-use MPA (544 kg/ha), assuming that these sites are fishing at MMSY or  
379 multispecies maximum sustainable yield and within limits of globally proposed  $B_{\text{MMSY}}$ <sup>43,45</sup> (see  
380 Figure S4 for a sensitivity analysis). Sensitivity of results to harvest rate assumptions are shown  
381 in Figure S5 (growth rates varying from 0.1 to 0.6). Regional patterns and percent changes in  
382 catch are not affected by assumptions of harvest rate. However, total absolute numbers of  
383 people affected by MPAs expansion are sensitive to harvest rate assumptions.

## 384 Assigning nutritional content to reef fish catch

385 To assign specific species to the predicted change in reef fish catch, we used the Sea  
386 Around Us database (Pauly et al 2020), allocating total catch estimates to species proportions  
387 based on the proportion of reef species caught in each country in 2014. We used SAU to  
388 account for country-level differences in catch and because our surveys did not cover all  
389 countries containing coral reef polygons. To obtain this information from SAU, we first  
390 separated production from artisanal and subsistence sectors. Next, we identified reef species as  
391 occurring in the following functional groups: "Medium reef assoc. fish (30 - 89 cm)", "Large reef  
392 assoc. fish ( $\geq 90$  cm)" and "Small reef assoc. fish ( $< 30$  cm)". In addition, we restricted the data  
393 to families that were recorded in the underwater visual surveys.

394 To assign nutritional content to reef fish species, we used the Aquatic Foods  
395 Composition Database (AFCD), a comprehensive database containing 3,750 records of nutrient  
396 content from global databases and peer-reviewed literature<sup>6</sup>. We then use a hierarchical  
397 approach<sup>6</sup> to match specific species taxonomic information with AFCD and fill nutrient  
398 information for species not present in the database. This hierarchy is based on the following  
399 order: 1) scientific name, and then the taxa-specific average of 2) genus, 3) family, 4) order,  
400 and 5) class. We then matched the following nutrients: iron, zinc, protein, vitamin A, vitamin B<sub>12</sub>  
401 and calcium (Figure S6). These nutrients were chosen because of their high concentration in  
402 aquatic species, their importance in human nutrition, and their inadequate intake across many  
403 countries<sup>6</sup>. We then multiplied the predicted catch by the edible portion of each species based  
404 on AFCD data and multiplied further by nutritional value to obtain the total nutrient supply for  
405 each nutrient.

## 406 Calculating per capita nutrient supply and catch from MPA expansion

407 We calculated the per capita nutrient supply by dividing total nutrient supply by the  
408 human population around reefs. Although some valuable reef species are traded in  
409 international or regional markets, we assumed for simplicity that all extra catch from MPA  
410 expansion will be consumed by coastal communities within a 10 km buffer around reefs. At a  
411 local scale, the per capita consumption will depend on accessibility: distance of the reef from  
412 the community, the size of boats, trade dynamics, etc. For example, a 20 km radius will capture  
413 the travel distance that most fishers take in subsistence/artisanal fisheries<sup>46,47</sup>. However,  
414 acknowledging the high uncertainty around this value, we tested multiple alternative buffer  
415 sizes around reefs to estimate per capita nutrient supply (Figure S7). Larger buffers around  
416 reefs increased the number of people impacted, and, thus, lowered per capita nutrient supply.  
417 Although the magnitude of impacts changed depending on buffer size, regional patterns were  
418 not affected by the assumed population around reefs.

419 To calculate the number of people supported by sustainable-use MPAs, we first  
420 estimated the per capita reef fish catch by dividing the predicted catch in each reef polygon by  
421 the population within a buffer around the reef. Next, we estimated the percent contribution of  
422 per capita reef catch relative to per capita national average consumption of aquatic animal  
423 sourced foods based on the Global Nutrient Database (GND)<sup>48</sup>. The GND used the Food and  
424 Agriculture Organization of the United Nations Supply and Utilization Accounts (SUAs) to obtain

425 estimates of apparent per capita consumption of 22 food groups and nutrient supply for 156  
426 nutrients across 195 countries. We considered coral reefs to provide a meaningful contribution  
427 when coral reef catch represented at least 5% of aquatic animal food intake (see Figure S8 for  
428 sensitivity analysis). We then summed across all reefs that provided a meaningful contribution  
429 to calculate the total number of people that could potentially be supported by existing and  
430 future sustainable-use MPAs.

### 431 Calculating the contribution of MPAs to human nutrition

432 To calculate potential nutritional effects of MPA expansion, we compared a baseline  
433 scenario with a scenario of increased reef fish consumption through an expansion of  
434 sustainable-use MPAs. Baseline conditions were calculated using estimates of nutrient  
435 consumption in 2017 from Global Nutrient Database (GND)<sup>48</sup>. The MPA expansion scenario was  
436 calculated by adding the per capita nutrient supply from MPA expansion to this baseline level of  
437 nutrient intake.

438 We then calculated the prevalence of inadequate intake for current conditions and MPA  
439 expansion scenarios to obtain the difference in inadequate intake across both scenarios.  
440 Prevalence of inadequate intake was calculated following three main steps. First, we  
441 disaggregated country-level mean intakes into age-sex mean intakes using the Global Expanded  
442 Nutrient Supply (GENuS) database for all nutrients except DHA+EPA and vitamin B12, which are  
443 not included in the GENuS database<sup>49</sup>. Second, using dietary recall data from SPADE (Statistical  
444 Program to Assess Habitual Dietary Exposure), we derived habitual dietary intake distributions  
445 across age-sex groups and geographies<sup>50</sup>. We used SPADE outputs to describe the shape  
446 (gamma or lognormal distribution) of intake distribution for each age–sex group and to derive  
447 age–sex mean intakes for DHA+EPA and vitamin B12. Lastly, we calculated the prevalence of  
448 inadequate intake using the summary exposure values, or SEVs<sup>6,21</sup>. SEVs estimate the  
449 population-level risk related to diets by comparing intake distributions with requirements. The  
450 latter are continuous risk curves with values of 1 for low intake, 0 for high intakes and 0.5 for  
451 intakes at the Estimated Average Requirement (EAR). These absolute risk curves are then  
452 constructed as the cumulative normal distribution function of requirements with a mean at the  
453 EAR and a coefficient of variation of 10%<sup>51</sup>. EAR estimates were derived from several sources  
454 (FAO, Institute of Medicine), and a coefficient of variation of 25% was used to account for  
455 uncertainties regarding recommended intakes. For DHA+EPA, we used the relative risk curves  
456 that are associated with ischaemic heart disease and have different values for adolescent and  
457 adult subpopulations (with no risk for children)<sup>21</sup>. Estimated prevalence of inadequate intake  
458 range from 0% (no risk) to full population-level risk (100%).

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## 565 Author contributions

566 DV, CDG, MM, AZ, and DG conceptualized the research idea, with significant methodological  
567 and design input from JZM, AS, CMF, and NK. DV led paper analysis and visualization. DV  
568 drafted the original manuscript, and all co-authors edited and revised the writing.

## 569 Data Availability

570 Reef Life Survey data available online (<https://reeflifesurvey.com/>)  
571 AGRRRA data available upon request (<https://www.agrra.org/>)  
572 Aquatic Food Composition Database available through Harvard dataverse  
573 (<https://dataverse.harvard.edu/dataset.xhtml?persistentId=doi:10.7910/DVN/KIONYM>)  
574 Sea Around Us database is available online (<https://www.searoundus.org/>)

## 575 Code Availability

576 All code used in the analysis is available on GitHub.

## 577 Competing interests

578 MM and AZ are employees of Conservation International, an organization whose mission is to  
579 empower societies to care for nature for the well-being of humanity through science,  
580 partnerships, and field demonstrations.

## 581 Additional Information

- 582 ● Supplementary Information is available for this paper.
- 583 ● Correspondence and requests for materials should be addressed to Daniel F. Viana.
- 584 ● Reprints and permissions information is available at [www.nature.com/reprints](http://www.nature.com/reprints).

## Supplementary Files

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